ANNUAL STOCK ASSESSMENT AND FISHERY EVALUATION REPORT: MARIANA ARCHIPELAGO FISHERY ECOSYSTEM PLAN 2022







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The ANNUAL STOCK ASSESSMENT AND FISHERY EVALUATION REPORT for the MARIANA ARCHIPELAGO FISHERY ECOSYSTEM PLAN 2022 was drafted by the Fishery Ecosystem Plan Team. This is a collaborative effort primarily between the Western Pacific Regional Fishery Management Council (WPRFMC), National Marine Fisheries Service (NMFS) Pacific Island Fisheries Science Center (PIFSC) and Pacific Islands Regional Office (PIRO), Hawaii Division of Aquatic Resources (HDAR), American Samoa Department of Marine and Wildlife Resources (DMWR), Guam Division of Aquatic and Wildlife Resources (DAWR), and Commonwealth of the Mariana Islands (CNMI) Division of Fish and Wildlife (DFW).

This report attempts to summarize annual fishery performance looking at trends in catch, effort and catch rates as well as provide a source document describing various projects and activities being undertaken on a local and federal level. The report also describes several ecosystem considerations, including fish biomass estimates, biological indicators, protected species, habitat, climate change, and human dimensions. Information like marine spatial planning and best scientific information available for each fishery are described. This report provides a summary of annual catches relative to the Annual Catch Limits established by the Council in collaboration with the local fishery management agencies.

Cover Image: Igor Cristino Silva Cruz

A report of the Western Pacific Regional Fishery Management Council 1164 Bishop Street, Suite 1400, Honolulu, HI 96813

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Published in the United States by the Western Pacific Regional Fishery Management Council under NOAA Award #NA20NMF4410013.

ISBN 978-1-950193-31-8

This document can be cited as follows:

WPRFMC. 2023. Annual Stock Assessment and Fishery Evaluation Report for the Mariana Archipelago Fishery Ecosystem Plan 2022. T Remington, M Seeley, A Ishizaki (Eds.). Honolulu: Western Pacific Regional Fishery Management Council.

The Western Pacific Regional Fishery Management Council acknowledges the valuable contributions of the following Plan Team members and others for drafting sections of this report:

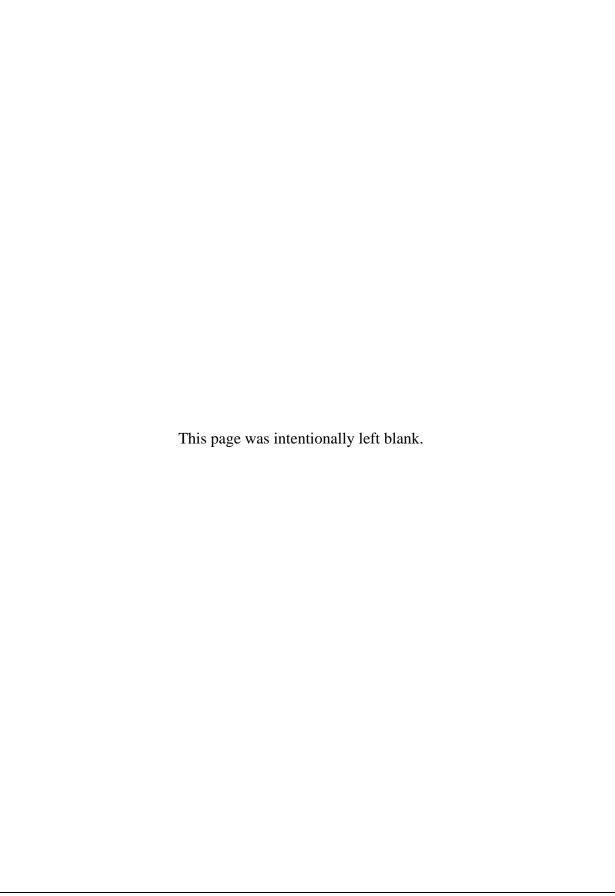
Guam Division of Aquatic and Wildlife Resources: Brent Tibbatts and Jason Biggs.

CNMI Division of Fish and Wildlife: Angela Delacruz.

NMFS Pacific Islands Fisheries Science Center: Marc Nadon, Felipe Carvalho, Jenny Suter, Joseph O'Malley, Eva Schemmel, Kisei Tanaka, Thomas Oliver, Hui Shi, Frank Parrish, Tye Kindinger, Keith Bigelow, Robert Ahrens, Marlowe Sabater, T. Todd Jones, Minling Pan, Danika Kleiber, Justin Hospital, Adam Ayers, and Kirsten Leong.

NMFS Pacific Islands Regional Office: Brett Schumacher, Heather Cronin, Savannah Lewis, Irene Kelly, Sean Hanser, and Keith Kamikawa.

The Council also acknowledges the staff of the **NMFS PIFSC Fisheries Research and Monitoring Division data programs** for providing the technical support to generate the data summaries.



EXECUTIVE SUMMARY

As part of its five-year fishery ecosystem plan (FEP) review, the Western Pacific Regional Fishery Management Council (WPRFMC; the Council) identified its annual reports as a priority for improvement. The former annual reports have been revised to meet National Standard regulatory requirements for Stock Assessment and Fishery Evaluation (SAFE) reports. The purpose of the reports is twofold: to monitor the performance of the fishery and ecosystem to assess the effectiveness of the FEP in meeting its management objectives; and to maintain the structure of the FEP living document. The reports are comprised of three chapters: Fishery Performance, Ecosystem Considerations, and Data Integration. The Council will iteratively improve the annual SAFE report as resources allow.

The Fishery Performance chapter of this report first presents a general description of the local fisheries within the Commonwealth of Northern Mariana Islands (CNMI) and Guam, focusing on the management unit species (MUS), particularly bottomfish MUS (BMUS), and accompanied by the monitoring of ecosystem component species (ECS). The fishery data collection system is explained, encompassing creel surveys and commercial receipt books. Fishery meta-statistics for BMUS and ECS are organized into summary dashboard tables showcasing the values for the most recent fishing year and a comparison to short-term (10-year) and long-term (20-year) averages. Time series for catch and effort statistics are also provided along with implemented annual catch limits (ACLs).

In 2019, NMFS developed a stock assessment for the Guam and CNMI BMUS stock complexes (Langseth et al. 2019). While the CNMI BMUS were determined to be healthy, the Guam complex was determined to be overfished but not experiencing overfishing. In response to a notification from NMFS of the change in stock status for Guam BMUS, the Council developed a rebuilding plan for the fishery, including an ACL of 31,000 lb and in-season accountability measures, which became effective on March 21, 2022 (87 FR 9271, February 18, 2022).

In the CNMI, neither the total estimated BMUS catch for 2022 (47,564 lb) nor the recent three-year average catch of BMUS (55,916 lb) exceeded the ACL of 84,000 lb or the annual catch target (ACT) of 78,000 lb. Total estimated BMUS catch in Guam for 2022 was relatively high at 44,788 lb and is currently undergoing validation by scientists at the Pacific Islands Fisheries Science Center (PIFSC).

There are no other MUS in Guam or the CNMI under the Mariana Archipelago FEP, as an amendment to the FEP in early 2019 reclassified most of the MUS as ECS except for the current BMUS (84 FR 2767, February 8, 2019). ECS do not require management under ACLs or accountability measures but are still to be monitored regularly in the annual SAFE report through a one-year snapshot of the ten most-caught ECS, complete catch time series of prioritized ECS as selected by the Guam Division of Aquatic and Wildlife Resources (DAWR) and the CNMI Division of Fish and Wildlife (DFW), as well as trophic and functional group biomass estimates from fishery independent surveys.

In the CNMI, total estimated BMUS catch notably decreased in 2022 to 47,564 lb, a 44% increase from the 10-year (i.e., short-term) average and a 18% increase from the 20-year (i.e., long-term) average. BMUS catch from commercial purchase data in 2022 also showed increases of 78% and 99% relative to the historical trends at 32,161 lb. However, CPUE for BMUS harvested by the bottomfish handline gear were lower than the decadal averages for both metrics

presented, pounds per trip and pounds per gear hour, except when considering CPUE in pound per gear hour relative to the 20-year average. There were 38 lb/trip of BMUS harvested by bottomfish fishing (14% decrease from the 10-year average), and approximately 3.84 lb/gear hour of BMUS harvested by bottomfish fishing (4% decrease and 28% increase from the short-and long-term averages, respectively). The number of bottomfish fishing trips that harvested BMUS as tallied in the creel surveys was 37 in 2022, a notable 85% increase from the 10-year average and 16% increase from the 20-year average. The tallied number of bottomfish fishing gear hours was 368 (44% increase and 27% decrease from the short- and long-term trends, respectively). There were 20 unique vessels tallied in the fishery in 2022 with an average of just two fishers per bottomfish fishing trip. There was no recorded bycatch in boat-based BMUS fisheries of the CNMI in 2022.

For the top ten landed ECS in CNMI in 2022, available data streams showed that the bigeye scad (*Selar crumenophthalmus*) had the most catch in the creel survey data (5,743 lb) and commercial landings (20,295 lb). The second most caught species were parrotfish for both data streams. Other species of note include unicornfish (*Naso* spp.) and several species of emperor. Most of the remainder of the top ten ECS from commercial purchase data were family groups (e.g., Acanthuridae) due to how the species are categorized by vendors on the commercial receipts.

For prioritized ECS (i.e., those selected by DFW) in CNMI, most species notably exceeded their short- and long-term averages, with *Naso lituratus* and *N. unicornis* having the highest catch levels. There were species codes for just six of the seven prioritized ECS species in CNMI commercial purchase data, as *Scarus ghobban* does not get actively recorded. In the data for the six available species, commercial purchase showed catches of zero for two species, *Mulloidichthys flavolineatus* and *L. harak*. For the species for which data were available, 2022 catches exceeded the historical averages for all four species.

For the BMUS fishery in Guam in 2022, total estimated BMUS catch was 45,071 lb, but this number is undergoing verification prior to finalization. No commercial catch trends were reported due to issues with data confidentiality (i.e., less than three dealers and/or vendors reporting data). CPUE for BMUS harvested by the bottomfish handline gear was presented using two metrics in the 2022 report, pounds per trip and pounds per gear hour. There was 21 pounds of BMUS caught per trip in Guam in 2022, a 24% increase from the recent 10-year average and a 11% increase from the 20-year average. CPUE in pounds per gear hour was 1.40 for BMUS harvested with the bottomfish handline gear, which coincided with increases relative to the recent 10- and 20-year averages (32% and 9%, respectively). The tallied number of fishing trips that harvested BMUS increased by 58% compared with the 10-year average to 95 trips, which also represented a 56% increase relative to the 20-year average. The number of bottomfish fishing gear hours on trips that harvested BMUS was 1,419, a 27% increase from the 10-year average and a 40% increase to the 20-year average. The tallied number of unique vessels harvested BMUS in Guam was 63, an increase to both the 10- and 20-year averages by 47% and 43%, respectively. The average number of fishers per trip was 2, which represented a decrease from the historical average of 3. There was one tallied BMUS release in 2022, and a relatively low amount of non-BMUS released. The overall bycatch rate for Guam boat-based fisheries was 0.82% in 2022, representing a decrease from historical averages.

For the top ten landed ECS in Guam in 2022, available data showed that the bigeye scad (*Selar crumenophthalmus*; 17,193 lb) had the most catch from creel survey data while "reef fish" had the most catch in the commercial purchase data (15,755 lb). The second most caught ECS in the

creel survey data was the bluespine unicornfish (*N. unicornis*; 14,046 lb) followed by assorted reef fish (6,361 lb). Several other species had notable catch estimates in the creel survey data, including *Etelis boweni* and *N. lituratus*. Many of the top ten ECS from commercial purchase data were family groups (e.g., Scaridae and Lethrinidae) due to how the species are categorized by vendors.

For prioritized ECS (i.e., those selected by DAWR) in Guam, 2022 creel survey catch estimates for *N. unicornis* and *Chlorurus frontalis* were much higher than both of their associated 10- and 20-year averages. Other prioritized ECS had mixed trends. Commercial purchase invoices were only able to capture *S. spinus* from the DAWR-prioritized ECS, which had non-disclosed catch information for 2022 due to data confidentiality rules.

Federal permit data show that there were nine bottomfish permit holders in the CNMI in 2022, but there were no active permits for lobster or shrimp. There were no active federal permits for Guam bottomfish, lobster, or shrimp fisheries. No catch data were reported in federal logbooks by permit holders for these fisheries in 2022.

An Ecosystem Considerations chapter was added to the annual SAFE report following the Council's review of its FEPs and revised management objectives. Fishery independent ecosystem survey data, socioeconomics, protected species, climate and oceanographic, essential fish habitat, and marine planning information are included in the Ecosystem Considerations chapter. A special section was added to the report in 2020 and 2021 describing the impacts of COVID-19 on archipelagic fisheries and fishing communities of Guam and the CNMI, but this section was removed from the 2022 report.

Fishery independent ecosystem data were acquired through visual surveys conducted by the National Marine Fisheries Service (NMFS) Pacific Islands Fisheries Science Center (PIFSC) through the National Coral Reef Monitoring Program (NCRMP) under the Ecosystem Sciences Division (ESD) in CNMI, the Pacific Remote Island Areas (PRIA), American Samoa, Guam, the Main Hawaiian Islands (MHI), and the Northwestern Hawaiian Islands (NWHI). This report describes mean fish biomass of functional, taxonomic, and trophic groups for coral reefs as well as habitat condition using mean coral coverage per island for each of these locations from 2010 to 2020. Surveys were conducted in the Mariana Archipelago in 2022. Generally, reef coverage and fish biomass increased in the CNMI in the northern islands relative to the southern islands. Guam also had slight increases in coral coverage and fish biomass relative to surveys conducted in 2016-2018.

For CNMI, life history parameters including maximum age, asymptotic length, growth coefficient, hypothetical age at length zero, natural mortality, age at 50% maturity, age at sex switching, length at which 50% of a fish species are capable of spawning, and length of sex switching are provided for eight prioritized ECS and the 13 BMUS where available. The same nine life history parameters are provided for nine prioritized ECS and the 13 BMUS in Guam where available. Length derived parameters summarized for coral reef ECS and bottomfish in CNMI and Guam include maximum fish length, mean length, sample size for L-W regression, and length-weight coefficients. Length derived values are presented for the same ECS and BMUS as the life history parameters for both CNMI and Guam where available. This year, research on age, growth, and reproduction for *Pristipomoides zonatus* and *Variola louti* were added, and there is ongoing work for aging criteria and reproduction for several species such as *Etelis coruscans* and *P. auricilla*.

The socioeconomics section outlines the pertinent economic, social, and community information available for assessing the successes and impacts of management measures or the achievements of the FEP for the Mariana Archipelago. It meets the objective "Support Fishing Communities" adopted at the 165th Council meeting; specifically, it identifies the various social and economic groups within the region's fishing communities and their interconnections. The section begins with an overview of the socioeconomic context for the region, provides a summary of relevant studies and data for the Mariana Archipelago, presents available socioeconomic data (including annual data for revenue, fish price, and cost of fishing), and then lists relevant socioeconomic studies for fisheries within the Mariana Archipelago. Considering the CNMI bottomfish fishery, there was an estimated total of 32,160 pounds sold for \$180,672. Fish price notably increased from 2021 to 2022 to \$5.62 per pound. The average cost of a bottomfish trip in CNMI in 2022 was higher than 2021 at \$43 due to decreased usage amid increased costs for fuel. The top 10 ECS in CNMI had 67,301 pounds sold for a revenue of \$244,011. The majority of socioeconomic information for Guam's bottomfish fishery were unavailable due to data confidentiality in 2022. The costs of fishing in Guam for 2022 were also not disclosed due to data confidentiality rules.

The protected species section of this report summarizes information and monitors protected species interactions in fisheries managed under the Mariana Archipelago FEP. These fisheries generally have limited impacts to protected species and do not have federal observer coverage. Consequently, this report tracks fishing effort and other characteristics to detect potential changes to the level of impacts to protected species. Fishery performance data contained in this report indicate that there have been no notable changes in the fisheries that would affect the potential for interactions with protected species, and there is no other information to indicate that impacts to protected species have changed in recent years in the Mariana Archipelago. On June 6, 2019 (extended on August 11, 2020, December 15, 2020, and February 9, 2022), NMFS determined that the conduct of these bottomfish fisheries during the period of consultation will not violate Endangered Species Act (ESA) Section 7(a)(2) and 7(d). In 2022, there were updates the section associated with information from the new biological opinion conducted for regional bottomfish fisheries. The consultation concluded that the Guam and CNMI bottomfish fisheries are likely to adversely affect oceanic whitetip sharks with no jeopardy, and additional information on this document was added to the protected species section of this report.

The climate change section of this report includes indicators of current and changing climate and related oceanic conditions in the geographic areas for which the Council has jurisdiction. In developing this section, the Council relied on a number of recent reports conducted in the context of the U.S. National Climate Assessment including, most notably, the 2012 Pacific Islands Regional Climate Assessment and the 'Ocean and Coasts' chapter of the 2014 report on a Pilot Indicator System prepared by the National Climate Assessment and Development Advisory Committee. The primary goal for selecting the indicators used in this report is to provide fisheries-related communities, resource managers, and businesses with climate-related situational awareness. In this context, indicators were selected to be fisheries relevant and informative, build intuition about current conditions considering changing climate, provide historical context, and recognize patterns and trends.

The trend of atmospheric concentration of carbon dioxide (CO₂) is increasing exponentially with a time series maximum at 419 ppm in 2022. Since 1989, the oceanic pH at Station ALOHA in Hawaii has shown a significant linear decrease of -0.045 pH units, or roughly a 10.9% increase

in acidity ([H+]), and was 8.05 in 2021. The Oceanic Niño Index, which is a measure of the El Niño – Southern Oscillation (ENSO) phase, indicated La Niña conditions throughout 2022. The Pacific Decadal Oscillation (PDO) was negative in 2022. The Accumulated Cyclone Energy (ACE) Index (x 10⁴ kt²) was average in the Eastern and Western North Pacific, below average in the Central North and South Pacific. Annual mean sea surface temperature (SST) around the Mariana Archipelago was 28.8 °C in 2022, and over the period of record, annual SST has increased at a rate of 0.0247 °C/year. The annual anomaly was 0.43 °C hotter than average, with intensification in the northern islands. The Mariana Archipelago experienced a minor coral heat stress event in the second half of 2022 with mass bleaching expected. Annual mean chlorophyll-*a* was 0.054 mg/m³ in 2022, and the annual anomaly was 0.0015 mg/m³ lower than average. Rainfall trends in the Mariana Archipelago were mixed over the course of 2022. The local trend in sea level rise is 4.17 millimeters/year based on monthly mean sea level data from 1993 to 2022, which is equivalent to a change of 1.37 feet in 100 years.

The Mariana Archipelago FEP and National Standard 2 guidelines require that this report include a report on the review of essential fish habitat (EFH) information. In the 2017 annual reports, a literature review of the life history and habitat requirements for each life stage for four species of reef-associated crustaceans that are landed in commercial fisheries Western Pacific region was presented, including information on two species of spiny lobster (*Panulirus marginatus* and *Scyllarides squammosus*), scaly slipper lobster (*Scyllarides squammosus*), and Kona crab (*Ranina ranina*). In the 2019 annual report, a review of EFH for reef-associated crustaceans in the MHI and Guam was included. The 2022 report also presents levels of EFH information available for Mariana Archipelago MUS. The National Standard guidelines also require a report on the condition of the habitat. In previous annual SAFE reports, data on benthic cover were included as indicators, pending development of habitat condition indicators for the Mariana Archipelago not represented in other sections of this report. The annual report addresses Council directives toward its Plan Team, but there were no directives associated with EFH in 2022.

The marine planning section of this report records activities with multi-year planning horizons and begins to track the cumulative impact of established facilities. Development of the report in the future will focus on identifying appropriate data streams. Military activities in the Marianas continue to impact fisheries and their access.

The Data Integration chapter of this report is under development. The chapter explores the potential association between fishery parameters and ecologically-associated variables that may be able to explain a portion of the variance in fishery-dependent data. A contractor completed preliminary evaluations in 2017, and results of exploratory analyses were included for the first time in the 2017 annual SAFE report. Going forward with the data integration analyses and presentation of results for Chapter 3 of the annual SAFE reports, the Council's Archipelagic Fishery Ecosystem Plan Team (Plan Team) suggested several improvements to implement in the future: standardizing and correcting values in the time series, incorporating longer stretches of phase lag, completing comparisons on the species-level and by dominant gear types, incorporating local knowledge on shifts in fishing dynamics over the course of the time series, and utilizing the exact environmental data sets presented in the Ecosystem Consideration chapter of this annual SAFE report. Many of these recommendations were applied to a revisited analysis in the Hawaii annual SAFE report in 2018 with similar plans for Mariana Archipelago data integration analyses in future report cycles. Implementation of these suggestions will allow for the preparation of a more finalized version of the data integration chapter in coming years. The

chapter will be updated in the future as resources allow. For the 2020 report, several recent relevant abstracts from primary publications related to data integration were added to the Data Integration chapter.

Plan Team members agreed to carry out the following recommendations, some of which are relevant to the Mariana Archipelago annual SAFE report:

Regarding the bycatch summary improvements, the APT

1. Recommends the Council approve the inclusion of new archipelagic bycatch summaries that describe both the amount and type of bycatch in Hawaii's bottomfish fisheries in the fishery performance module of the Hawaii Archipelago annual SAFE report.

Regarding the development of the territorial non-commercial modules for the American Samoa and Mariana Archipelago annual SAFE reports, the APT

2. Recommends the Council request NMFS PIFSC continue its effort to develop the territorial non-commercial module and related R scripts for approval and inclusion in the annual SAFE reports for 2023, noting that other time series data streams (e.g., commercial receipt book) may also be updated in pursuit of a single data summarization and/or expansion process for the Western Pacific region.

Regarding the draft Hawaii non-commercial module, the APT

3. Recommends the Council approve the inclusion of the draft Hawaii non-commercial module based on HMRFS data into the Hawaii Archipelago annual SAFE report as presented, noting that additional investigation is needed to determine if there may be biases in the interview-derived data.

Regarding the refinement of uku EFH in the MHI, the APT

4. Recommends the Council select Option 5 to refine the EFH designation for uku in the Hawaii Archipelago FEP based on an overlay of Level 1 and 2 modeling products alongside fishery-dependent CPUE data. The APT noted that there may also be forthcoming information on the spatial distribution of egg and post-hatch pelagic life stages of uku for further refinement of the EFH designations for the species in the next one to three years.

Regarding the establishment of SDC for MHI Kona crab, the APT

5. Recommends the Council select Alternative 2 to establish SDC for Kona Crab in the Hawaii Archipelago FEP based on the SDC utilized in the previous stock assessment (Kapur et al. 2019) and NMFS technical guidance (Restrepo et al. 1998).

Regarding the territorial BMUS revision, the APT

6. Recommends the Council select Alternative 2 to revise the American Samoa BMUS list in the American Samoa FEP based on the results of the hierarchical cluster analysis by PIFSC, a review of the ten non-exhaustive factors for determining which species require federal conservation and management as specified in National Standard 1, and the life history synthesis, as well as the five related Magnuson-Stevens Act management components (i.e., SDC, ACLs/AMs, EFH, monitoring and bycatch, and fishing communities) based on the generation of MSA component reports developed by the APT. The APT agreed to move forward with territorial BMUS revisions in alignment with the

current schedule stock assessments for each island area such that the list revisions will occur separately for each jurisdiction.

Regarding CNMI BMUS ACL specifications, the APT

7. Recommends the Council select Option 3 that would retain the previous risk of overfishing of 39% based on the previous P* analysis, associated with an ACL of 82,000 lb and an ACT of 75,000 lb for 2024-2025. The APT noted that the risk of overfishing was presented by the SSC and Council through their standardized P* and SEEM processes, though these processes are subject to change based on the availability of new fishery information.

Regarding Kona crab ACL specifications, the APT

8. Recommends the Council select Option 2 that would rollover the previous ACL of 30,802 lb alongside an ACT of 25,491 lb for 2024-2025, maintaining the risk of overfishing of 38% and 20%, respectively, from the previous P* and SEEM evaluations. The APT noted that the current ACT of 25,491 lb have not been reached since their implementation in 2020 and are unlikely to in the next two years.

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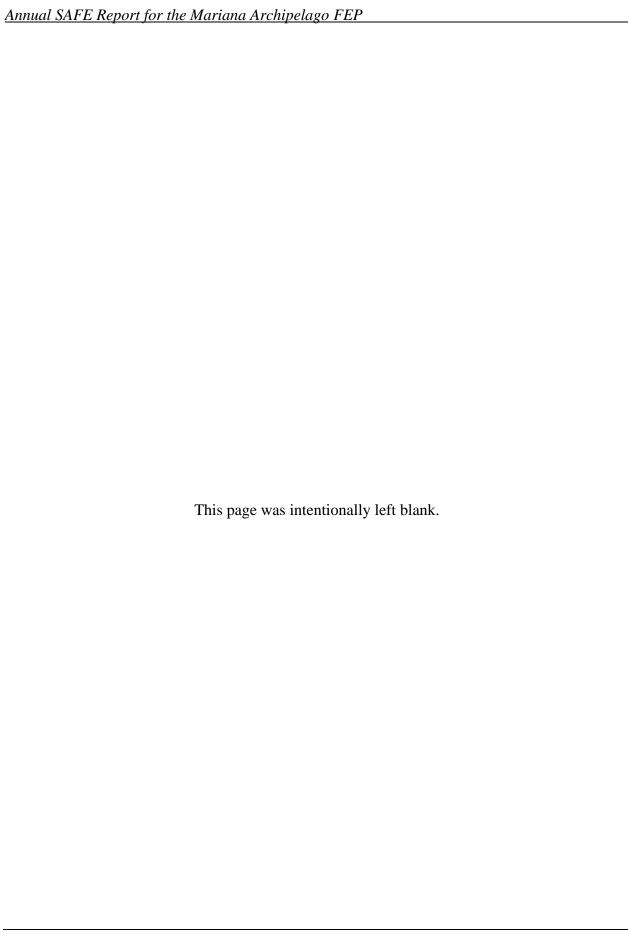
ACRONYMS AND ABBREVIATIONS

| Acronym | Meaning |
|---------------------|--|
| A ₅₀ | Age at 50% Maturity |
| $A\Delta_{50}$ | Age at 50% Sex Reversal |
| ABC | Acceptable Biological Catch |
| ACE | Accumulated Cyclone Energy |
| ACL | Annual Catch Limits |
| ACT | Annual Catch Target |
| AM | Accountability Measure |
| AVHRR | Advanced Very High Resolution Radiometer (NOAA) |
| В | Biomass |
| BE | Biological Evaluation |
| B_{FLAG} | Reference point indicating low biomass |
| BiOp | Biological Opinion |
| BMUS | Bottomfish Management Unit Species |
| BRFA | Bottomfish Restricted Fishing Areas |
| BSIA | Best Scientific Information Available |
| CFEAI | Commercial Fishing Economic Assessment Index |
| CFR | Code of Federal Regulations |
| CMAP | Merged Analysis of Precipitation (CPC) |
| CMUS | Crustacean Management Unit Species |
| CNMI | Commonwealth of the Northern Mariana Islands |
| CO-OPS | Center for Operational Oceanographic Products and Services |
| | (NOAA) |
| CPC | Climate Prediction Center (NOAA) |
| CPI | Consumer Price Index |
| CPUE | Catch per Unit Effort |
| CRED | Coral Reef Ecosystem Division (PIFSC) |
| CREP | Coral Reef Ecosystem Program (PIFSC) |
| CREMUS | Coral Reef Ecosystem Management Unit Species |
| CRW | Coral Reef Watch (NOAA) |
| CV | Coefficient of Variation |
| DAR | Division of Aquatic Resources (Hawaii) |
| DAWR | Division of Aquatic and Wildlife Resources (Guam) |
| DFW | Division of Fish and Wildlife (CNMI) |
| DGI | Daily Growth Increments |
| DHW | Degree Heating Weeks |
| DIC | Dissolved Inorganic Carbon |
| DMWR | Department of Marine and Wildlife Resources (American Samoa) |
| DOD | Department of Defense |
| DOJ | Department of Justice |
| DON | Department of Navy |
| DPS | Distinct Population Segment |
| E | Effort |
| EA | Environmental Assessment |
| | Zii i Ziiiiiiiiiiii i issoosiiiiiii |

| Acronym | Meaning |
|------------------|--|
| EBFM | Ecosystem Based Fisheries Management |
| ECS | Ecosystem Component Species |
| EEZ | Exclusive Economic Zone |
| EFH | Essential Fish Habitat |
| EIS | Environmental Impact Statement |
| ENSO | El Niño - Southern Oscillation |
| EO | Executive Order |
| ESA | Endangered Species Act |
| ESRL | Earth Systems Research Laboratory (NOAA) |
| F | Fishing Mortality |
| FAD | Fish Aggregating Device |
| FL | Fork Length |
| FDM | Farallon de Medinilla |
| FEP | Fishery Ecosystem Plan |
| FMP | Fishery Management Plan |
| FR | Federal Register |
| FRMD | Fisheries Research and Monitoring Division (PIFSC) |
| FSM | Federated States of Micronesia |
| FSWP | Fisheries Statistics of the Western Pacific |
| GFCA | Guam Fisherman's Cooperative Association |
| GLM | General Linear Modeling |
| GOES | Geostationary Operational Environmental Satellite (NOAA) |
| GPS | Global Positioning System |
| Н | Harvest |
| HAPC | Habitat Area of Particular Concern |
| HOT | Hawaii Ocean Time Series (UH) |
| HURL | Hawaii Undersea Research Laboratory (NOAA and UH) |
| k | von Bertalanffy Growth Coefficient |
| L ₅₀ | Length at 50% Maturity |
| $L\Delta_{50}$ | Length at 50% Sex Reversal |
| L_{∞} | Asymptotic Length |
| L _{bar} | Mean Fish Length |
| L_{max} | Maximum Fish Length |
| LAA | Likely to Adversely Affect |
| LOC | Letter of Concurrence |
| LOF | List of Fisheries |
| M | Natural Mortality |
| MBTA | Migratory Bird Treaty Act |
| MCP | Marine Conservation Plan |
| MFMT | Maximum Fishing Mortality Threshold |
| MHI | Main Hawaiian Islands |
| MITT | Mariana Islands Training and Testing |
| MLCD | Marine Life Conservation District |
| MMA | Marine Managed Area |
| MMPA | Marine Mammal Protection Act |

| Acronym | Meaning |
|--------------|--|
| MODIS | Moderate Resolution Imaging Spectroradiometer (NASA) |
| Monument | Marianas Trench Marine National Monument (also MTMNM) |
| MOU | Memorandum of Understanding |
| MPA | Marine Protected Area |
| MPCC | Marine Planning and Climate Change |
| MPCCC | MPCC Committee (WPRFMC) |
| MSA | Magnuson-Stevens Fishery Conservation and Management Act |
| MSL | Mean Sea Level |
| MSST | Minimum Stock Size Threshold |
| MSU | Microwave Sounding Unit |
| MSY | Maximum Sustainable Yield |
| MUS | Management Unit Species |
| n | Sample Size |
| N_{L-W} | Sample Size for Length-Weigh Regression |
| N/A | Not Applicable |
| NAF | No Active Fishery |
| NASA | National Aeronautics and Space Administration |
| NCADAC | National Climate Assessment and Development Advisory |
| | Committee |
| NCDC | National Climatic Data Center (NOAA) |
| ND | Not Detected |
| NEPA | National Environmental and Policy Act |
| NESDIS | National Environmental Satellite, Data, and Information Service (NOAA) |
| NLAA | Not Likely to Adversely Affect |
| NMFS | National Marine Fisheries Service (NOAA) |
| NMI | Northern Marina Islands |
| NMS | Non-metric Multidimensional Scaling; or National Marine |
| 11110 | Sanctuary |
| NOAA | National Oceanic and Atmospheric Administration |
| NOI | Notice of Intent |
| NOS | National Ocean Service (NOAA) |
| NPDES | National Pollutant Discharge Elimination System |
| NS | National Standard |
| NTM | Notice to Mariners |
| NWHI | Northwestern Hawaiian Islands |
| NWS | National Weather Service |
| OEIS | Overseas Environmental Impact Statement |
| OFL | Overfishing Limits |
| ONI | Ocean Niño Index |
| ONMS | Office of National Marine Sanctuaries (NOAA) |
| ONMS | OLR Precipitation Index (NOAA) |
| OLR | |
| | Outgoing Longwave Radiation |
| OY DCOP 1 | Optimum Yield Pendamia Condition of Pandinass 1 |
| PCOR 1 | Pandemic Condition of Readiness 1 |

| Acronym | Meaning | | |
|-------------|---|--|--|
| PCMUS | Precious Coral Management Unit Species | | |
| PDO | Pacific Decadal Oscillation | | |
| Pelagic FEP | Fishery Ecosystem Plan for the Pacific Pelagic Fisheries | | |
| PIAFA | Pacific Insular Area Fishery Agreement | | |
| PIFSC | Pacific Island Fisheries Science Center (NMFS) | | |
| PIRCA | Pacific Islands Regional Climate Assessment | | |
| PIRO | Pacific Islands Regional Office (NMFS) | | |
| PMEL | Pacific Marine Environmental Laboratory (NOAA) | | |
| PMUS | Pelagic Management Unit Species | | |
| POES | Polar Operational Environmental Satellite (NOAA) | | |
| PRIA | Pacific Remote Island Areas | | |
| RAMP | Reef Assessment and Monitoring Program (CRED) | | |
| ROD | Record of Decision | | |
| RPB | Regional Planning Body | | |
| SAFE | Stock Assessment and Fishery Evaluation | | |
| SBRM | Standardized Bycatch Reporting Methodologies | | |
| Secretary | Secretary of Commerce | | |
| SEEM | Social, Economic, Ecological, Management (Uncertainty) | | |
| SEIS | Supplemental Environmental Impact Statement | | |
| SFA | Sustainable Fisheries Act or Saipan's Fishermen Association | | |
| SFD | Sustainable Fisheries Division (PIRO) | | |
| SODA | Simple Ocean Data Assimilation | | |
| SPC | Stationary Point Count | | |
| SSC | Scientific and Statistical Committee (WPRFMC) | | |
| SSM/I | Special Sensor Microwave/Imager | | |
| SST | Sea Surface Temperature | | |
| SSBPR | Spawning Stock Biomass Proxy Ratio | | |
| SUA | Special Use Airspace | | |
| t_0 | Hypothetical Age at Length Zero | | |
| T_{max} | Maximum Age | | |
| TA | Total Alkalinity | | |
| TALFF | Total Allowable Level of Foreign Fishing | | |
| TBA | To Be Assigned | | |
| TBD | To Be Determined | | |
| TSI | Territory Science Initiative | | |
| UH | University of Hawaii | | |
| USAF | United States Air Force | | |
| USFWS | United States Fish and Wildlife Service | | |
| VBGF | von Bertalanffy Growth Function | | |
| VFP | Visual Fox Pro | | |
| WPacFIN | Western Pacific Fishery Information Network | | |
| WPR | Western Pacific Region | | |
| WPRFMC | Western Pacific Regional Fishery Management Council | | |
| WPSAR | Western Pacific Stock Assessment Review | | |



1 FISHERY PERFORMANCE

1.1 CNMI FISHERY DESCRIPTIONS

1.1.1 Background

The Commonwealth of the Northern Mariana Islands (CNMI) is a chain of islands in the Western Pacific Ocean. Along with the island of Guam, the chain is historically known as the Mariana Islands. The CNMI consists of 14 small islands situated in a north-south direction, stretching a distance of about 500 km. The surrounding waters of the CNMI play an integral role in the everyday lives of its citizens. The ocean is a major source of food and leisure activities for residents and tourists alike. Archeological research has also revealed evidence of fishing activities in the CNMI dating back 3,000 years. Although the composition of fishing activities in the Marianas has changed significantly since then, a common view of its importance remains.

Fisheries during the German occupation

During the German occupational period (1899-1914) a majority of the economic focus in the Northern Marianas was on the copra industry. Few commercial fisheries were noted during this period of time, as the German administration focused efforts on crop production and feral cattle trade (Russell 1999). Chamorro and Carolinians utilized the protected lagoon and open waters with several fishing methods: talaya (cast net), chinchulu (surround net), gigao (fish weir), tokcha (spear), tupak (hook and line), and Carolinians additionally gleaned sea cucumbers for the Asian Markets. Most of these activities were for subsistence purposes, with the catch being distributed and bartered among relatives and acquaintances.

Fisheries during the Japanese occupation

Fisheries development prospered during the Japanese administration (1914-1945), becoming the nation's second largest industry. Small pelagic fishing operations were established and the Garapan port became the main area for drying fish. Large scale fishing activities occurred during the 1930s, shown as Saipan produced 11 percent of total tuna landed in Micronesia (Bowers 2001). However, efforts to develop the tuna fishery shifted to Palau and Federated States of Micronesia (FSM) due to the availability of bait fish in the region. Subsistence fishing still persisted within the lagoon and fringing reefs and was mainly conducted by the natives though a large extraction of sea cucumbers did occur. There were several main fishing methods used during this period: cast net, spear, gill net, surround net, hook and line, and gleaning. During this period, the topshell (*Trochus niloticus*) was also introduced into the Marianas.

Fisheries during the U.S. military occupation

The fishing industry was destroyed during World War 2, but quickly rebuilt afterwards with support from the U.S. military. Okinawans who operated the fishery prior to the war were hired to operate and train locals to fish commercially, targeting pelagic species. A company called Saipan Fishing Company operated during this time and contributed to the early re-development of post-war commercial fisheries in the CNMI (Bowers 2001). Most of the fishing activities were for *Katsuwanus pelamis* (bonito) and other tuna species. However, other resources, such as bigeye scad, reef fish, and lobster, were also harvested during calm weather. The Chamorro and Carolinians continued subsistence fishing in the lagoon after the war. Although limited quantities of monofilament nets were available during this period, they were used to capture lagoon fish

and along the reef lines. The use of modern fishing gear such as masks, rubber fins, and flashlights made it much easier to harvest coral reef resources during this time.

Fisheries activities within the past two decades

The CNMI has had numerous changes in its fisheries over the past twenty years. In the mid-1990s, commercial fishing activities increased significantly. Commercial SCUBA fishing became a common method, not only to support local demand for reef fish, but to bolster exports to Guam as well. Large-scale commercial bottomfish fishing in the Northern Islands of the CNMI peaked starting in the mid-1990s through 2002, with landings being both sold locally and exported to Japan. Troll fishing continued to be dominant during this period. An exploratory, deepwater shrimp fishery also developed, but did not last due to internal company issues and gear losses. Around this time, a sea cucumber fishery also began on Rota before migrating to Saipan; ultimately, however, this fishery was found to be unstable and was subsequently halted.

Several fishing companies entered the fisheries only to close down a few years later. The CNMI reached its highest population during the last two decades, most of whom have been migrant workers from Asia. The tourism industry has also been increasing, which contributes to high demand for fresh fish. Subsistence fishing within the nearshore waters of Saipan, Tinian, and Rota has also increased.

In the 2000s, small-scale troll, bottom and reef fish fisheries persisted, with landings sold locally. Federal and state support was provided multiple times to further develop fisheries in the CNMI with intermittent success. An exploratory longline fishery was funded and operated in the CNMI in the mid-2000 for about two years, but eventually closed down due to low productivity of high-value, pelagic fish, among other issues within the business. A few larger (40-80') bottomfish fishing vessels were also operational during this period, with a majority of them fishing the northern islands and offshore banks. A few of these vessels were recipients of financial assistance to improve their fishing capacities.

Fisheries in the CNMI have generally been relatively small and fluid, with 16-20' boats fishing within 20 miles from Saipan. Many of these small vessels conduct multiple fishing activities during a single trip. For example, a company that is supported mainly by troll fishing may also conduct bottomfish fishing and spearfishing to supplement their income. Fishing businesses tend to enter and exit the fishery when it is economically beneficial to do so, as they are highly sensitive to changes in the economy, development, population, and regulations. Subsistence fishing continues; however, fishing methods and target species have shifted in step with population demographics and fishery restrictions. Nearshore hook and line, cast net, and spear fishing are common activities, but fishing methods such as gill net, surround net, drag net, and SCUBA-spear have been restricted or outright banned in the CNMI since the early 2000s.

1.1.1.1 Bottomfish Fishery

The bottomfish fishery has not changed much from its early years in certain aspects. Relatively small (<25 ft.) fishing vessels are still being used to access bottom fishing grounds around Saipan and Tinian, while the larger (>25 ft.) vessels are used to access bottomfish resources in the Northern Islands. Only a handful of these larger bottom fishing vessels are operating within the CNMI. Most of the small bottomfish fishing vessels are owned by commercial purchasers; there are, however, a few subsistence bottomfish fishermen that participate in the fishery

intermittently. More recently, improved technologies, such as sophisticated electronics to locate fish and various types of reels replacing handlines, have entered the CNMI bottomfish fishery.

Two distinct types of bottomfish fisheries are identified in the CNMI: shallow-water bottom fishing, which targets fish at depths down to 150 m, and deepwater bottom fishing, which targets fish at depths greater than 150 m. Species targeted by the shallow-water fishery consist of the Redgill Emperor (*Lethrinus rubrioperculatus*), Black Jack (*Caranx lugubris*), Matai (*Epinephelus fasciatus*), Sas (*Lutjanus kasmira*), and Lunartail Grouper (*Variola louti*), among other fish residing at similar depths. Species targeted by the deepwater bottom fishing depths (>150m) include onaga (*Etelis corsucans*), ehu (*E. carbunculus*), yellowtail kalekale (*Pristipomiodes auricilla*), amberjack (*Seriola dumerili*), blueline gindai (*P. argyrogrammicus*), gindai (*P. zonatus*), opakapaka (*P. filamentosus*), and eightbanded grouper (*Hyporthordus octofasciatus*), among other fish residing at similar depths.

Bottomfish management unit species (BMUS) are not the only species caught in the shallow-bottom fishery. Deep-water bottomfish fishing requires more efficient fishing gears, such as hydraulic reels. Bottomfish fishing trips generally return during the day, but there is an unmeasured amount that occurs outside of survey hours from 2 AM to 10 AM. Fishing trips to the Northern Islands can take two to four days depending on vessel size and refrigeration capacity. These trips are most productive during calm weather months. Successful fishermen targeting deep-water bottomfish tend to fish for one to four years before leaving the fishery, whereas the majority of fishermen targeting shallow-water bottomfish tend to leave the fishery after the first year.

The overall participation of fishermen in the bottomfish fishery tends to occur on a relatively short-term basis (i.e., less than four years). The slight difference between shallow-water and deepwater fishermen likely reflects the greater skill and investment required to participate in the deepwater bottomfish fishery. In addition, deepwater bottomfish fishing tends to include larger ventures that are more buffered from the impulses of individual choice and are usually dependent on a skilled captain and fishermen. Overall, the long-term commitment to hard work, maintenance and repairs, and staff retention appear to be challenging for CNMI bottomfish fishermen to sustain their efforts for more than a few years. A full list of BMUS species is provided in Appendix A.

1.1.1.2 Coral Reef Fishery

Coral reef fisheries have been generally steady in recent years relative to previous decades. Small-scale nearshore fisheries in the CNMI continue to be important socially, culturally, recreationally, financially, and for subsistence. Most fishermen are subsistence fishers, with a number of them selling a portion of their catch to roadside commercial purchasers and some of these commercial purchasers employing the fishermen to maintain a constant supply of reef fish. Most of the fishing for coral reef species occurs within the Saipan lagoon and fringing reefs around the islands, targeting mainly finfish and invertebrates. All reef fish catches are sold to local markets or used for personal consumption with a minimal portion exported for off-island residents. Shoreline access is the most common way to harvest coral reef resources. Vessels are generally used during calm weather to fish areas not as accessible other times of the year, with fishing trips to other islands being made when the weather is favorable. Fishing methods have not changed significantly compared to previous years; hook and line, cast netting, spear fishing, and gleaning are methods still being used today. Some of the common families found in the

CNMI reef fish markets are Acanthuridae (surgeonfish), scaridae (parrotfish), mullidae (goatfish), serranidae (grouper), labridae (wrasse), holocentridae (soldier/squirrelfish), carangidae (jacks), scombridae (scad), haemulidae (sweetlips), gerridae (mojarra), kyphosidae (rudderfish), and mugilidae (mullet), as well as other non-finfish families.

In 2018, the Council drafted an Amendment 5 to the Mariana Archipelago Fishery Ecosystem Plan (FEP) that reclassified a large number of management unit species (MUS) as Ecosystem Component Species (ECS; WPRFMC 2018). The final rule was published in the *Federal Register* in early 2019 (84 FR 2767, February 8, 2019). This amendment reduces the number of MUS from 227 species and families to 13 in the Mariana Archipelago FEP. All former coral reef ecosystem MUS (CREMUS) and crustacean MUS (CMUS) were reclassified as ECS that do not require annual catch limit (ACL) specifications or accountability measures but are still to be monitored regularly to prioritize conservation and management efforts and to improve efficiency of fishery management in the region. All existing management measures, including reporting and record keeping, prohibitions, and experimental fishing regulations apply to ECS. If an ECS stock becomes a target of a federal fishery in the future, the National Marine Fisheries Service (NMFS) and the Council may consider including that stock as a MUS to actively manage that stock. These species are still regularly monitored via other means (see Sections 1.1.5.3 and 2.2).

1.1.2 Fishery Data Collection System

A majority of the information collected by the CNMI Division of Fish and Wildlife (DFW) is fishery dependent. Since the early-1980s, attempts were made to establish a data collection program for the nearshore fisheries but failed due to intergovernmental issues. Over recent decades, significant time and effort has been made to further develop nearshore fishery data collection. This effort has resulted in the re-establishment of the shore-based creel survey program by DFW in collaboration with other local and federal agencies. To further improve data collection efforts, the CNMI instituted mandatory data submission for commercial fisheries. The CNMI is working on improving commercial licensing and data submission processes to meet recent data collection mandates. The CNMI is working with NOAA to further improve this mandate through exploring alternative fishery data collection programs.

1.1.2.1 Creel Surveys

Currently the CNMI maintains both a boat- and shore-based creel survey for the island of Saipan, with plans for expansion to the populated neighboring islands. The programs were established in 2000 and 2005, respectively, in order to strengthen the capacity of DFW in providing sufficient information to the public regarding local fisheries. Other programs, such as the invoicing system and importation monitoring, provide supplemental information on harvest and demand for the fishery.

Effective management of Saipan's marine fishery resources requires the collection of fishing effort, methods used, and harvest. The CNMI boat- and shore-based creel surveys are some of the major data collection systems used by DFW to estimate the total annual boat-based participation, effort, and harvest while surveying nearshore fishery resources. These surveys were formerly known as the "CNMI offshore creel survey" but are now referred to as "boat-based" because they cover all fishing done from a boat (including non-registered makeshift boats and kayaks since October 2022). This is an important distinction because where the fishing activity is initiated (i.e., boat vs. shore) determines how that type of activity will be accounted for in the survey systems. For instance, very small boats launched from non-standard launching

areas (e.g., from the back of a pickup truck on a beach) are not included in the boat-based creel survey.

The objective of the boat-based creel survey program is to quantify fishing participation, effort, and catch done from on a vessel in CNMI's waters. DFW had an early creel survey data collection program in 1984, and 1990 to 1994, however since the methods were not standardized, the data collected with that early program is not currently being used. The early program was eventually terminated due to a lack of resources. On April 2, 2000, the DFW fishery staff reinitiated the boat-based creel survey program on the island's boat-based fishery following a three-year hiatus. The fishery survey collects data on the island's boating activities and interviews returning commercial and noncommercial fishermen at the three most active launching ramps/docks on the island: Smiling Cove, DFW Ramp, and Fishing Base. DFW Ramp had recently been added as an active launching site as of August 2022 to replace Sugar Dock as it has not been active for quite some time due to the lack of launching access. Essential fishery information is collected and processed from both commercial and noncommercial vessels to help better inform management decisions. The two types of data collection programs utilized by Saipan's boat-based creel survey program include: boat-based participation count to collect participation data, and a boat-based access point survey to collect catch and effort data (through survey maps, boat logs, and interviews) at the three major boat ramp areas listed above. The data collected are then expanded at a stratum level (quarterly vs. annually, charter vs. non-charter, weekday vs. weekend, etc.) to create estimated landings by gear type for CNMI's boat-based fishery.

The shore-based survey currently covers the western lagoon of Saipan. Some pilot surveys were conducted on Saipan's Eastern beaches such as Laolao Bay, Obyan Beach, and Ladder Beach. However, effort to collect data from these areas have been very sporadic. Other accessible areas are not covered at this time due to existing limited resource availability and logistical constraints. With the assistance of the Fisheries Research and Monitoring Division (FRMD) at the Pacific Islands Fisheries Science Center (PIFSC), data processing software and a database were developed to process these survey data.

In May 2005, DFW fishery staff reinitiated the creel survey program for the island's shore-based fishery following an eleven year hiatus. The western lagoon starts from the northwest (Wing Beach) and extends to the southwest (Agingan Point) of Saipan, encompassing over twenty accessible and highly active shoreline access points. Saipan's shore-based creel survey is also a stratified, randomized data collection program. This program collects two types of data to estimate catch and effort information in the shore-based fishery: participation counts and interviews. The participation counts involve counting the number of people fishing on randomly selected days and their method of fishing along the shoreline. The interviews involve some dialog with fishermen to determine catch, method used, length and weights of fish, species composition, catch disposition, and if any fish were not kept (i.e., bycatch). The data collected from this program have been used to expand and create annual estimated landings for the shorebased fishery in the CNMI.

On August 2015, Saipan was impacted by Typhoon Soudelor and on October 2018, Saipan and Tinian were directly impacted by Super Typhoon Yutu. The damage inflicted by the typhoon delayed both creel surveys, collection of commercial receipt invoices and data entry. About a month after the typhoon, creel surveys were regularly conducted again, and boat-based surveys

followed soon thereafter. commercial purchasers prioritized repairing typhoon-related damages to their businesses, and the number of invoices collected decreased as a result.

In March 2020, the CNMI issued community restrictions to address the COVID-19 pandemic concerns. A number of measures, such as curfew restrictions, gathering restrictions, sanitation restrictions, office closures, travel restrictions, as well as fishing restrictions, were implemented during this time. These restrictions were reduced as the COVID -19 situation improved in the CNMI. This also significantly affected commerce within the CNMI as tourism is the main source of income. Fishing activities and businesses gradually opened back up as the situation improved. Participation increased as people entered the fishery due to being displaced by the pandemic and turned to fishing for alternative income. The DFW fishery data collection program activities were also limited by the COVID-19 restrictions. Sample days and hours were limited during the first few months of the restrictions. As restrictions were lifted, sampling effort for all fishery data collection programs increased and coverage improved.

There were 51 boat-based surveys conducted between January 1 and December 31, 2022. A total of 170 regular interviews and 7 opportunistic interviews were completed with an expanded catch of 47,417 lbs. The vessel/trailer participation survey is ongoing and includes all launching areas on the west coast of Saipan, where all boat-based fishing occurs. No vessels were marked as "out fishing" for several years prior to 2022.

1.1.2.2 Purchaser Invoice

The DFW has been collecting fishery statistics on Saipan's commercial fishing fleet since the mid-1970s. With the assistance of NMFS, the DFW also expanded its fisheries monitoring programs to include the other two major inhabited islands in the CNMI: Rota and Tinian. The DFW's principal method of collecting domestic commercial fisheries data is an invoicing system, sometimes referred to as a "trip ticket" system. The DFW provides numbered two-part invoices to all commercial purchasers of fresh fishery products (including hotels, restaurants, stores, fish markets, and roadside vendors). commercial purchasers then complete an invoice each time they purchase fish directly from fishers; one copy goes to the DFW, and one copy goes to their records. Some advantages of this data collection method are that it is relatively inexpensive to implement and maintain, and it is fairly easy to completely cover the commercial fisheries. The DFW can also provide feedback to commercial purchasers and fishers to ensure data accuracy and continued cooperation over time.

There are some disadvantages to the trip ticket system, including: (1) dependency on non-DFW personnel to identify the catch and record the data, (2) restrictions on the types of data that can be collected, (3) required education and cooperation of all fish commercial purchasers, and (4) limited recordings of fish actually sold to commercial purchasers. Therefore, a potentially important portion of the total landings typically goes unrecorded. Since 1982, the DFW has tried to minimize these disadvantages in several ways by (1) maintaining a close working relationship with commercial purchasers, (2) adding new commercial purchasers to their list and educating them, and (3) implementing a creel survey to help estimate total catch (including recreational and subsistence portion). The current system collects data from commercial purchasers in Saipan, where the DFW estimates more than 90 percent of all CNMI commercial landings are made. The DFW also estimates that the proportion of total commercial landings that have been recorded in the Saipan database since 1983 is about 90 percent; however, coverage has been relatively mottled over the years. Previous volumes of FSWP reported only recorded landings, but in recent

volumes, the data have been adjusted to represent 100 percent coverage and are referenced as "estimated commercial landings" in the tables and figures.

These data elements are collected for all purchases of fishery products; however, species identification is frequently identified only to a group level, especially for reef fish.

1.1.2.3 Bio-Sampling

The bio-sampling database contains general and specific bio-data obtained from individual commercial spearfish catches landed on Saipan from six different commercial purchasers over the course of 2011. The following data was captured for each fishing trip sampled: date, fishing gear type, time/hours fished, location fished, number/names of fishers, lengths/weights of individual fish, number/weight of octopus and squid, number/carapace size/weight/sex of lobster, and whether it was boat- or shore-based fishing trip.

Although sampling effort was intended to be spread evenly among all participating commercial purchasers, smaller commercial purchasers were inherently much more difficult to sample within the time constraints allowed. Therefore, a regular sampling schedule was implemented for the island's two largest commercial purchasers that included two weekdays and one weekend day each week starting in January-February 2011. Problems encountered in sampling the smaller commercial purchasers included: more days in any given month where no fish were purchased, the work area was not conducive for sampling, and communication problems. The bio-sampling database focuses on nighttime (non-SCUBA) spearfishing activities. Due to purchaser-imposed limitations, other gear types that typically land their catch during normal business hours were not sampled.

1.1.2.4 Exemption Netting

In 2003, the use of gill nets was prohibited in the CNMI. In 2005, the DFW decided to allow gill netting under special circumstances. Gill netting is now allowed under strict conditions provided by the DFW with their permission such that all gill netting activities are to be monitored and recorded by DFW personnel.

In 2010, a law was passed allowing for the use of gill nets for the purpose of subsistence on the island of Rota. The following year, a regulation allowing subsistence net fishing was passed for the island of Tinian.

For a majority of the permitted gillnet activities, length and weight measurements were taken at the fishing site. Fork lengths were measured in millimeters and weights were measured in grams. If time did not permit for individual measurements, then length measurements were taken for each fish and total weight was taken for each species. Length/weight ratios were used to estimate weights of sampled fish. Information has been collected for activities conducted on the island of Saipan, but no official collection of information has been collected for Rota or Tinian.

1.1.2.5 Life History

The CNMI DFW life history program began in 1996 sampling the redgill emperors (*Lethrinus rubrioperculatus*). Since then, sampling has been conducted on other species, including *A. lineatus*, Myriprestinae (*Myripristis violacea*, *M. kuntee*, *M. pralinea*, *M. bernti*, *M. murdjan*), *L. harak*, *Naso lituratus*, *Chlorurus sordidus*, and *C. undulatus*. Other life history programs have also developed over the past years. In collaboration with NMFS, DFW personnel collect life history information on *Scarus rubroviolaceus*, *Lethrinus atkinsoni*, and *Parupeneus barbarinus*

through funding provided by NOAA-NMFS. The life history survey captures biological information, including reproductive cycle, age at length, and age at maturity. The DFW is continually working to improve the understanding of reef fish life history in the CNMI through these types of programs.

1.1.2.6 Monitoring of Imported Fish

The DFW Fisheries Data Sections collect fisheries-related importation invoices from the Department of Customs at the end of every month. The data is then entered into a ticket receipt system and reviewed prior to being sent out for compilation by PIFSC. Most of the information entered into the system can only be identified to the family taxa.

1.1.2.7 Vessel Inventory

The most recent records obtained from CNMI Department of Public Safety (DPS) are from 2022. Their records are hand-written and do not exist electronically. 340 vessels (including jetskis) were registered. 107 vessels were registered as commercial fishing vessels. 297 were registered for personal use, although an unknown amount was and continue to be used for commercial fishing regardless of their intended use specified on the registration. Others were registered for commercial recreation and government use. This work is also impacted by policies of the DPS, which manages vessel licensing. Going forward, additional emphasis on continuing to improve the vessel inventory project, especially once the open data technicians and data manager positions are filled at the CNMI DFW.

1.1.3 Meta-Data Dashboard Statistics

The meta-data dashboard statistics describe the amount of data used or available to calculate the fishery-dependent information. Creel surveys are sampling-based systems that require a random-stratified design applied to pre-scheduled surveys. The number of sampling days, participation runs, and catch interviews can be used to determine if there are sufficient samples to run the expansion algorithm. The trends of these parameters over time may infer survey performance. Monitoring the survey performance is critical for explaining the reliability of the expanded information.

Commercial receipt book information depends on the number of invoices submitted and the number of vendors participating in the program. Variations in these meta-data affect the commercial landing and revenue estimates.

1.1.3.1 Creel Survey Meta-Data Statistics

Calculations:

Sample days: Count of the total number of unique dates found in the boat log sampling date data in boat-based creel surveys.

Catch Interviews: In boat-based creel surveys, count of the total number of data records found in the interview header data (number of interview headers). This is divided into two categories, interviews conducted during scheduled survey days (Regular) and opportunistic interviews (Opportunistic), which are collected on non-scheduled days.

Table 1. Summary of CNMI boat-based creel survey meta-data

| Year # Sample Days | # Catch Interviews |
|--------------------|--------------------|
|--------------------|--------------------|

| | | Regular | Opportunistic |
|------------|----|---------|---------------|
| 2000 | 44 | 168 | 9 |
| 2001 | 67 | 285 | 0 |
| 2002 | 75 | 200 | 25 |
| 2003 | 90 | 299 | 40 |
| 2004 | 77 | 272 | 16 |
| 2005 | 78 | 417 | 29 |
| 2006 | 71 | 342 | 22 |
| 2007 | 62 | 314 | 1 |
| 2008 | 55 | 250 | 1 |
| 2009 | 64 | 241 | 25 |
| 2010 | 65 | 161 | 82 |
| 2011 | 67 | 162 | 87 |
| 2012 | 72 | 166 | 0 |
| 2013 | 71 | 191 | 0 |
| 2014 | 71 | 166 | 0 |
| 2015 | 57 | 119 | 2 |
| 2016 | 65 | 117 | 3 |
| 2017 | 66 | 120 | 6 |
| 2018 | 54 | 126 | 1 |
| 2019 | 33 | 65 | 8 |
| 2020 | 58 | 126 | 52 |
| 2021 | 69 | 205 | 51 |
| 2022 | 51 | 170 | 7 |
| 10-yr avg. | 60 | 141 | 13 |
| 10-yr SD | 11 | 40 | 19 |
| 20-yr avg. | 65 | 201 | 22 |
| 20-yr SD | 12 | 88 | 27 |

1.1.3.2 Commercial Receipt Book Statistics

Calculations:

Vendors: Count of the number of unique buyer codes found in the commercial purchase header data from the Commercial Receipt Book; BMUS vendors are only from vendors that landed BMUS species.

Invoices: Count of the number of unique invoice numbers found in the commercial header data from the Commercial Receipt Book; BMUS vendors are only from vendors that landed BMUS species.

Table 2. Summary of CNMI commercial receipt book meta-data

| Year | # Vendors | # Invoices Collected | # BMUS Vendors | # BMUS Invoices Collected |
|------|-----------|-------------------------|-------------------|---------------------------------|
|------|-----------|-------------------------|-------------------|---------------------------------|

| Year | # Vendors | # Invoices Collected | # BMUS Vendors | # BMUS Invoices Collected |
|------|-----------|-------------------------|-------------------|---------------------------------|
| 1983 | 42 | 2,930 | 13 | 55 |
| 1984 | 45 | 3,452 | 11 | 50 |
| 1985 | n.d. | n.d. | n.d. | n.d. |
| 1986 | n.d. | n.d. | n.d. | n.d. |
| 1987 | 27 | 1,908 | 11 | 30 |
| 1988 | 16 | 2,204 | 7 | 23 |
| 1989 | 24 | 2,454 | 8 | 51 |
| 1990 | 23 | 2,218 | 5 | 19 |
| 1991 | 30 | 2,240 | 4 | 16 |
| 1992 | 55 | 3,233 | 3 | 4 |
| 1993 | 48 | 3,426 | 15 | 53 |
| 1994 | 55 | 3,722 | 17 | 89 |
| 1995 | 61 | 4,637 | 21 | 167 |
| 1996 | 73 | 5,870 | 25 | 231 |
| 1997 | 56 | 4,920 | 20 | 171 |
| 1998 | 53 | 6,374 | 21 | 220 |
| 1999 | 52 | 5,771 | 21 | 213 |
| 2000 | 49 | 6,892 | 16 | 210 |
| 2001 | 42 | 5,820 | 19 | 431 |
| 2002 | 33 | 5,611 | 17 | 268 |
| 2003 | 27 | 4,726 | 14 | 172 |
| 2004 | 25 | 3,720 | 13 | 99 |
| 2005 | 24 | 4,245 | 11 | 116 |
| 2006 | 21 | 4,541 | 10 | 154 |
| 2007 | 18 | 3,688 | 11 | 212 |
| 2008 | 13 | 3,242 | 10 | 221 |
| 2009 | 6 | 2,649 | 6 | 238 |
| 2010 | 5 | 1,708 | 5 | 134 |
| 2011 | 3 | 1,210 | 3 | 143 |
| 2012 | 20 | 1,630 | 12 | 192 |
| 2013 | 17 | 2,277 | 13 | 223 |
| 2014 | 17 | 2,034 | 12 | 152 |
| 2015 | 15 | 1,045 | 4 | 19 |
| 2016 | 16 | 2,407 | 9 | 175 |
| 2017 | 32 | 2,831 | 14 | 134 |
| 2018 | 39 | 4,581 | 16 | 98 |
| 2019 | 36 | 3,963 | 11 | 109 |
| 2020 | 33 | 3,321 | 9 | 288 |
| 2021 | 41 | 6,192 | 18 | 466 |

| Year | # Vendors | # Invoices Collected | # BMUS Vendors | # BMUS Invoices Collected |
|------------|-----------|-------------------------|-------------------|---------------------------------|
| 2022 | 46 | 6,850 | 21 | 592 |
| 10-yr avg. | 29 | 3,550 | 13 | 226 |
| 10-yr SD | 11 | 1,766 | 5 | 169 |
| 20-yr avg. | 23 | 3,343 | 11 | 197 |
| 20-yr SD | 12 | 1,528 | 4 | 127 |

^{&#}x27;n.d.' indicates that data are non-disclosed due to confidentiality rules (i.e., less than three dealers and/or vendors).

1.1.4 Fishery Summary Dashboard Statistics

The Fishery Summary Dashboard Statics section consolidates all fishery-dependent information comparing the most recent year with short-term (recent 10 years) and long-term (recent 20 years) average (shown bolded in [brackets]). Trend analysis of the past 10 years will dictate the trends (increasing, decreasing, or no trend). The right-most symbol indicates whether the mean of the short-term and long-term years were above, below, or within one standard deviation of the mean of the full time series.

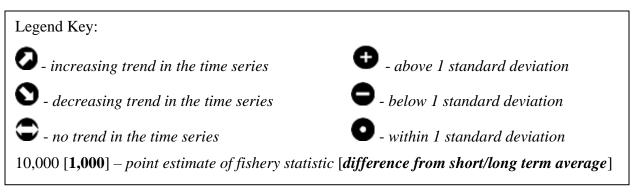


Table 3. Annual indicators for CNMI bottomfish fisheries describing performance and comparing estimates from 2022 with short- (10-year) and long-term (20-year) averages

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) | |
|--------------------------------------|---|---------------------------------|----------------------------------|--|
| Bottomfish | Total estimated catch (lb) | | | |
| All gears (BMUS only) | All BMUS from creel survey data | 47,564[▲44%]�� | 47,564[▲18%] | |
| | All BMUS from commercial purchase data | 32,161[▲ 78%] ⊘⊕ | 32,161[▲ 99%] ⊘ ⊕ | |
| | Catch-per-unit-effort (from boat-based creel surveys) | | | |
| Bottomfish fishing (BMUS only) | Bottomfish fishing lb/trip | 38[▼14%]�� | 38[▼3%] 🛇 🔾 | |
| | Bottomfish fishing lb/gr-h. | 3.84[▼4%]�� | 3.84[128%] | |
| | Fishing effort (from boat-based creel surveys) | | | |
| Bottomfish | Tallied bottomfish trips | 37[▲85%]�� | 37[▲16%] ۞ | |

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) |
|------------------------|--|--------------------------|----------------------|
| fishing (BMUS only) | Tallied bottomfish gear hours | 368[▲44%] ❷ ● | 368[▼27%] �� |
| | Fishing participants (from b | oat-based creel surveys) | |
| Bottomfish | Tallied number of bottomfish fishing vessels | 20[▲25%] | 20[▼5%]�� |
| fishing (BMUS only) | Estimated average number of fishermen per bottomfish fishing trip | 2[no change] | 2[▼60%]�� |
| | Bycatch | <u> </u> | |
| | # fish caught | 561[▲57%]�� | 561[▼4%]�� |
| BMUS | # fish discarded/released | 0[no change] | 0[▼100%]�� |
| | % bycatch | 0[no change] | 0[▼100%]�� |

Table 4. Annual indicators for CNMI ECS fisheries describing performance and comparing 2022 estimates with short- (10-year) and long-term (20-year) averages

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) | | | | |
|-----------------|--|-----------------------------------|----------------------------------|--|--|--|--|
| ECS | Estimated catch (lb) | | | | | | |
| | Acanthurus lineatus from creel survey data | 808[▲83%] | 808[▲155%] | | | | |
| | Acanthurus lineatus from commercial data | 498[▲840%] ② ⊕ | 498[▲703%] ⊘ ⊕ | | | | |
| | Naso lituratus from creel survey data | 4,787[▲ 259%] ⊘ ↔ | 4,787[▲336%] | | | | |
| | Naso lituratus from commercial data | _{7,594[▲212%]} ⊘⊕ | 7,594[▲406%] ② ⊕ | | | | |
| | Naso unicornis from creel survey data | 4,192[▲460%] ⊘ ↔ | 4,192[▲ 495%] ⊘ ⊕ | | | | |
| Prioritized ECS | Naso unicornis from commercial data | 5,340[▲543%] ② ⊕ | 5,340[▲ 911%] ⊘ ⊕ | | | | |
| Filontized ECS | Scarus ghobban from creel survey data | 13[▲44%] ❷ ● | 13[▲160%] ⊘ • | | | | |
| | Lethrinus harak from creel survey data | 0[▼100%]�� | _{0[▼100%]} �� | | | | |
| | Lethrinus harak from commercial data | 0[no change] | 0[no change] | | | | |
| | Siganus argenteus from creel survey data | 276[▼ 1%] ⊘⊙ | ²⁷⁶ [▼31%] ○ ○ | | | | |
| | Siganus argenteus from commercial data | 1,801[▲137%] 🗸 🖨 | 1,801[▲91%] | | | | |
| | Mulloidichthys flavolineatus from creel | 548[▲162%] ② ◆ | 548[▲129%] ⑤ ⑥ | | | | |

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) |
|---------|---|-----------------------|----------------------|
| | survey data | | |
| | Mulloidichthys flavolineatus from commercial data | 0[▼100%] 🗢 💽 | 0[▼100%] ⊘⊙ |

1.1.5 Catch Statistics

The following section summarizes the catch statistics for bottomfish, the top ten landed ECS, and seven prioritized ECS in CNMI as decided by DFW. Estimates of catch are summarized from the creel survey and commercial receipt book data collection programs. Catch statistics provide estimates of annual harvest from the different fisheries. Estimates of fishery removals can provide proxies for the level of fishing mortality and a reference level relative to established quotas.

1.1.5.1 Catch by Data Stream

This section describes the estimated total catch from the boat-based creel survey programs as well as the commercial landings from the commercial receipt book system. The difference between the creel total and the commercial landings is assumed to be the non-commercial component. However, there are cases where the commercial landing may be higher than the estimated creel total of the commercial receipt book program. In this case, the commercial receipt books can capture the fishery better than the creel surveys. While the reporting of commercial landings for Guam boat-based archipelagic fisheries is often constrained by rules associated with data confidentiality (i.e., commercial data must be sourced by at least three commercial purchasers to be reported), the relative lack of purchaser reports from Guam is likely related to non-participation by commercial purchasers rather than being reflective of a paucity of commercial purchasers.

Calculations: Estimated landings are based on a pre-determined list of species (Appendix A) identified as BMUS regardless of the gear used, for each data collection (creel surveys and the commercial purchase reports).

Table 5. Summary of CNMI BMUS total catch (lb) from expanded boat-based and shore-based creel surveys and the commercial purchase system for all gear types

| Year | Boat-Based Creel Survey Estimates | Shore-Based Creel Survey Estimates | Total Creel Survey Estimates | Commercial Landings |
|------|---|--|------------------------------------|------------------------|
| 1983 | - | - | - | 3,407 |
| 1984 | - | - | - | 3,463 |
| 1985 | - | - | - | n.d. |
| 1986 | - | - | - | n.d. |
| 1987 | - | - | - | 1,889 |
| 1988 | - | - | - | 2,413 |
| 1989 | - | - | - | 4,021 |
| 1990 | - | - | - | 1,273 |
| 1991 | - | - | - | 781 |
| 1992 | - | - | - | 158 |

| Year | Boat-Based Creel Survey Estimates | Shore-Based Creel Survey Estimates | Total Creel Survey Estimates | Commercial Landings |
|------------|---|--|------------------------------------|------------------------|
| 1993 | - | - | - | 1,722 |
| 1994 | - | - | - | 5,459 |
| 1995 | - | - | - | 17,564 |
| 1996 | - | - | - | 32,294 |
| 1997 | - | - | - | 21,607 |
| 1998 | - | - | - | 25,529 |
| 1999 | - | - | - | 33,622 |
| 2000 | 67,252 | - | 67,252 | 14,751 |
| 2001 | 24,637 | - | 24,637 | 24,817 |
| 2002 | 24,603 | - | 24,603 | 24,296 |
| 2003 | 12,726 | - | 12,726 | 17,144 |
| 2004 | 30,407 | - | 30,407 | 11,292 |
| 2005 | 34,311 | 168 | 34,479 | 15,025 |
| 2006 | 35,279 | 5 | 35,284 | 11,837 |
| 2007 | 54,257 | 648 | 54,905 | 14,805 |
| 2008 | 21,118 | 69 | 21,187 | 15,098 |
| 2009 | 65,269 | 21 | 65,290 | 18,313 |
| 2010 | 56,007 | 2 | 56,009 | 12,971 |
| 2011 | 25,799 | 22 | 25,821 | 16,115 |
| 2012 | 137,495 | 84 | 137,579 | 10,591 |
| 2013 | 20,390 | 0 | 20,390 | 16,500 |
| 2014 | 7,740 | 166 | 7,906 | 16,334 |
| 2015 | 10,386 | 215 | 10,601 | 4,121 |
| 2016 | 54,335 | 36 | 54,371 | 17,717 |
| 2017 | 48,007 | 59 | 48,066 | 11,923 |
| 2018 | 650 | 2 | 652 | 7,258 |
| 2019 | 21,012 | 2 | 21,014 | 15,697 |
| 2020 | 45,547 | 36 | 45,583 | 20,071 |
| 2021 | 73,861 | 739 | 74,600 | 38,946 |
| 2022 | 47,417 | 147 | 47,564 | 32,161 |
| 10-yr avg. | 32,935 | 156 | 33,075 | 18,073 |
| 10-yr SD | 22,834 | 218 | 22,946 | 9,994 |
| 20-yr avg. | 40,101 | 142 | 40,222 | 16,196 |
| 20-yr SD | 29,804 | 212 | 29,856 | 7,513 |

^{&#}x27;-' = No data available; 'n.d.' indicates that data are non-disclosed due to confidentiality rules (i.e., less than three dealers and/or vendors).

1.1.5.2 Expanded Catch Estimates by Fishing Method

Catch information is provided for the top boat-based fishing methods that comprise most of the annual BMUS catch in CNMI.

Calculations: The creel survey catch time series are the sum of the estimated weight for selected gear in all strata for all species all BMUS species.

Table 6. Total catch time series estimates (lb) for all species and BMUS only using CNMI expanded boat-based creel survey data for bottomfish fishing gears

| Voor | Botto | mfish | Spearfishing (Snorkel) | | |
|------------|---------|---------|------------------------|-------|--|
| Year | All | BMUS | All | BMUS | |
| 2000 | 99,106 | 62,990 | 27,918 | 4,262 | |
| 2001 | 40,556 | 24,574 | 8,693 | 63 | |
| 2002 | 37,621 | 23,945 | 9,990 | 159 | |
| 2003 | 15,406 | 12,547 | 5,528 | 178 | |
| 2004 | 40,060 | 30,407 | 7,452 | 0 | |
| 2005 | 48,699 | 34,266 | 6,567 | 46 | |
| 2006 | 61,157 | 34,951 | 8,553 | 15 | |
| 2007 | 83,677 | 54,059 | 11,849 | 198 | |
| 2008 | 51,075 | 19,744 | 15,516 | 1,334 | |
| 2009 | 99,523 | 64,979 | 18,801 | 217 | |
| 2010 | 82,211 | 56,007 | 5,814 | 0 | |
| 2011 | 60,432 | 25,799 | 7,289 | 0 | |
| 2012 | 157,445 | 137,495 | 8,513 | 0 | |
| 2013 | 34,954 | 20,390 | 2,456 | 0 | |
| 2014 | 15,291 | 7,740 | 2,257 | 0 | |
| 2015 | 17,554 | 10,374 | 4,820 | 0 | |
| 2016 | 56,983 | 53,906 | 0 | 0 | |
| 2017 | 50,177 | 47,883 | 0 | 0 | |
| 2018 | 4,347 | 90 | 4,087 | 0 | |
| 2019 | 25,556 | 16,831 | 10,486 | 0 | |
| 2020 | 73,773 | 45,358 | 6,892 | 189 | |
| 2021 | 89,963 | 70,013 | 31,608 | 32 | |
| 2022 | 68,470 | 46,808 | 32,224 | 374 | |
| 10-yr avg. | 43,707 | 31,939 | 9,483 | 60 | |
| 10-yr SD | 27,059 | 22,392 | 11,605 | 119 | |
| 20-yr avg. | 56,838 | 39,482 | 9,536 | 129 | |
| 20-yr SD | 34,988 | 29,797 | 8,729 | 295 | |

1.1.5.3 Top and Prioritized ECS in Boat-Based Fishery Catch

Catch can act as an indicator of fishery performance. Variations in the catch can be attributed to several factors, and there is no single explanatory variable for the observed trends. A one-year

reflection of the top ten harvested species (by weight) is included to monitor which ECS are being caught the most annually. Additionally, CNMI DFW selected seven species that were reclassified as ECS that are still of priority to CNMI DFW for regular monitoring, and complete catch time series of these species are included in the report as well.

Calculations: Catch tallied from the boat-based expanded species composition data combining gear types for all species excluding BMUS and pelagic MUS species.

Table 7a. Top ten landed species (lb) in CNMI ECS fisheries from expanded boat-based creel survey data in 2022

| Common Name | Scientific Name | Catch |
|-------------------------|--------------------------|-------|
| Bigeye scad | Selar crumenophthalmus | 5,743 |
| Tan-faced parrotfish | Chlorurus frontalis | 4,719 |
| Orangespine unicornfish | Naso lituratus | 3,810 |
| Bluespine unicornfish | Naso unicornis | 3,310 |
| Yellowtail kalikali | Pristipomoides auricilla | 3,300 |
| Yellowband parrotfish | Scarus schlegeli | 3,207 |
| Bluefin trevally | Caranx melampygus | 2,395 |
| Longnose emperor | Lethrinus olivaceus | 2,157 |
| Blackspot emperor | Lethrinus harak | 1,635 |
| Stareye parrotfish | Calotomus carolinus | 1,091 |

Calculations: Catch tallied from commercial receipt data combining gear types for all species excluding BMUS and pelagic MUS species.

Table 7b. Top ten landed species (lb) in CNMI ECS fisheries from commercial landings data in 2022

| Common Name | Scientific Name | Catch |
|-------------------------------|---------------------------|--------|
| Bigeye scad | Selar crumenophthalmus | 20,295 |
| Parrotfish (palakse/la/misc.) | Scaridae (family) | 15,342 |
| Surgeonfish (misc.) | Acanthuridae (family) | 7,429 |
| Goatfish (satmoneti) | Mullidae (family) | 6,325 |
| Emperor (mafute/misc.) | Lethrinidae (family) | 4,824 |
| Rudderfish (guili) | Kyphosus spp. | 4,123 |
| Squirrelfish (sagamelon) | Holocentrinae (subfamily) | 2,399 |
| Unicornfish (tataga) | Naso spp. | 2,368 |
| Spiny lobster | Panulirus spp. | 2,183 |
| Jacks (misc.) | Carangidae (family) | 2,013 |

Calculations: Catch tallied from boat-based expanded species composition data for species identified as priority ECS by DFW (Appendix A).

Table 8a. Catch (lb) from expanded boat-based creel survey data for prioritized species in CNMI ECS fisheries

| Year | Acanthurus | Naso | Naso | Scarus | Lethrinus | Siganus | Mulloidichthys |
|------|------------|-----------|-----------|---------|-----------|-----------|----------------|
| | lineatus | lituratus | unicornis | ghobban | harak | argenteus | flavolineatus |
| 2000 | 0 | 1,189 | 43 | 0 | 0 | 0 | 955 |

| Year | Acanthurus lineatus | Naso lituratus | Naso unicornis | Scarus ghobban | Lethrinus harak | Siganus argenteus | Mulloidichthys flavolineatus |
|--------------|------------------------|-------------------|-------------------|-------------------|--------------------|----------------------|---------------------------------|
| 2001 | 0 | 849 | 222 | 0 | 0 | 0 | 136 |
| 2002 | 0 | 2,238 | 981 | 0 | 0 | 0 | 1,034 |
| 2003 | 345 | 1,125 | 965 | 0 | 0 | 136 | 227 |
| 2004 | 601 | 458 | 323 | 0 | 0 | 0 | 11 |
| 2005 | 339 | 451 | 250 | 0 | 0 | 272 | 0 |
| 2006 | 249 | 375 | 1,662 | 0 | 0 | 2,676 | 28 |
| 2007 | 200 | 1,139 | 1,125 | 0 | 0 | 4,640 | 114 |
| 2008 | 0 | 636 | 135 | 0 | 0 | 7,318 | 317 |
| 2009 | 0 | 3,555 | 524 | 0 | 0 | 8,996 | 1,385 |
| 2010 | 0 | 600 | 0 | 0 | 0 | 1,063 | 615 |
| 2011 | 40 | 81 | 1,611 | 0 | 0 | 1,648 | 0 |
| 2012 | 155 | 190 | 0 | 0 | 0 | 6,941 | 0 |
| 2013 | 0 | 77 | 0 | 0 | 0 | 1,224 | 0 |
| 2014 | 34 | 223 | 0 | 0 | 0 | 1,819 | 736 |
| 2015 | 87 | 383 | 64 | 0 | 48 | 386 | 29 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 408 | 0 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 45 | 0 |
| 2018 | 0 | 412 | 0 | 0 | 0 | 1,896 | 489 |
| 2019 | 0 | 346 | 0 | 0 | 0 | 1,979 | 0 |
| 2020 | 113 | 1,382 | 2,186 | 0 | 0 | 5,387 | 0 |
| 2021 | 3,363 | 5,735 | 1,051 | 80 | 16 | 3,297 | 284 |
| 2022 | 808 | 4,787 | 4,192 | 13 | 0 | 1,801 | 548 |
| 10-year avg. | 441 | 1,335 | 749 | 9 | 6 | 1,824 | 209 |
| 10-year SD | 1,056 | 2,119 | 1,408 | 25 | 15 | 1,581 | 284 |
| 20-year avg. | 317 | 1,098 | 704 | 5 | 3 | 2,597 | 239 |
| 20-year SD | 750 | 1,633 | 1,065 | 18 | 11 | 2,677 | 361 |

Calculations: Catch tallied from commercial purchase data for species identified as priority ECS by DFW (Appendix A). From the prioritized ECS list, *Scarus ghobban* is not included because there is no specific code for that species in the CNMI commercial coding system.

Table 8b. Catch (lb) from commercial purchase data for prioritized species in CNMI ECS fisheries

| Year | Acanthurus lineatus | Naso lituratus | Naso unicornis | Lethrinus harak | Siganus argenteus | Mulloidichthys flavolineatus |
|------|---------------------|-------------------|-------------------|--------------------|----------------------|------------------------------|
| 1983 | 0 | 0 | 0 | 0 | 7,644 | 0 |
| 1984 | 0 | 0 | 0 | 0 | 9,792 | 0 |
| 1985 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. |
| 1986 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. |
| 1987 | 0 | 0 | 0 | 0 | 4,061 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 6,653 | 0 |

| Year | Acanthurus lineatus | Naso lituratus | Naso unicornis | Lethrinus harak | Siganus argenteus | Mulloidichthys flavolineatus |
|------------|------------------------|-------------------|-------------------|--------------------|----------------------|---------------------------------|
| 1989 | 0 | 0 | 0 | 0 | 8,434 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 5,678 | 0 |
| 1991 | 0 | 0 | 0 | 0 | 3,858 | 0 |
| 1992 | 0 | 0 | 0 | 0 | 3,151 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 1,603 | 0 |
| 1994 | 0 | 0 | 0 | 0 | 2,181 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 904 | 0 |
| 1996 | 0 | 1,434 | 0 | 0 | 1,338 | 0 |
| 1997 | 0 | 3,173 | 0 | 0 | 1,093 | 0 |
| 1998 | 0 | 106 | 0 | 0 | 5,956 | 0 |
| 1999 | 0 | 1,756 | 0 | 0 | 6,442 | 0 |
| 2000 | 0 | 4,883 | 0 | 0 | 12,677 | 0 |
| 2001 | 0 | 4,500 | 0 | 0 | 8,408 | 0 |
| 2002 | 0 | 1,041 | 0 | 0 | 9,141 | 0 |
| 2003 | 0 | 143 | 0 | 0 | 7,161 | 0 |
| 2004 | 0 | 2 | 0 | 0 | 3,714 | 0 |
| 2005 | 0 | 64 | 0 | 0 | 2,571 | 0 |
| 2006 | 0 | 70 | 0 | 0 | 8,354 | 0 |
| 2007 | 0 | 426 | 0 | 0 | 5,909 | 0 |
| 2008 | 0 | 323 | 0 | 0 | 2,599 | 0 |
| 2009 | 0 | 313 | 0 | 0 | 1,312 | 0 |
| 2010 | 717 | 1,123 | 462 | 0 | 1,880 | 0 |
| 2011 | 0 | 2,804 | 1,804 | 0 | 2,185 | 0 |
| 2012 | 0 | 451 | 0 | 0 | 1,467 | 0 |
| 2013 | 0 | 759 | 0 | 0 | 2,331 | 0 |
| 2014 | 0 | 1,827 | 0 | 0 | 2,329 | 0 |
| 2015 | 0 | 1,380 | 0 | 0 | 1,569 | 0 |
| 2016 | 0 | 1,018 | 0 | 0 | 2,319 | 0 |
| 2017 | 0 | 1,664 | 0 | 0 | 3,063 | 18 |
| 2018 | 0 | 415 | 0 | 0 | 1,008 | 0 |
| 2019 | 0 | 320 | 0 | 0 | 293 | 0 |
| 2020 | 0 | 2,887 | 0 | 0 | 390 | 0 |
| 2021 | 33 | 6,465 | 2,962 | 0 | 5,820 | 0 |
| 2022 | 498 | 7,594 | 5,340 | 0 | 5,949 | 0 |
| 10-yr avg. | 53 | 2,433 | 830 | 0 | 2,507 | 2 |
| 10-yr SD | 157 | 2,550 | 1,838 | 0 | 1,993 | 6 |
| 20-yr avg. | 62 | 1,502 | 528 | 0 | 3,111 | 1 |
| 20-yr SD | 190 | 2,081 | 1,361 | 0 | 2,298 | 4 |

'n.d.' indicates that data are non-disclosed due to confidentiality rules (i.e., less than three dealers and/or vendors).

1.1.6 Catch-per-Unit-Effort (CPUE) Statistics

This section summarizes the estimates for CPUE in the boat-based BMUS fisheries. The boat-based fisheries include the bottomfish fishing (handline gear) and spearfishing (snorkel). CPUE is reported as pounds per gear hour in the boat-based fishery.

Calculations: CPUE is calculated from interview data by gear type using \sum catch $/\sum$ (number of gears used*number of hours fished) or \sum catch $/\sum$ trips for boat-based data. If the value is blank (i.e., zero), then there was no interview collected for that method. Landings from interviews without fishing hours or number of gears are excluded from the calculations.

All - lb/trip: All catch and trips are tallied from landings by gear level, including non-BMUS species.

All - lb/gr-hr.: All catch and trips are tallied from trips with data on the number of gears used and numbers of hours fished, including non-BMUS species.

BMUS - lb/trip: Only BMUS catch and trips that landed BMUS species are tallied from landings by gear level.

BMUS - lb/gr-hr.: Only BMUS catch and trips that landed BMUS are tallied from trips with data on the number of gears used and numbers of hours fished.

Table 9. CPUE (lb/trip and lb/gear hour) for bottomfish fishing gears in the CNMI boatbased fishery for all species and BMUS only

| | | Botto | mfish | | Spearfishing (Snorkel) | | | | |
|------|---------|------------|---------|------------|------------------------|----------|---------|------------|--|
| Year | A | All | BN | IUS | | All | T | IUS | |
| | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | |
| 2000 | 50 | 4.44 | 55 | 4.76 | 35 | 2.43 | 64 | 5.33 | |
| 2001 | 17 | 1.64 | 21 | 1.89 | 19 | 1.48 | 2 | 0.11 | |
| 2002 | 28 | 2.22 | 32 | 2.35 | 20 | 1.55 | 3 | 0.38 | |
| 2003 | 21 | 1.76 | 21 | 1.64 | 29 | 2.07 | 4 | 0.29 | |
| 2004 | 25 | 2.03 | 20 | 1.55 | 15 | 0.91 | 0 | 0.00 | |
| 2005 | 26 | 2.01 | 26 | 1.72 | 21 | 1.82 | 1 | 0.15 | |
| 2006 | 18 | 1.43 | 17 | 1.22 | 12 | 1.25 | 1 | 0.10 | |
| 2007 | 28 | 2.65 | 28 | 2.42 | 15 | 1.05 | 2 | 0.12 | |
| 2008 | 16 | 1.03 | 13 | 0.88 | 21 | 1.19 | 6 | 0.23 | |
| 2009 | 19 | 0.77 | 34 | 1.47 | 21 | 1.39 | 3 | 0.08 | |
| 2010 | 12 | 0.40 | 11 | 0.39 | 15 | 1.32 | 0 | 0.00 | |
| 2011 | 11 | 0.34 | 16 | 0.54 | 38 | 2.76 | 0 | 0.00 | |
| 2012 | 108 | 8.83 | 156 | 9.85 | 13 | 1.03 | 0 | 0.00 | |
| 2013 | 46 | 4.30 | 44 | 3.59 | 20 | 1.33 | 0 | 0.00 | |
| 2014 | 18 | 1.87 | 32 | 3.63 | 33 | 1.89 | 0 | 0.00 | |
| 2015 | 34 | 2.77 | 43 | 3.00 | 19 | 3.26 | 0 | 0.00 | |
| 2016 | 69 | 5.28 | 78 | 5.68 | 0 | 0.00 | 0 | 0.00 | |
| 2017 | 81 | 8.16 | 115 | 12.97 | 0 | 0.00 | 0 | 0.00 | |
| 2018 | 5 | 0.41 | 1 | 0.14 | 9 | 0.88 | 0 | 0.00 | |

| | | Botto | mfish | | Spearfishing (Snorkel) | | | | |
|------------|---------|------------|---------|-------------|------------------------|----------|---------|------------|--|
| Year | A | All | BN | BMUS | | All | | IUS | |
| | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | |
| 2019 | 26 | 2.19 | 23 | 2.42 | 10 | 0.83 | 0 | 0.00 | |
| 2020 | 28 | 2.03 | 29 | 1.89 | 14 | 0.84 | 2 | 0.09 | |
| 2021 | 39 | 2.73 | 41 | 2.51 | 41 | 2.11 | 2 | 0.05 | |
| 2022 | 40 | 4.19 | 38 | 3.84 | 42 | 3.22 | 6 | 0.58 | |
| 10-yr avg. | 39 | 3 | 44 | 4 | 24 | 2 | 3 | 0 | |
| 10-yr SD | 22 | 2 | 30 | 3 | 13 | 1 | 2 | 0 | |
| 20-yr avg. | 34 | 3 | 39 | 3 | 22 | 2.00 | 3 | 0 | |
| 20-yr SD | 25 | 2 | 36 | 3 | 10 | 1.00 | 2 | 0 | |

1.1.7 Effort Statistics

This section summarizes the effort trends in the CNMI bottomfish fishery. Fishing effort trends provide insights on the level of fishing pressure through time. Effort information is provided for the top boat-based fishing methods that comprise most of the annual catch.

Calculations: Effort estimates (in both trips and gear hours) are calculated from boat-based interview data. Trips are tallied according to the interview data in boat-based creel surveys. Gear hours are generated by summing the data on number of gears used*number of hours fished collected from interviews by gear type. For the boat-based estimates, data collection started in 2000.

All - Trips: All trips tallied by gear type.

All - Gear-hr: Gear hours tallied by gear type.

BMUS - Trips: Trips that landed BMUS tallied by gear type.

BMUS - Gr-hr: Gear hours tallied by gear type for trips landed BMUS with data on both number of gears used and numbers of hours fished.

Table 10. Effort (trips and gear hours) for bottomfish fishing gears in the CNMI boatbased fishery for all species and BMUS only

| | | Bott | tomfish | | Spear Snorkel | | | | |
|------|-------|-------|---------|-------|---------------|-------|-------|-------|--|
| Year | All | | BMUS | | All | | BMUS | | |
| | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | |
| 2000 | 35 | 392 | 24 | 276 | 13 | 186 | 1 | 12 | |
| 2001 | 50 | 529 | 20 | 221 | 14 | 181 | 1 | 18 | |
| 2002 | 40 | 505 | 22 | 299 | 12 | 156 | 1 | 8 | |
| 2003 | 34 | 403 | 25 | 323 | 8 | 112 | 2 | 28 | |
| 2004 | 53 | 656 | 45 | 579 | 17 | 274 | 0 | 0 | |
| 2005 | 124 | 1,600 | 85 | 1,285 | 25 | 286 | 3 | 27 | |
| 2006 | 101 | 1,248 | 59 | 810 | 27 | 253 | 1 | 10 | |
| 2007 | 81 | 852 | 48 | 552 | 32 | 464 | 4 | 66 | |
| 2008 | 57 | 881 | 23 | 351 | 9 | 159 | 3 | 78 | |

| | | Bott | tomfish | | Spear Snorkel | | | | |
|------------|-------|-------|---------|-------|---------------|-------|-------|-------|--|
| Year | | All | BM | IUS | 1 | All | BN | MUS | |
| | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | |
| 2009 | 100 | 1,901 | 34 | 488 | 19 | 280 | 2 | 24 | |
| 2010 | 116 | 3,510 | 63 | 1,743 | 5 | 56 | 0 | 0 | |
| 2011 | 134 | 4,439 | 37 | 1,097 | 4 | 55 | 0 | 0 | |
| 2012 | 26 | 318 | 16 | 253 | 10 | 124 | 0 | 0 | |
| 2013 | 29 | 309 | 16 | 197 | 5 | 74 | 0 | 0 | |
| 2014 | 17 | 160 | 6 | 52 | 3 | 53 | 0 | 0 | |
| 2015 | 14 | 170 | 7 | 100 | 4 | 23 | 0 | 0 | |
| 2016 | 20 | 263 | 16 | 219 | 0 | 0 | 0 | 0 | |
| 2017 | 13 | 127 | 7 | 61 | 0 | 0 | 0 | 0 | |
| 2018 | 12 | 140 | 2 | 14 | 4 | 41 | 0 | 0 | |
| 2019 | 13 | 156 | 9 | 85 | 2 | 23 | 0 | 0 | |
| 2020 | 51 | 710 | 30 | 463 | 8 | 130 | 2 | 35 | |
| 2021 | 101 | 1,349 | 67 | 995 | 21 | 364 | 2 | 76 | |
| 2022 | 51 | 486 | 37 | 368 | 16 | 207 | 2 | 19 | |
| 10-yr avg. | 32 | 387 | 20 | 255 | 6 | 91 | 1 | 13 | |
| 10-yr SD | 27 | 366 | 19 | 282 | 7 | 109 | 1 | 24 | |
| 20-yr avg. | 57 | 984 | 32 | 502 | 11 | 149 | 1 | 18 | |
| 20-yr SD | 41 | 1,128 | 23 | 455 | 9 | 129 | 1 | 26 | |

1.1.8 Participants

This section summarizes the estimated participation in the bottomfish fishery. The information presented here can be used in the impact analysis of potential amendments in the FEPs associated with the bottomfish fisheries. The trend in participation over time can also be used as an indicator for fishing pressure.

Calculations: For boat-based data, the estimated number of unique vessels is calculated by tallying the number of vessels recorded in the interview data via vessel registration or name.

All: Total unique vessels by gear type.

BMUS: Unique vessels from trips that landed BMUS by gear type.

Table 11a. Estimated number of unique vessels for bottomfish fishing gears in the CNMI boat-based fishery for all species and BMUS only

| Year | Botto | mfish | Spearfishing (Snorkel) | | |
|-------|-------|-------|------------------------|------|--|
| i ear | All | BMUS | All | BMUS | |
| 2000 | 24 | 18 | 12 | 1 | |
| 2001 | 35 | 15 | 10 | 1 | |
| 2002 | 25 | 15 | 11 | 1 | |
| 2003 | 22 | 15 | 6 | 2 | |
| 2004 | 29 | 24 | 13 | 0 | |

| V | Bott | omfish | Spearfishii | ng (Snorkel) |
|------------|------|--------|-------------|--------------|
| Year | All | BMUS | All | BMUS |
| 2005 | 67 | 51 | 22 | 3 |
| 2006 | 60 | 42 | 18 | 1 |
| 2007 | 58 | 36 | 26 | 4 |
| 2008 | 40 | 22 | 9 | 3 |
| 2009 | 55 | 27 | 16 | 2 |
| 2010 | 26 | 19 | 5 | 0 |
| 2011 | 31 | 15 | 4 | 0 |
| 2012 | 23 | 15 | 9 | 0 |
| 2013 | 25 | 15 | 4 | 0 |
| 2014 | 14 | 5 | 3 | 0 |
| 2015 | 12 | 6 | 4 | 0 |
| 2016 | 16 | 13 | 0 | 0 |
| 2017 | 12 | 6 | 0 | 0 |
| 2018 | 11 | 2 | 3 | 0 |
| 2019 | 12 | 8 | 2 | 0 |
| 2020 | 44 | 27 | 8 | 2 |
| 2021 | 87 | 58 | 17 | 2 |
| 2022 | 31 | 20 | 13 | 2 |
| 10-yr avg. | 26 | 16 | 5 | 1 |
| 10-yr SD | 23 | 16 | 5 | 1 |
| 20-yr avg. | 34 | 21 | 9 | 1 |
| 20-yr SD | 21 | 15 | 7 | 1 |

Calculations: For boat-based data, the estimated number of fishermen per trip is calculated by filtering interviews that recorded the number of fishers, and then \sum fishers/ \sum trips.

All: Average fishers from all trips by gear type.

BMUS: Average fishers from trips that landed BMUS by gear type.

Table 11b. Estimated number of fishermen per trip for bottomfish fishing gears in the CNMI boat-based fishery for all species and BMUS only

| Voor | Botto | mfish | Spearfishing (Snorkel) | | |
|------|-------|-------|------------------------|------|--|
| Year | All | BMUS | All | BMUS | |
| 2000 | 4 | 3 | 4 | 8 | |
| 2001 | 3 | 3 | 3 | 2 | |
| 2002 | 4 | 4 | 3 | 2 | |
| 2003 | 5 | 5 | 3 | 2 | |
| 2004 | 4 | 5 | 4 | 0 | |
| 2005 | 5 | 5 | 3 | 2 | |
| 2006 | 4 | 4 | 3 | 3 | |
| 2007 | 3 | 3 | 3 | 3 | |

| Vaar | Bott | tomfish | Spearfishi | ng (Snorkel) |
|------------|------|---------|------------|--------------|
| Year | All | BMUS | All | BMUS |
| 2008 | 6 | 6 | 4 | 4 |
| 2009 | 10 | 6 | 4 | 3 |
| 2010 | 21 | 19 | 2 | 0 |
| 2011 | 21 | 17 | 3 | 0 |
| 2012 | 2 | 2 | 4 | 0 |
| 2013 | 2 | 2 | 2 | 0 |
| 2014 | 2 | 2 | 3 | 0 |
| 2015 | 2 | 2 | 2 | 0 |
| 2016 | 2 | 2 | 0 | 0 |
| 2017 | 2 | 2 | 0 | 0 |
| 2018 | 3 | 5 | 3 | 0 |
| 2019 | 2 | 2 | 3 | 0 |
| 2020 | 2 | 2 | 3 | 4 |
| 2021 | 2 | 2 | 4 | 6 |
| 2022 | 2 | 2 | 3 | 3 |
| 10-yr avg. | 2 | 2 | 2 | 1 |
| 10-yr SD | 0 | 1 | 1 | 2 |
| 20-yr avg. | 5 | 5 | 3 | 2 |
| 20-yr SD | 6 | 5 | 1 | 2 |

1.1.9 Bycatch Estimates

This section focuses on Magnuson-Stevens Fishery Conservation and Management Act (MSA) § 303(a)(11), which requires that all FMPs establish a standardized reporting methodology to assess the amount and type of bycatch occurring in the fishery, and include conservation and management measures that, to the extent practicable, minimize bycatch and bycatch mortality. The MSA § 303(a)(11) standardized reporting methodology is commonly referred to as a "Standardized Bycatch Reporting Methodology" (SBRM) and was added to the MSA by the Sustainable Fisheries Act of 1996 (SFA). The Council implemented omnibus amendments to FMPs in 2003 to address MSA bycatch provisions and established SBRMs at that time.

The following are recent bycatch estimates for the boat-based BMUS and non-BMUS fisheries. The bycatch estimates presented here are self-reported by fishers during creel survey interviews, and thus, the data are likely biased downward.

Calculations: The number caught is the sum of the total number of individuals found in the raw data including bycatch. The number discarded or released is number of individuals marked as bycatch. Percent bycatch is the sum of all released divided by the number caught.

Table 12. Time series of catch and bycatch in the CNMI boat-based BMUS and non-BMUS fisheries

| | | BMUS | | I | Non-BMUS | 8 | BMU | JS + Non-B | MUS |
|---------------|--------|-------------|---------|--------|----------|---------|--------|------------|---------|
| Year | | # | | | # | | | # | |
| 1 cai | # | Discard | % | # | Discard | % | # | Discard | % |
| | Caught | or | Bycatch | Caught | or | Bycatch | Caught | or | Bycatch |
| 2000 | 402 | Release | 2.42 | 205 | Release | 2.77 | 010 | Release | 2.57 |
| 2000 | 493 | 12 | 2.43 | 325 | 9 | 2.77 | 818 | 21 | 2.57 |
| 2001 | 268 | 0 | 0.00 | 663 | 1 | 0.15 | 931 | 1 | 0.11 |
| 2002 | 474 | 0 | 0.00 | 430 | 14 | 3.26 | 904 | 14 | 1.55 |
| 2003 | 627 | 3 | 0.48 | 250 | 33 | 13.20 | 877 | 36 | 4.10 |
| 2004 | 756 | 0 | 0.00 | 623 | 20 | 3.21 | 1,379 | 20 | 1.45 |
| 2005 | 2,206 | 4 | 0.18 | 1,019 | 0 | 0.00 | 3,225 | 4 | 0.12 |
| 2006 | 874 | 0 | 0.00 | 971 | 3 | 0.31 | 1,845 | 3 | 0.16 |
| 2007 | 1,325 | 0 | 0.00 | 785 | 0 | 0.00 | 2,110 | 0 | 0.00 |
| 2008 | 241 | 0 | 0.00 | 917 | 0 | 0.00 | 1,158 | 0 | 0.00 |
| 2009 | 596 | 0 | 0.00 | 1,183 | 0 | 0.00 | 1,779 | 0 | 0.00 |
| 2010 | 614 | 0 | 0.00 | 860 | 0 | 0.00 | 1,474 | 0 | 0.00 |
| 2011 | 482 | 0 | 0.00 | 1,252 | 0 | 0.00 | 1,734 | 0 | 0.00 |
| 2012 | 456 | 0 | 0.00 | 326 | 0 | 0.00 | 782 | 0 | 0.00 |
| 2013 | 519 | 0 | 0.00 | 338 | 0 | 0.00 | 857 | 0 | 0.00 |
| 2014 | 57 | 0 | 0.00 | 159 | 0 | 0.00 | 216 | 0 | 0.00 |
| 2015 | 102 | 0 | 0.00 | 94 | 0 | 0.00 | 196 | 0 | 0.00 |
| 2016 | 636 | 0 | 0.00 | 85 | 0 | 0.00 | 721 | 0 | 0.00 |
| 2017 | 120 | 0 | 0.00 | 194 | 0 | 0.00 | 314 | 0 | 0.00 |
| 2018 | 6 | 0 | 0.00 | 101 | 0 | 0.00 | 107 | 0 | 0.00 |
| 2019 | 139 | 0 | 0.00 | 105 | 0 | 0.00 | 244 | 0 | 0.00 |
| 2020 | 516 | 0 | 0.00 | 692 | 0 | 0.00 | 1,208 | 0 | 0.00 |
| 2021 | 913 | 0 | 0.00 | 568 | 2 | 0.35 | 1,481 | 2 | 0.14 |
| 2022 | 561 | 0 | 0.00 | 439 | 0 | 0.00 | 1,000 | 0 | 0.00 |
| 10-yr avg. | 357 | 0 | 0.00 | 278 | 0 | 0.04 | 634 | 0 | 0.01 |
| 10-yr SD | 294 | 0 | 0.00 | 209 | 1 | 0.11 | 462 | 1 | 0.04 |
| 20-yr avg. | 587 | 0 | 0.03 | 548 | 3 | 0.85 | 1,135 | 3 | 0.30 |
| 20-yr SD | 491 | 1 | 0.11 | 381 | 8 | 2.92 | 762 | 9 | 0.93 |

1.1.10 Federal Logbook Data

1.1.10.1 Number of Federal Permit Holders

In the CNMI, the following federal permits are required for fishing in the exclusive economic zone (EEZ) under the Mariana Archipelago FEP. Regulations governing fisheries under this FEP are in the Code of Federal Regulations (CFR), Title 50, Part 665.

1.1.10.1.1 Northern Mariana Island Bottomfish Permit

Regulations require this permit for any vessel commercially fishing for, landing, or transshipping BMUS or bottomfish ECS in the EEZ around CNMI. Commercial fishing is prohibited within the boundaries of the Islands Unit of the Marianas Trench Marine National Monument.

1.1.10.1.2 Special Coral Reef Ecosystem Permit

Regulations require the coral reef ecosystem special permit for anyone fishing for coral reef ECS in a low-use marine protected area (MPA), fishing for species on the list of Potentially Harvested Coral Reef Taxa or using fishing gear not specifically allowed in the regulations. NMFS will make an exception to this permit requirement for any person issued a permit to fish under any FEP who incidentally catches CNMI coral reef ECS while fishing for BMUS, crustacean ECS, western Pacific pelagic MUS, precious coral, or seamount groundfish. Regulations require a transshipment permit for any receiving vessel used to land or transship potentially harvested coral reef taxa, or any coral reef ECS caught in a low-use MPA.

1.1.10.1.3 Western Pacific Precious Corals Permit

Regulations require this permit for anyone harvesting or landing black, bamboo, pink, red, or gold corals in the EEZ in the western Pacific.

1.1.10.1.4 Western Pacific Crustaceans Permit (Lobster or Deepwater Shrimp)

Regulations require a permit by the owner of a U.S. fishing vessel used to fish for lobster or deepwater shrimp in the EEZ around American Samoa, Guam, Hawaii, and the Pacific Remote Islands Area (PRIA), and in the EEZ seaward of 3 nautical miles of the shoreline of the CNMI.

There is no record of special coral reef or precious coral fishery permits issued for the EEZ around the CNMI since 2007. Table 13 provides the number of permits issued for CNMI fisheries between 2013 and 2022. Data are from the NMFS Pacific Islands Regional Office (PIRO) Sustainable Fisheries Division (SFD) permits program.

Table 13. Number of federal permit holders for the CNMI crustacean and bottomfish fisheries

| CNMI Fisheries | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Bottomfish | 5 | 7 | 7 | 18 | 25 | 14 | 9 | 14 | 18 | 9 |
| Lobster | 0 | 0 | 0 | 1* | 0 | 1* | 0 | 0 | 0 | 0 |
| Shrimp | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |

Source: PIRO SFD unpublished data.

^{*} Permits apply to multiple areas and may include American Samoa, Guam, the CNMI, and PRIA.

1.1.10.2 Summary of Catch and Effort for FEP Fisheries

The Mariana Archipelago FEP requires fishermen to obtain a federal permit to fish for certain MUS and ECS in federal waters and to report all catch and discards. While NMFS annually issues permits for various FEP fisheries, there is currently limited data available on the level of catch or effort made by federal non-longline permit holders. Determining the level of fishing activity through the required federal logbook reporting for each fishery helps establish the level of non-longline fishing occurring in federal waters to assess whether there is a continued need for active conservation and management measures (e.g., annual catch limits) for these fisheries. For each FEP fishery, the number of federal permits issued since the federal permit and logbook reporting requirements became effective as well as available catch and effort data are presented in Table 14 through Table 16. NMFS has never issued a federal permit for precious coral or coral reef fishing in federal waters around CNMI. Therefore, catch and effort data are not presented for these fisheries.

1.1.10.2.1 Commercial Bottomfish Fishery

Table 14. Summary of available federal logbook data for the commercial bottomfish fishery in the CNMI

| Year | No. of Federal Bottomfish | No. of Federal Bottomfish Permits | No. of Trips in CNMI EEZ | | eported L Catch (lb) | O | Total Reported Logbook MUS Release/Discard (#s) | | |
|------|---------------------------------|---|-----------------------------------|---|-----------------------------------|----------------|---|--------------------------------|--|
| Tear | Permits Issued ¹ | Reporting Catch | | Bottomfish MUS & ECS ² | Coral Reef ECS ² | Pelagic MUS | Bottomfish MUS & ECS ² | Coral Reef ECS ² | |
| 2009 | 3 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | |
| 2010 | 12 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | |
| 2011 | 9 | 3 | 16 | 1,985 | 1,420 | 1,115 | | 13 | |
| 2012 | 14 | 5 | 40 | 2,309 | 1,765 | 159 | 52 | 10 | |
| 2013 | 5 | 4 | 9 | 3,103 | 632 | 300 | | | |
| 2014 | 7 | 0 | | | | | | | |
| 2015 | 7 | 0 | | | | | | | |
| 2016 | 18 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | |
| 2017 | 25 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | |
| 2018 | 14 | 0 | | | | | | | |
| 2019 | 9 | 0 | | | | | | | |
| 2020 | 14 | 0 | | | | | | | |
| 2021 | 18 | 0 | | | | | | | |
| 2022 | 9 | 0 | | | | | | | |

¹ Source: PIRO SFD unpublished data.

Notes: Federal permit and reporting requirements for CNMI bottomfish became effective on May 6, 2009 (74 FR 15373, April 6, 2009); n.d. = Not disclosed due to confidentiality.

1.1.10.2.2 Spiny and Slipper Lobster

Table 15. Summary of available federal logbook data for lobster fisheries in the CNMI

² On February 8, 2019, NMFS published a final rule (84 FR 2767) to reclassify some BMUS and all CREMUS in the Mariana Archipelago as ECS.

| Year | No. of Federal Lobster | No. of Federal Lobster Permits | No. of Trips in CNMI | Total Repor | | Total Reported Logbook Release/Discard (lb) | | |
|------|--------------------------------|-----------------------------------|----------------------------|-----------------------------------|-------------------------------------|--|-------------------------------------|--|
| | Permits Issued ¹ | Reporting Catch | EEZ | Spiny lobster ECS ² | Slipper lobster ECS ² | Spiny lobster ECS ² | Slipper lobster ECS ² | |
| 2006 | 2 | 0 | | | | | | |
| 2007 | 2 | 0 | | | | | | |
| 2008 | 7 | 0 | | | | | | |
| 2009 | 0 | - | | | | | | |
| 2010 | 0 | - | | | | | | |
| 2011 | 0 | - | | | | | | |
| 2012 | 0 | - | | | | | | |
| 2013 | 0 | - | | | | | | |
| 2014 | 0 | - | | | | | | |
| 2015 | 0 | - | | | | | | |
| 2016 | 1* | 0 | | | | | | |
| 2017 | 0 | - | | | | | | |
| 2018 | 1* | 0 | | | | | | |
| 2019 | 0 | - | | | | | | |
| 2020 | 0 | - | | | | | | |
| 2021 | 0 | - | | | | | | |
| 2022 | 0 | - | | | | | | |

¹ Source: PIRO Sustainable Fisheries unpublished data.

Note: Federal permit and reporting requirements for CNMI lobster fisheries became effective on December 4, 2006 (71 FR 69496, December 1, 2006).

1.1.10.2.3 Deepwater Shrimp

Table 16. Summary of available federal logbook data for deepwater shrimp fisheries in the $\overline{\text{CNMI}}$

| Year | No. of Federal Shrimp Permits Issued ¹ | No. of Federal Shrimp Permits Reporting Catch | No. of Trips in CNMI EEZ | Total Reported Logbook Shrimp ECS ² Catch (lb) | Total Reported Logbook Shrimp ECS ² Release/Discard (lb) |
|------|---|--|-----------------------------|---|---|
| 2009 | 0 | - | | | |
| 2010 | 2 | n.d. | n.d. | n.d. | n.d. |
| 2011 | 2 | 0 | | | |
| 2012 | 0 | - | _ | _ | _ |
| 2013 | 0 | - | _ | _ | _ |
| 2014 | 0 | - | _ | _ | |

² On February 8, 2019, NMFS published a final rule (84 FR 2767) to reclassify all CMUS in the Mariana Archipelago as ECS.

^{*} Permits apply to multiple areas and may include American Samoa, Guam, CNMI, and PRIA. Note: Federal permit and reporting requirements for CNMI lobster fisheries became effective on December 4, 2006

| Year | No. of Federal Shrimp Permits Issued ¹ | No. of Federal Shrimp Permits Reporting Catch | No. of Trips in CNMI EEZ | Total Reported Logbook Shrimp ECS ² Catch (lb) | Total Reported Logbook Shrimp ECS ² Release/Discard (lb) |
|------|---|--|-----------------------------|---|---|
| 2015 | 1 | 0 | | | |
| 2016 | 1 | 0 | | | |
| 2017 | 0 | - | | | |
| 2018 | 0 | - | | | |
| 2019 | 0 | - | | | |
| 2020 | 0 | - | | | |
| 2021 | 0 | - | | | |
| 2022 | 0 | - | _ | _ | |

¹ Source: PIRO SFD unpublished data.

Notes: Federal permit and reporting requirements for CNMI bottomfish became effective on June 29, 2009 (74 FR 25650, May 29, 2009); n.d. = Not disclosed due to confidentiality.

1.1.11 Status Determination Criteria

1.1.11.1 Bottomfish Fishery

Overfishing criteria and control rules are specified and applied to individual species within the multi-species stock whenever possible. When this is not possible, they are based on an indicator species for the multi-species stock. It is important to recognize that individual species would be affected differently based on the control rule, and it is important that for any given species, fishing mortality does not currently exceed a level that would result in excessive depletion of that species. No indicator species are used for the bottomfish multi-species stock complexes and the coral reef species complex. Instead, the control rules are applied to each entire stock complex.

The MSY control rule is used as the maximum fishing mortality threshold (MFMT). The MFMT and minimum stock size threshold (MSST) are specified based on the recommendations of Restrepo et al. (1998) and both are dependent on the natural mortality rate (M). The value of M used to determine the reference point values are not specified in this section. The latest estimate is used, and the value is occasionally re-estimated using the best available information. The range of M among species within a stock complex is taken into consideration when estimating and choosing the M to be used for the purpose of computing the reference point values.

In addition to the thresholds MFMT and MSST, a warning reference point, B_{FLAG} , is specified at some point above the MSST to provide a trigger for consideration of management action prior to B reaching the threshold. MFMT, MSST, and B_{FLAG} are specified as indicated in Table 17.

Table 17. Overfishing threshold specifications for the BMUS in the CNMI

| MFMT | MSST | $\mathbf{B}_{	ext{FLAG}}$ |
|--|-----------------------------|-------------------------------|
| $F(B) = \frac{F_{MSY}B}{c B_{MSY}} \text{for } B \le c B_{MSY}$ | $c~{ m B}_{	ext{	iny MSY}}$ | $\mathbf{B}_{	ext{	iny MSY}}$ |
| $F(B) = F_{MSY}$ for $B > c B_{MSY}$ | | |

² On February 8, 2019, NMFS published a final rule (84 FR 2767) to reclassify all CMUS in the Mariana Archipelago as ECS.

where $c = \max(1-M, 0.5)$

Standardized values of fishing effort (E) and CPUE can be used as proxies for F and B, respectively, so E_{MSY} , $CPUE_{MSY}$, and $CPUE_{FLAG}$ can be used as proxies for F_{MSY} , B_{MSY} , and B_{FLAG} , respectively.

In cases where reliable estimates of $CPUE_{MSY}$ and E_{MSY} are not available, they can be estimated from catch and effort times series, standardized for all identifiable biases. $CPUE_{MSY}$ would be calculated as half of a multi-year average reference CPUE, called $CPUE_{REF}$. The multi-year reference window would be objectively positioned in time to maximize the value of $CPUE_{REF}$. E_{MSY} would be calculated using the same approach or, following Restrepo et al. (1998), by setting E_{MSY} equal to E_{AVE} , where E_{AVE} represents the long-term average effort prior to declines in CPUE. When multiple estimates are available, the more precautionary one is used.

Since the MSY control rule specified here applies to multi-species stock complexes, it is important to ensure that no species within the complex has a mortality rate that leads to excessive depletion. In order to accomplish this, a secondary set of reference points is specified to evaluate stock status with respect to recruitment overfishing. A secondary "recruitment overfishing" control rule is specified to control fishing mortality with respect to that status. The rule applies only to those component stocks (species) for which adequate data are available. The ratio of a current spawning stock biomass proxy (SSBPt) to a given reference level (SSBPREF) is used to determine if individual stocks are experiencing recruitment overfishing. SSBP is CPUE scaled by percent mature fish in the catch. When the ratio SSBP_t/SSBP_{REF}, or the "SSBP ratio" (SSBPR) for any species drops below a certain limit (SSBPR_{MIN}), that species is considered to be recruitment overfished and management measures will be implemented to reduce fishing mortality on that species. The rule applies only when the SSBPR drops below the SSBPR_{MIN}, but it will continue to apply until the ratio achieves the "SSBP ratio recovery target" (SSBPR_{TARGET}), which is set at a level no less than SSBP_{RMIN}. These two reference points and their associated recruitment overfishing control rule, which prescribe a target fishing mortality rate (F_{RO-REBUILD}) as a function of the SSBPR, are specified as indicated in Table 18. Again, E_{MSY} is used as a proxy for F_{MSY}.

 $\begin{aligned} F_{\text{RO-REBUILD}} & SSBPR_{\text{MIN}} & SSBPR_{\text{TARGET}} \\ F(SSBPR) &= 0 & \text{for SSBPR} \leq 0.10 \\ F(SSBPR) &= 0.2 \, F_{\text{MSY}} & \text{for } 0.10 < SSBPR \leq SSBPR_{\text{MIN}} \\ F(SSBPR) &= 0.4 \, F_{\text{MSY}} & \text{for SSBPR}_{\text{MIN}} < SSBPR \leq SSBPR_{\text{TARGET}} \end{aligned} \qquad 0.20 \qquad 0.30$

Table 18. Rebuilding control rules for the BMUS in the CNMI

1.1.11.2 Current Stock Status

Bottomfish

Biological and other fishery data are poor for all bottomfish species in the Mariana Archipelago. Generally, data are only available on commercial landings by species and CPUE for the multispecies complexes as a whole. At this time, it is not possible to partition these effort measures among the various BMUS. The most recent stock assessment (Langseth et al. 2019) for the CNMI BMUS complex (comprised of 11 species of shallow and deep species of snapper, grouper, jacks, and emperors) was based on estimate of total catch, an abundance index derived

from the nominal CPUE generated from the creel surveys. The assessments used a state-space Bayesian surplus production model within the modeling framework Just Another Bayesian Biomass Assessment (JABBA), which included biological information and fishery-dependent data through 2017. Determinations of overfishing and overfished status can then be made by comparing current biomass and harvest rates to MSY level reference points. To date, the CNMI BMUS is not subject to overfishing and is not overfished.

Table 19. Stock assessment parameters for the CNMI BMUS complex (from Langseth et al. 2019)

| Parameter | Value | Notes | Status |
|--------------------|---------------------|--|--------------------------|
| MSY | 93.6 (48.8-205.3) | Expressed in 1000 lb (with 95% confidence interval) | |
| H ₂₀₁₇ | 0.12 | Expressed in percentage | |
| H _{CR} | 0.167 (0.084-0.315) | Expressed in percentage (with 95% confidence interval) | |
| H/H _{CR} | 0.79 | | No overfishing occurring |
| B ₂₀₁₇ | 569.2 | Expressed in 1000 lb | |
| B _{MSY} | 570.6 (271.8-1,287) | Expressed in 1000 lb (with 95% confidence interval) | |
| B/B _{MSY} | 1.08 | | Not overfished |

1.1.12 Overfishing Limit, Acceptable Biological Catch, and Annual Catch Limits

1.1.12.1 Brief Description of the ACL Process

The Council developed a tiered system of control rules to guide the specification of ACLs and Accountability Measures (AMs; WPRFMC 2011). The process starts with the use of the best scientific information available (BSIA) in the form of, but not limited to, stock assessments, published papers, reports, and/or available data. These data are categorized into the different tiers in the control rule ranging from Tier 1 (i.e., most information available, typically a stock assessment) to Tier 5 (i.e., catch-only information). The control rules are applied to the BSIA. Tiers 1 to 3 involve conducting a Risk of Overfishing Analysis (denoted by P*) to quantify the scientific uncertainties associated with the assessment to specify the Acceptable Biological Catch (ABC), lowering the MSY-based OFL to the ABC. A Social, Ecological, Economic, and Management (SEEM) Uncertainty Analysis is performed to quantify the uncertainties associated with the SEEM factors, and a buffer is used to lower the ABC to an ACL. For Tier 4, which is comprised of stocks with MSY estimates but no active fisheries, the control rule is 91 percent of MSY. For Tier 5, which has catch-only information, the control rule is a one-third reduction in the median catch depending on a qualitative evaluation of stock status via expert opinion. ACL specification can choose from a variety of methods including the above mentioned SEEM analysis or a percentage buffer (i.e., percent reduction from ABC based on expert opinion) or the use of an Annual Catch Target (ACT). Specifications are done on an annual basis, but the Council normally produces a multi-year specification.

The usual AM for CNMI bottomfish fisheries is an overage adjustment in which the next year's ACL is downward adjusted by the amount of overage from the previous ACL based on a three-year running average.

1.1.12.2 Current OFL, ABC, ACL, and Recent Catch

On May 7, 2021, NMFS implemented an ACL of 84,000 lb for CNMI BMUS from 2020 to 2023 (86 FR 24511), and an ACT of 78,000 lb was also implemented. If the recent three-year average catch exceeds the ACT but remains below the ACL, then an overage adjustment would not be applied. The catch shown in Table 20 takes the average catch of the most recent three years as recommended by the Council at its 160th meeting to avoid large fluctuations in catch due to high interannual variability in creel survey estimates.

Table 20. CNMI 2022 ACL table with three-year average catch (lb)

| Fishery | MUS | OFL | ABC | ACL | ACT | Catch |
|------------|--------------------------------------|--------|--------|--------|--------|--------|
| Bottomfish | Bottomfish multi- species complex | 95,000 | 84,000 | 84,000 | 78,000 | 55,916 |

1.1.13 Best Scientific Information Available

1.1.13.1 Bottomfish Fishery

1.1.13.1.1 Stock Assessment Benchmark

The benchmark stock assessment for the Territory BMUS complexes was developed and finalized by Langseth et al. (2019). The assessments used a state-space Bayesian surplus production model within the JABBA modeling framework. Estimates of harvest rate (H), annual biomass (H), the harvest rate associated with overfishing as determined by the harvest control rule (H_{CR}), maximum sustainable yield (H_{CR}), and the biomass at maximum sustainable yield (H_{CR}) allowed for determination of stock status relative to reference points determining overfishing (H_{CR} > 1) and overfished (H_{CR} > 1) status. Stock projections were conducted for 2020–2025 for a range of hypothetical 6-year catches, and the corresponding risk of overfishing was calculated.

1.1.13.1.2 Stock Assessment Updates

Updates to the 2007 benchmark done in 2012 (Brodziak et al. 2012) and 2015 (Yau et al. 2016). These included a two-year stock projection table used for selecting the level of risk the fishery will be managed under ACLs. Yau et al. (2016) was considered the BSIA for the Territory bottomfish MUS complex after undergoing a Western Pacific Stock Assessment Review (WPSAR) Tier 3 panel review (Franklin et al. 2015) prior to the Langseth et al. (2019) benchmark stock assessment. This was the basis for the P* and SEEM analyses that previously determined risk levels to specify past ABCs and ACLs.

1.1.13.1.3 Other Information Available

Approximately every five years, PIFSC administers a socioeconomic survey to small boat fishermen in CNMI. This survey consists of about 60 questions regarding a variety of topics, including fishing experiences, market participation, vessels and gear, demographics and household income, and fishermen perspectives. The survey requests participants to identify which MUS they primarily targeted during the previous 12 months by percentage of trips. Full reports of these surveys can be found at the PIFSC Socioeconomics webpage.

1.1.14 Harvest Capacity and Extent

The MSA defines the term "optimum," with respect to the yield from a fishery, as the amount of fish which:

- Will provide the greatest overall benefit to the Nation, particularly with respect to food production and recreational opportunities, and taking into account the protection of marine ecosystems.
- Is prescribed on the basis of the MSY from the fishery, as reduced by any relevant social, economic, or ecological factor.
- In the case of an overfished fishery, provides for rebuilding to a level consistent with producing the MSY in such fishery [50 CFR §600.310(f)(1)(i)].

Optimum yield (OY) in the bottomfish fisheries is prescribed based on the MSY from the stock assessment and the best available scientific information. In the process of specifying ACLs, social, economic, and ecological factors were considered and the uncertainties around those factors defined the management uncertainty buffer between the ABC and ACL. OY for the bottomfish MUS complex is defined to be the level of harvest equal to the ACL consistent with the goals and objectives of the FEPs and used by the Council to manage the stock.

The Council recognizes that MSY and OY are long-term values whereas the ACLs are yearly snapshots based on the level of fishing mortality at MSY (F_{MSY}). There are situations when the long-term means around MSY are lower than ACLs especially if the stock is known to be productive, relatively pristine, or lightly fished. A stock can have catch levels and rates exceeding that of MSY over the short-term to lower the biomass to a level around the estimated MSY and still not jeopardize the stock. The harvest extent, in this case, is defined as the level of catch harvested in a fishing year relative to the ACL or OY. The harvest capacity is the level of catch remaining in the annual catch limit that can potentially be used for the total allowable level of foreign fishing (TALFF). Table 21 summarizes the harvest extent and harvest capacity for the CNMI, tracking annual catch of BMUS against the most recently implemented ACL (86 FR 24511, May 7, 2021).

Table 21. CNMI proportion of harvest capacity and extent relative to the ACL in 2022

| Fishery | MUS | ACL | Catch | Harvest extent (%) | Harvest capacity (%) |
|------------|----------------------------------|--------|--------|--------------------|----------------------|
| Bottomfish | Bottomfish multi-species complex | 84,000 | 47,564 | 56.6 | 43.4 |

1.1.15 Administrative and Regulatory Actions

NMFS did not implement any management actions for insular fisheries in the CNMI during calendar year 2022.

1.2 GUAM FISHERY DESCRIPTIONS

1.2.1 Bottomfish Fishery

Bottomfish fishing in Guam is a combination of recreational, subsistence, and small-scale commercial fishing. It can be separated into two distinct fisheries separated by depth and species composition. The shallow water complex (< 500 ft.) comprises the largest portion of the total bottomfish harvest and effort, and primarily includes: reef-dwelling snappers of the genera *Lutjanus*, *Aphareus*, and *Aprion*; groupers of the genera *Epinephelus*, *Variola*, and *Cephalopholis*; jacks of the genera *Caranx* and *Carangoides*; Holocentrids (*Myripristis* spp. and *Sargocentron* spp.); emperors of the genera *Lethrinus* and *Gymnocranius*; and Dogtooth Tuna (*Gymnosarda unicolor*). The deep-water complex (>500 ft.) consists primarily of groupers of the genera *Hyporthodus* and *Cephalopholis*, jacks of the genera *Caranx* and *Seriola*, and snappers of the genera *Pristipomoides*, *Etelis*, and *Aphareus*. In recent years, deepwater species have made up a significant portion of the total expanded bottomfish fishing catch.

Most fishers that participate in the bottomfish fishery are either subsistence or part-time commercial fishermen, operate boats less than 25 feet in length, and target primarily the shallow water bottomfish complex. It is not uncommon to intercept fishermen combining bottomfish fishing with other methods such as trolling, spearing, and jigging to maximize their catch. High demand has made it profitable to sell locally caught bottomfish, although overhead costs including fuel and gear may be significant factors for in determining a fisherman's selection of fishing method. The demand for local bottomfish, when combined with environmental pressures, however, may cause stress to local bottomfish stocks.

The majority of bottomfish fishing around Guam takes place on offshore banks, though practically no information exists on the condition of the reefs on offshore banks. On the basis of anecdotal information, most of the offshore banks are in good condition due to their isolation. According to Myers (1997), less than 20 percent of the total coral reef resources harvested in Guam are taken from the exclusive economic zone (EEZ), primarily because the reefs are often associated with less accessible offshore banks. As such, finfish make up most of the catch in the EEZ. Most offshore banks are deep, remote, and subject to strong currents. Generally, these banks are only accessible during calm weather in the summer months (May to August/September). Galvez Bank is the closest and most accessible and, consequently, fished most frequently. In contrast, other banks (White Tuna and Santa Rosa, Rota) are remote and generally are fished only during exceptional weather conditions (Green 1997). Local fishermen report that up to ten commercial boats, with two to three people per boat, and some recreational boats, make use of the banks when the weather is good (Green 1997).

At present, the banks are fished using two methods: bottomfish fishing by hook and line and jigging at night for bigeye scad (*Selar crumenophthalmus*; Myers 1997). In recent years, the estimated annual catch in these fisheries has ranged from 14 to 22 metric tons of shallow bottomfish and 3 to 15 metric tons of bigeye scad (Green 1997). The shallow water component accounted for nearly 68 percent (35,002 to 65,162 lb) of the aggregate bottomfish landings in fiscal years 1992–1994 (Myers 1997). Catch composition of the shallow water bottomfish complex (and coral reef species) is dominated by lethrinids, with a single species (*Lethrinus rubrioperculatus*) alone accounting for 28 percent of the total catch. Other important components of the bottomfish catch include lutjanids, carangids, other lethrinids, and serranids. Holocentrids, mullids, labrids, scombrids, and balistids are minor components of the shallow water bottomfish

complex. It should be noted that at least two of these species (*Aprion virescens* and *Caranx lugubris*) are also found in deeper waters, and as a result comprise a portion of the catch of the deep-water fishery.

Species that are commonly taken in the shallow-bottom fishery of Guam are: Aphareus furca, Aprion virescens, Lutjanus kasmira, L. fulvus, Carangoides orthogrammus, Caranx lugubris, C. melampygus, C. ignobilis, Selar crumenophthalmus, Cephalopholis argus, C. spiloparaea, C. urodeta, Epinephelus fasciatus, Gymnocranius spp., Lethrinus atkinsoni, L. erythracanthus, L. olivaceus, L. rubrioperculatus, L. xanthochilus, Gymnosarda unicolor, Sargocentron spp., Myripristis spp., Variola albimarginata, and V. louti.

Species that are commonly taken in the deep-bottom fishery of Guam are: *Aphareus rutilans*, *Aprion virescens*, *Caranx lugubris*, *Seriola dumerilii*, *Cephalopholis igarashiensis*, *C. sonnerati*, *Hyporthodus octofasciatus*, *Etelis carbunculus*, *E. coruscans*, and *Pristipomoides* spp.

1.2.2 Ecosystem Component (formerly Coral Reef) Fishery

Shore-based fishing accounts for most of the fish and invertebrate harvest from coral reefs around Guam. The coral reef fishery harvests more than 100 species of fish, including members of the families Acanthuridae, Carangidae, Gerreidae, Holocentridae, Kyphosidae, Labridae, Lethrinidae, Lutjanidae, Mugilidae, Mullidae, Scaridae, and Siganidae (Hensley and Sherwood 1993). There are several pulse fisheries for juvenile fish that can be major components of the coral reef fishery, but totals in these can vary year to year. These include juvenile rabbitfish (manahak and lesso'), juvenile jacks (i'e), and juvenile goatfish (ti'ao).

Species that are commonly taken in the coral reef fishery of Guam are: Naso unicornis, N. lituratus, Acanthurus xanthopterus, A. lineatus, A. triostegus, Caranx melampygus, C. papuensis, Selar crumenophthalmus, Gerres acinaces, Myripristis spp., Sargocentron spp., Neoniphon spp., Kyphosus cinerascens, K. vaigiensis, Cheilinus undulatus, Cheilinus spp., Halichoeres spp., Lethrinus harak, L. obseletus, L. atkinsoni, Gnathodentex aurolineatus, Lutjanus fulvus, L. monostigma, L. bohar, L. argentimaculatus, Mulloidichthys flavolineatus, M. vanicolensis (ti'ao), Parupeneus multifasciatus, P. barberinus, P. cyclostomus, Ellechelon vaigiensis, Moolgarda engeli, M. seheli, Chlorurus spilurus, C. frontalis, Scarus psittacus, S. altipinnis, S. rubroviolaceus, S. ghobban, S. schlegeli, Siganus spinus (manahak), and S. argenteus (lesso).

Hook and line is the most common method of fishing for coral reef fish in Guam. In 2022, hook and line fishing accounted for around 73% of fishers and 76% of gear in inshore participation surveys. Throw net (talaya) is the second most common method, accounting for about 16% of fishers and 14% of gear. Other methods include gill net, snorkel spearfishing, surround net, drag net, hooks and gaffs, and gleaning.

Guam has continued to experience high levels of commercial activity targeting reef fish. This has primarily been performed by recent migrants from the Federated States of Micronesia. The fishers are generally hired by retail shops to fish six days per week; there have been as many as eight or nine of these stores open at a time. Gathering commercial sales data from these vendors has been difficult due to vendor anxiety surrounding the reason data is being collected and the lack of perceived benefit to the vendor for reporting sales. There have been several instances during data collection where the vendors were not able to comfortably communicate in English. Data collected from these vendors is of limited value, as fish are not identified to species level, and are frequently labeled simply as "reef fish". In 2022, there was one vendor reporting sales. In

order to improve this situation, the Council, Division of Aquatic and Wildlife Resources (DAWR), and PIFSC partnered to increase vendor participation in the data collection program through the Territory Science Initiative. Extensive training, follow-ups, education, and outreach efforts were conducted to vendors and fishermen to increase participation in data collection.

In 2018, the Council drafted an Amendment 5 to the Mariana Archipelago FEP that reclassified a large number MUS as ECS (WPRFMC 2018). The final rule was published in the *Federal Register* in early 2019 (84 FR 2767, February 8, 2019), and reduced the number of MUS from 227 species/families to 13 in the Mariana Archipelago FEP. All former CREMUS and CMUS were reclassified as ECS that do not require ACL specifications or accountability measures but are still to be monitored regularly to prioritize conservation and management efforts and to improve efficiency of fishery management in the region. All existing management measures, including reporting and record keeping, prohibitions, and experimental fishing regulations apply to ECS. If an ECS stock becomes a target of a federal fishery in the future, NMFS and the Council may consider including that stock as a MUS to actively manage that stock. These species are still regularly monitored via other means (see Sections 1.2.6.3 and 2.3.3).

In 2022, there were 59 Broadcast Notice to Mariners (BNM) regarding military exercises on and around Guam. While most of these do not affect inshore fishing, they do affect access to offshore banks for bottomfishing, and some firing ranges being activated restrict access to inshore fishing locations as well. 126 warnings for area W-517, a large area south and east of Guam, that borders several offshore banks where bottom fishing occurs, were issued in 2022.

1.2.3 Fishery Data Collection System

Guam currently has three fishery-dependent collection programs which can be described as long-term data collection programs with different approaches for gathering important information on fishery harvest methods performed by fishermen. The programs are the shore-based and boat-based data programs and the commercial fishery program. The Sportfish Restoration Grant from the U.S. Fish and Wildlife Service (USFWS) provides the significant portion of the funding for these programs. Training of the fishery staff to collect information is rigorous, and year-end totals are calculated by an expansion process done with in collaboration with NMFS PIFSC. Identification of fish to the species level is the goal of Guam's fishery staff.

The boat- and shore-based creel surveys are part of a long-term program that collects participation, effort, and catch data from fishermen. Collaboration with PIFSC has resulted in a reproducible computer database program that can analyze the data to produce various types of trends that describe status of both charter and non-charter fisheries in federal and local waters. The commercial receipt book program is an important source of information for fish that enter the commercial market; however, obtaining information from dealers has been sporadic, occasionally with less than three dealers providing data. In order to improve this situation, the Council, DAWR, and PIFSC partnered to increase vendor participation in the data collection program through the Territory Science Initiative (TSI).

Oram et al. (2011) and Jasper et al. (2016) describe the fishery data collection process for the offshore program on Guam. In general, DAWR staff collect fishery information through a series of random-stratified surveys for participation (i.e., accounting for fishing effort) and catch interviews (i.e., accounting for catch composition, size frequency, and CPUE). These data are transcribed into the Western Pacific Fisheries Information Network (WPacFIN) database, and the annual catch estimates are expanded from the effort and CPUE information. Monthly

commercial vendor reports are tallied at the end of the year and adjusted based on the coverage estimates provided by the vendor and/or the data collection program staff.

1.2.3.1 Effects of COVID-19 on DAWR Creel Survey Data Collection

2022 marked the return to normal survey schedules for DAWR. Normally, there are six shore-based creel surveys and two participation surveys completed per month. Boat-based creel surveys similarly returned to a normal schedule with eight boat based creel (Table 22).

Table 22. Number of inshore creel and participation surveys completed by DAWR in 2022

| Month | Inshore Creel Surveys Completed | Participation Surveys Completed | Aerial Surveys Completed |
|-----------|------------------------------------|------------------------------------|-----------------------------|
| January | 6 | 2 | 0 |
| February | 6 | 2 | 0 |
| March | 6 | 2 | 0 |
| April | 6 | 2 | 0 |
| May | 6 | 2 | 0 |
| June | 6 | 2 | 0 |
| July | 6 | 2 | 0 |
| August | 6 | 2 | 0 |
| September | 6 | 2 | 0 |
| October | 6 | 2 | 0 |
| November | 6 | 2 | 0 |
| December | 6 | 2 | 0 |
| Total | 72 | 24 | 0 |

1.2.4 Meta-Data Dashboard Statistics

The meta-data dashboard statistics describe the amount of data used or available to calculate the fishery-dependent information. Creel surveys are sampling-based systems that require random-stratified design applied to pre-scheduled surveys. The number of sampling days, participation runs, and catch interviews would determine if there are enough samples to run the expansion algorithm. The trends of these parameters over time may infer survey performance. Monitoring the survey performance is critical for explaining the reliability of the expanded information.

Commercial receipt book information depends on the number of invoices submitted and the number of vendors participating in the program. Variations in these meta-data affect the commercial landing and revenue estimates.

1.2.4.1 Creel Survey Meta-Data Statistics

Calculations:

Sample days: Count of the total number of unique dates found in the boat log sampling date data in boat-based creel surveys.

Catch Interviews: In boat-based creel surveys, count of the total number of data records found in the interview header data (number of interview headers). This is divided into two categories, interviews conducted during scheduled survey days (Regular) and opportunistic interviews (Opportunistic), which are collected on non-scheduled days.

Table 23. Summary of Guam boat-based creel survey meta-data

| \$ 7 | #G 1 D | # Catch Interviews | | |
|-------------|---------------|--------------------|---------------|--|
| Year | # Sample Days | Regular | Opportunistic | |
| 1982 | 46 | 469 | 8 | |
| 1983 | 47 | 431 | 34 | |
| 1984 | 53 | 531 | 0 | |
| 1985 | 66 | 812 | 0 | |
| 1986 | 49 | 522 | 0 | |
| 1987 | 48 | 612 | 0 | |
| 1988 | 48 | 949 | 0 | |
| 1989 | 48 | 931 | 2 | |
| 1990 | 48 | 1,028 | 0 | |
| 1991 | 48 | 1,019 | 1 | |
| 1992 | 48 | 1,110 | 0 | |
| 1993 | 52 | 1,119 | 0 | |
| 1994 | 55 | 1,168 | 0 | |
| 1995 | 96 | 1,613 | 4 | |
| 1996 | 96 | 1,608 | 0 | |
| 1997 | 96 | 1,358 | 0 | |
| 1998 | 96 | 1,581 | 0 | |
| 1999 | 96 | 1,367 | 3 | |
| 2000 | 96 | 1,246 | 1 | |
| 2001 | 96 | 908 | 6 | |
| 2002 | 84 | 610 | 1 | |
| 2003 | 78 | 446 | 0 | |
| 2004 | 95 | 530 | 1 | |
| 2005 | 97 | 552 | 0 | |
| 2006 | 96 | 556 | 0 | |
| 2007 | 96 | 500 | 0 | |
| 2008 | 96 | 571 | 2 | |
| 2009 | 96 | 803 | 0 | |
| 2010 | 96 | 902 | 0 | |
| 2011 | 96 | 645 | 0 | |
| 2012 | 74 | 371 | 0 | |
| 2013 | 96 | 561 | 1 | |
| 2014 | 90 | 635 | 9 | |
| 2015 | 97 | 651 | 13 | |
| 2016 | 93 | 900 | 2 | |
| 2017 | 92 | 820 | 10 | |
| 2018 | 89 | 795 | 11 | |
| 2019 | 93 | 786 | 3 | |

| Vaar | # Sample Days | # Catch Interviews | |
|--------------|---------------|--------------------|---------------|
| Year | | Regular | Opportunistic |
| 2020 | 96 | 349 | 1 |
| 2021 | 96 | 884 | 2 |
| 2022 | 97 | 798 | 0 |
| 10-year avg. | 94 | 718 | 5 |
| 10-year SD | 3 | 161 | 5 |
| 20-year avg. | 93 | 653 | 3 |
| 20-year SD | 6 | 169 | 4 |

1.2.4.2 Commercial Receipt Book Statistics

Calculations:

Vendors: Count of the number of unique buyer codes found in the commercial purchase header data from the Commercial Receipt Book; BMUS vendors are only from vendors that landed BMUS species.

Invoices: Count of the number of unique invoice numbers found in the commercial header data from the Commercial Receipt Book; BMUS vendors are only from vendors that landed BMUS species.

Table 24. Summary of Guam commercial receipt book meta-data

| Year | # Vendors | # Total Invoices Collected | # BMUS Vendors | # BMUS Invoices Collected |
|------|-----------|-------------------------------|-------------------|------------------------------|
| 1982 | n.d. | n.d. | n.d. | n.d. |
| 1983 | 3 | 2,312 | n.d. | n.d. |
| 1984 | 3 | 2,587 | 3 | 48 |
| 1985 | n.d. | n.d. | n.d. | n.d. |
| 1986 | n.d. | n.d. | n.d. | n.d. |
| 1987 | n.d. | n.d. | n.d. | n.d. |
| 1988 | n.d. | n.d. | n.d. | n.d. |
| 1989 | n.d. | n.d. | n.d. | n.d. |
| 1990 | 4 | 2,803 | 3 | 72 |
| 1991 | 3 | 2,512 | n.d. | n.d. |
| 1992 | 3 | 2,737 | n.d. | n.d. |
| 1993 | 3 | 2,664 | n.d. | n.d. |
| 1994 | n.d. | n.d. | n.d. | n.d. |
| 1995 | 3 | 1,565 | n.d. | n.d. |
| 1996 | 6 | 1,965 | 3 | 27 |
| 1997 | 7 | 2,923 | 4 | 41 |
| 1998 | 4 | 3,591 | 3 | 69 |
| 1999 | 5 | 3,410 | 3 | 177 |
| 2000 | 3 | 3,868 | 3 | 174 |
| 2001 | 3 | 4,155 | 3 | 286 |

| Year | # Vendors | # Total Invoices Collected | # BMUS Vendors | # BMUS Invoices Collected |
|--------------|-----------|-------------------------------|-------------------|------------------------------|
| 2002 | 3 | 3,498 | n.d. | n.d. |
| 2003 | n.d. | n.d. | n.d. | n.d. |
| 2004 | 3 | 3,107 | n.d. | n.d. |
| 2005 | 3 | 2,649 | n.d. | n.d. |
| 2006 | 4 | 2,589 | n.d. | n.d. |
| 2007 | n.d. | n.d. | n.d. | n.d. |
| 2008 | n.d. | n.d. | n.d. | n.d. |
| 2009 | n.d. | n.d. | n.d. | n.d. |
| 2010 | n.d. | n.d. | n.d. | n.d. |
| 2011 | n.d. | n.d. | n.d. | n.d. |
| 2012 | n.d. | n.d. | n.d. | n.d. |
| 2013 | n.d. | n.d. | n.d. | n.d. |
| 2014 | 8 | 1,355 | n.d. | n.d. |
| 2015 | 9 | 1,361 | n.d. | n.d. |
| 2016 | 8 | 1,661 | n.d. | n.d. |
| 2017 | 11 | 1,996 | 4 | 104 |
| 2018 | 10 | 1,748 | 4 | 56 |
| 2019 | 6 | 1,200 | n.d. | n.d. |
| 2020 | n.d. | n.d. | n.d. | n.d. |
| 2021 | n.d. | n.d. | n.d. | n.d. |
| 2022 | n.d. | n.d. | n.d. | n.d. |
| 10-year avg. | 6 | 1,227 | 2 | 71 |
| 10-year SD | 4 | 509 | 1 | 33 |
| 20-year avg. | 4 | 1,696 | 2 | 101 |
| 20-year SD | 3 | 715 | 1 | 46 |

^{&#}x27;n.d.' indicates that data are non-disclosed due to confidentiality rules (i.e., less than three dealers and/or vendors).

1.2.5 Fishery Summary Dashboard Statistics

The Fishery Summary Dashboard Statics section consolidates fishery-dependent information comparing the most recent year with short-term (recent 10-year) and long-term (recent 20-year) average (shown bolded in [brackets]). Trend analysis of the past 10 years will dictate the trends (increasing, decreasing, or no trend). The right-most symbol indicates whether the mean of the short-term and long-term years were above, below, or within one standard deviation of the mean of the full time series.

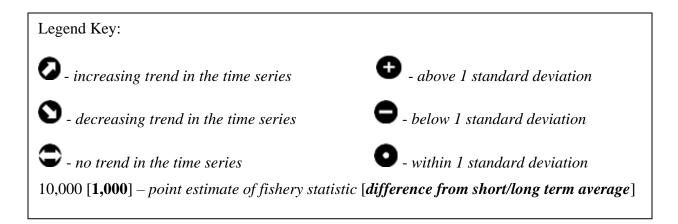


Table 25. Annual indicators for Guam bottomfish fisheries describing performance and comparing 2022 estimates with short- (10-year) and long-term (20-year) averages

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) |
|---------------------------|---|----------------------------|--|
| Bottomfish | Total estimated catch (lb | (a) | |
| All gears | All BMUS from creel survey data | 45,071[▲138%] 🗸 🖸 | 45,071[▲116%] |
| (BMUS only) | All BMUS from commercial purchase data | n.d. | n.d. |
| | Catch-per-unit-effort (fr | om boat-based creel survey | ys) |
| Bottomfish fishing | Bottomfish fishing lb/trip | 21[▲24%] 🗸 🖸 | 21[▲11%] 🛇 🖸 |
| (BMUS only) | Bottomfish fishing lb/gr-hr | 1.40[▲32%] � € | 1.40[▲9%] �� |
| | Fishing effort (from boar | t-based creel surveys) | |
| Bottomfish fishing | Tallied bottomfish trips | 95[▲58%] �� | 95[▲56%] 🕢 🕀 |
| (BMUS only) | Tallied bottomfish gear hours | 1,419[▲27%] ❷ • | 1,419[▲40%] 🏈 🕀 |
| | Fishing participants (fro | m boat-based creel surveys | s) |
| Bottomfish | Tallied number of bottomfish fishing vessels | 63[▲47%] 🕢 🕀 | 63[▲43%] 🗢 <table-cell-rows></table-cell-rows> |
| fishing (BMUS only) | Estimated average number of fishermen per bottomfish fishing trip | 2[▼33%] 🗢 🖨 | 2[▼33%]�� |
| | Bycatch | | |

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) |
|------------|------------------------------|-----------------------|----------------------|
| Bottomfish | Total estimated catch (lb | 0) | |
| BMUS | # fish caught | 1,248[▲100%] | 1,248[▲94%] 🗢 🛨 |
| | # fish discarded/released | 1[no change] | 1[▼75%] 🛇 🔾 |
| | % bycatch | 0.08[▼62%] 🗢 🖸 | 0.08[▼85%] �� |

Table 26. Annual indicators for Guam ECS fisheries describing performance and comparing 2022 estimates with short- (10-year) and long-term (20-year) averages

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) | | | | |
|--------------------|--|--------------------------|----------------------|--|--|--|--|
| ECS | Total estimated boat-based catch (lb) | | | | | | |
| | Naso unicornis from creel survey data | 14,046[▲192%] ② ◆ | 14,046[▲138%] | | | | |
| | Siganus spinus from creel survey data | 415[▼45%] 🗸 🖸 💽 | 415[▼17%] 🗸 🖸 | | | | |
| | Siganus spinus from commercial purchase data | n.d. | n.d. | | | | |
| | Lethrinus harak from creel survey data | 1,227[▼55%] �� | 1,227[▼65%] �� | | | | |
| Prioritized ECS | Chlorurus frontalis from creel survey data | 3,955[▲337%] 🗗 🕀 | 3,955[▲374%] | | | | |
| | Epinephelus fasciatus from creel survey data | 716[▼41%]�� | 716[▼66] �� | | | | |
| | Caranx melampygus from creel survey data | 447[▼82%]�� | 447[▼85%]�� | | | | |
| | Lethrinus olivaceus from creel survey data | 380[▼54%]�� | 380[▼60%]�� | | | | |
| | Lutjanus fulvus | 55[▼84%]�� | 55[▼87%]�� | | | | |

| Fishery | Fishery statistics | Short-term (10 years) | Long-term (20 years) |
|---------|--|-----------------------|----------------------|
| ECS | Total estimated boat-b | ased catch (lb) | |
| | from creel survey data | | |
| | Scarus rubroviolaceus from creel survey data | 1,207[▲361%] 🎝 🖨 | 1,207[▲164%] 🍑 💽 |

1.2.6 Catch Statistics

The following section summarizes the catch statistics for bottomfish, the top ten landed species, and nine prioritized species in Guam as decided by DAWR. Estimates of catch are summarized from the creel survey and commercial receipt book data collection programs. Catch statistics provide estimates of annual harvest from the different fisheries. Estimates of fishery removals can provide proxies for the level of fishing mortality and a reference level relative to established quotas. This section also provides detailed levels of catch for fishing methods and the top species complexes harvested in the ECS and bottomfish fisheries.

1.2.6.1 Catch by Data Stream

This section describes the estimated total catch from the boat-based creel survey programs as well as the commercial landings from the commercial receipt book system. The difference between the creel total and the commercial landings is assumed to be the non-commercial component. However, there are cases where the commercial landing may be higher than the estimated creel total of the commercial receipt book program. In this case, the commercial receipt books can capture fishery data better than the creel surveys.

Calculations: Estimated landings are based on a pre-determined list of species (Appendix A) identified as BMUS regardless of the gear used, for each type of data collection (boat-based creel and the commercial purchase reports).

Table 27. Summary of Guam BMUS total catch (lb) from expanded boat- and shore-based creel surveys and the commercial purchase system for all gear types

| Year | Boat-Based Creel Survey Estimates | Shore-Based Creel Survey Estimates | Total Creel Survey Estimates | Commercial Landings |
|------|--------------------------------------|---------------------------------------|---------------------------------|------------------------|
| 1982 | 20,677 | - | 20,677 | n.d. |
| 1983 | 36,150 | - | 36,150 | n.d. |
| 1984 | 14,655 | - | 14,655 | 3,445 |
| 1985 | 38,960 | 4 | 38,964 | n.d. |
| 1986 | 16,404 | 386 | 16,790 | n.d. |
| 1987 | 24,279 | 12 | 24,291 | n.d. |
| 1988 | 33,986 | 3,092 | 37,078 | n.d. |
| 1989 | 44,799 | 76 | 44,875 | n.d. |
| 1990 | 33,816 | 1,635 | 35,451 | 4,277 |
| 1991 | 31,546 | 1,641 | 33,187 | n.d. |
| 1992 | 36,316 | 2,337 | 38,653 | n.d. |

| Year | Boat-Based Creel Survey Estimates | Shore-Based Creel Survey Estimates | Total Creel Survey Estimates | Commercial Landings |
|-----------------|--------------------------------------|---------------------------------------|---------------------------------|------------------------|
| 1993 | 39,073 | 368 | 39,441 | n.d. |
| 1994 | 40,719 | 222 | 40,941 | n.d. |
| 1995 | 27,194 | 892 | 28,086 | n.d. |
| 1996 | 40,498 | 1 | 40,499 | 1,251 |
| 1997 | 21,255 | 24 | 21,279 | 1,957 |
| 1998 | 22,296 | 34 | 22,330 | 4,576 |
| 1999 | 40,773 | 46 | 40,819 | 20,940 |
| 2000 | 58,640 | 79 | 58,719 | 12,184 |
| 2001 | 43,696 | 34 | 43,730 | 10,554 |
| 2002 | 20,366 | 30 | 20,396 | n.d. |
| 2003 | 29,506 | 0 | 29,506 | n.d. |
| 2004 | 25,233 | 20 | 25,253 | n.d. |
| 2005 | 29,087 | 2 | 29,089 | n.d. |
| 2006 | 33,414 | 3 | 33,417 | n.d. |
| 2007 | 22,576 | 3 | 22,579 | n.d. |
| 2008 | 31,103 | 4 | 31,107 | n.d. |
| 2009 | 35,029 | 46 | 35,075 | n.d. |
| 2010 | 23,928 | 211 | 24,139 | n.d. |
| 2011 | 52,230 | 50 | 52,280 | n.d. |
| 2012 | 17,518 | 4 | 17,522 | n.d. |
| 2013 | 27,277 | 218 | 27,495 | n.d. |
| 2014 | 20,687 | 24 | 20,711 | n.d. |
| 2015 | 10,782 | 73 | 10,855 | n.d. |
| 2016 | 24,479 | 1 | 24,480 | n.d. |
| 2017 | 14,653 | 82 | 14,735 | 4,002 |
| 2018 | 28,364 | 363 | 28,727 | 3,029 |
| 2019 | 28,849 | 143 | 28,992 | n.d. |
| 2020 | 16,844 | 0 | 16,844 | n.d. |
| 2021 | 50,864 | 0 | 50,864 | n.d. |
| 2022 | 44,788* | 283 | 45,071 | n.d. |
| 10-year avg. | 26,759 | 132 | 28,437 | 3,464 |
| 10-year SD | 12,064 | 123 | 10,695 | 2,299 |
| 20-year avg. | 28,361 | 85 | 28,437 | 5,143 |
| 20-year SD | 10,749 | 109 | 10,760 | 2,859 |

^{&#}x27;-' indicates no data available; 'n.d.' indicates that data are non-disclosed due to confidentiality rules (i.e., less than three dealers and/or vendors).

^{*} Value is preliminary and undergoing data validation.

1.2.6.2 Expanded Catch Estimates by Fishing Method

Catch information is provided for the top boat-based fishing methods that comprise most of the annual BMUS catch in Guam.

Calculations: The creel survey catch time series are the sum of the estimated weight for selected gear in all strata for all species and all BMUS species.

Table 28. Total catch time series estimates (lb) for all species and BMUS only using Guam expanded boat-based creel survey data for bottomfish fishing gears

| *7 | Bottomfish | | Spearfishing | g (Snorkel) | Spearfishing (SCUBA) | | |
|------|------------|--------|--------------|-------------|----------------------|-------|--|
| Year | All | BMUS | All | BMUS | All | BMUS | |
| 1982 | 41,329 | 20,677 | 420 | 0 | 0 | 0 | |
| 1983 | 50,415 | 36,150 | 1,355 | 0 | 4,399 | 0 | |
| 1984 | 57,412 | 14,525 | 14,108 | 87 | 5,460 | 43 | |
| 1985 | 88,047 | 36,660 | 18,737 | 481 | 12,761 | 76 | |
| 1986 | 34,515 | 14,904 | 12,545 | 10 | 5,145 | 92 | |
| 1987 | 44,459 | 23,510 | 12,448 | 261 | 7,474 | 198 | |
| 1988 | 67,038 | 32,204 | 24,712 | 1,717 | 10,649 | 50 | |
| 1989 | 79,973 | 43,732 | 30,931 | 46 | 13,985 | 9 | |
| 1990 | 61,401 | 32,827 | 28,871 | 0 | 22,273 | 393 | |
| 1991 | 60,753 | 31,113 | 27,898 | 49 | 37,027 | 339 | |
| 1992 | 78,174 | 33,303 | 35,162 | 179 | 25,226 | 1,938 | |
| 1993 | 107,130 | 37,092 | 39,435 | 0 | 22,848 | 293 | |
| 1994 | 105,283 | 40,310 | 37,554 | 0 27,2 | | 247 | |
| 1995 | 101,075 | 25,125 | 40,554 | 60 | 74,735 | 1,246 | |
| 1996 | 129,708 | 38,618 | 67,446 | 255 | 91,810 | 698 | |
| 1997 | 109,345 | 20,779 | 37,363 | 82 | 41,920 | 177 | |
| 1998 | 99,601 | 21,618 | 56,442 | 272 | 68,198 | 314 | |
| 1999 | 122,930 | 39,717 | 45,200 | 168 | 80,859 | 263 | |
| 2000 | 115,837 | 56,095 | 42,403 | 282 | 116,072 | 1,052 | |
| 2001 | 123,975 | 43,119 | 74,369 | 0 | 65,105 | 535 | |
| 2002 | 55,447 | 19,092 | 21,712 | 39 | 34,766 | 347 | |
| 2003 | 82,224 | 29,057 | 22,649 | 0 | 40,093 | 77 | |
| 2004 | 61,874 | 23,268 | 33,601 | 130 | 50,442 | 1,726 | |
| 2005 | 62,651 | 27,838 | 15,036 | 256 | 27,934 | 896 | |
| 2006 | 89,865 | 32,132 | 12,796 | 1,178 | 4,129 | 0 | |
| 2007 | 57,750 | 20,363 | 18,516 | 357 | 11,316 | 1,835 | |
| 2008 | 59,639 | 30,872 | 29,715 | 124 | 24,647 | 0 | |
| 2009 | 89,997 | 34,369 | 22,669 | 305 | 28,947 | 0 | |
| 2010 | 56,164 | 22,958 | 23,635 | 233 | 1,775 | 0 | |
| 2011 | 88,694 | 50,576 | 26,483 | 0 | 67,431 | 26 | |
| 2012 | 40,214 | 17,518 | 23,986 | 0 | 12,204 | 0 | |

| Year | Botton | nfish | Spearfishing | Spearfishing (Snorkel) | | g (SCUBA) |
|--------------|--------|----------------------|--------------|------------------------|--------|-----------|
| rear | All | BMUS | All | BMUS | All | BMUS |
| 2013 | 42,602 | 14,425 | 20,816 | 0 | 2,771 | 0 |
| 2014 | 69,299 | 18,011 | 28,088 | 274 | 32,316 | 0 |
| 2015 | 29,395 | 10,253 | 22,371 | 0 | 30,654 | 0 |
| 2016 | 51,475 | 23,872 | 28,985 | 376 | 21,517 | 0 |
| 2017 | 46,715 | 14,096 | 17,045 | 17,045 88 | | 0 |
| 2018 | 57,904 | 27,022 | 23,051 | 23,051 130 65,998 | | 672 |
| 2019 | 44,208 | 28,448 | 13,557 | 557 18 15,532 | | 0 |
| 2020 | 33,030 | 16,565 | 8,336 | 25 | 2,801 | 0 |
| 2021 | 86,886 | 50,460 | 30,453 | 92 | 0 | 0 |
| 2022 | 67,662 | 40,643 | 47,882 | 3,144 | 0 | 0 |
| 10-year avg. | 52,918 | 52,918 24,380 24,058 | | 415 | 18,144 | 67 |
| 10-year SD | 16,841 | 12,143 | 10,344 | 917 | 19,658 | 202 |
| 20-year avg. | 60,912 | 26,637 | 23,484 | 337 | 22,518 | 262 |
| 20-year SD | 18,509 | 10,865 | 8,472 694 | | 20,294 | 559 |

1.2.6.3 Top and Prioritized ECS in Boat-Based Fishery Catch

Catch time series can act as indicators of fishery performance. Variations in the catch can be attributed to various factors, and there is no single explanatory variable for the observed trends. A one-year reflection of the top ten harvested species (by weight) is included to monitor which ECS are being caught the most annually. Additionally, Guam DAWR selected nine species that were reclassified as ECS that are still of priority to Guam DAWR for regular monitoring, and complete catch time series of these species are included in the report as well.

Calculations: Catch tallied from the boat-based expanded species composition data combining gear types for all species excluding BMUS and pelagic MUS species.

Table 29a. Top ten landed ECS in Guam from boat-based creel survey data in 2022

| Common Name | Scientific Name | Catch (lb) |
|-------------------------|-----------------------------|------------|
| Bigeye scad | Selar crumenophthalmus | 17,193 |
| Bluespine unicornfish | Naso unicornis | 14,046 |
| Assorted reef fish | Assorted reef fish | 6,361 |
| Giant ruby snapper | Etelis boweni | 6,126 |
| Orangespine unicornfish | Naso lituratus | 5,302 |
| Tan-faced parrotfish | Chlorurus frontalis | 3,955 |
| Parrotfish | Hipposcarus longiceps | 3,480 |
| Shallow bottomfish | Assorted shallow bottomfish | 3,015 |
| Bluebanded surgeonfish | Acanthurus lineatus | 2,805 |
| Bullethead parrotfish | Chlorurus spilurus | 1,757 |

Calculations: Catch tallied from commercial receipt data combining gear types for all species excluding BMUS and pelagic MUS species.

Table 29b. Top ten landed ECS in Guam from estimated commercial landings data in 2022

| Common Name | Scientific Name | Catch (lb) |
|----------------------|------------------------|------------|
| Reef fish | Actinopterygii (class) | 15,755 |
| Bigeye scad (atulai) | Selar crumenophthalmus | 2,976 |
| Parrotfish | Scaridae (family) | 1,279 |
| Mafute (emperor) | Lethrinidae (family) | 844 |
| Unicornfish | Naso spp. | 649 |
| Surgeonfish | Acanthuridae (family) | 550 |
| Miscellaneous | Assorted species | 386 |
| Snapper | Lutjanidae (family) | 282 |
| Rabbitfish (hitting) | Siganus argenteus | 252 |
| Jacks | Carangidae (family) | 225 |

Calculations: Catch tallied from boat-based expanded species composition data for species identified as priority ECS (Appendix A).

Table 30a. Catch (lb) from boat-based expansion data for prioritized species in Guam ECS fisheries

| Year | Naso unicornis | Siganus spinus | Lethrinus harak | Chlorurus frontalis | Epinephelus fasciatus | Caranx melampygus | Lethrinus olivaceus | Lutjanus fulvus | Scarus rubroviolaceus |
|------|-------------------|-------------------|--------------------|------------------------|-----------------------|----------------------|------------------------|--------------------|--------------------------|
| 1982 | 0 | 0 | 0 | 0 | 335 | 490 | 43 | 8 | 0 |
| 1983 | 10 | 0 | 0 | 16 | 1,505 | 670 | 0 | 109 | 0 |
| 1984 | 383 | 0 | 0 | 0 | 669 | 96 | 174 | 0 | 0 |
| 1985 | 1,177 | 0 | 296 | 502 | 3,313 | 2,961 | 765 | 100 | 175 |
| 1986 | 305 | 0 | 33 | 572 | 610 | 512 | 458 | 95 | 288 |
| 1987 | 227 | 66 | 21 | 517 | 1,482 | 1,286 | 77 | 103 | 138 |
| 1988 | 1,219 | 84 | 127 | 2,409 | 3,967 | 869 | 214 | 192 | 1,906 |
| 1989 | 4,402 | 422 | 1,185 | 105 | 2,046 | 1,451 | 397 | 1,269 | 892 |
| 1990 | 4,648 | 670 | 2,628 | 2 | 1,348 | 2,861 | 3,757 | 202 | 628 |
| 1991 | 6,683 | 570 | 2,022 | 225 | 2,827 | 1,936 | 744 | 2,024 | 2,395 |
| 1992 | 15,510 | 418 | 1,544 | 3,157 | 2,126 | 735 | 1,484 | 1,018 | 1,594 |
| 1993 | 5,335 | 2,103 | 2,263 | 181 | 5,950 | 2,087 | 353 | 617 | 1,126 |
| 1994 | 6,089 | 426 | 3,098 | 832 | 2,342 | 2,606 | 5,470 | 3,108 | 809 |
| 1995 | 23,433 | 2,133 | 3,268 | 1,874 | 7,747 | 5,038 | 1,628 | 1,514 | 1,262 |
| 1996 | 40,676 | 935 | 6,523 | 1,221 | 6,017 | 8,961 | 2,700 | 1,853 | 983 |
| 1997 | 18,354 | 1,541 | 6,151 | 197 | 4,581 | 3,843 | 2,073 | 704 | 457 |
| 1998 | 26,540 | 1,464 | 3,293 | 2,478 | 8,678 | 2,913 | 586 | 749 | 708 |
| 1999 | 23,985 | 2,096 | 4,185 | 1,114 | 6,348 | 2,985 | 2,309 | 477 | 495 |
| 2000 | 34,700 | 646 | 4,188 | 78 | 3,607 | 4,846 | 4,081 | 920 | 1,941 |
| 2001 | 17,222 | 989 | 4,705 | 508 | 3,590 | 2,822 | 3,615 | 625 | 940 |
| 2002 | 12,329 | 1,012 | 3,675 | 158 | 2,030 | 4,179 | 11,890 | 172 | 49 |
| 2003 | 8,643 | 740 | 4,108 | 1,911 | 9,998 | 3,376 | 629 | 504 | 830 |
| 2004 | 18,734 | 24 | 5,669 | 30 | 3,608 | 5,622 | 2,700 | 238 | 0 |
| 2005 | 12,089 | 71 | 5,451 | 956 | 1,446 | 4,460 | 1,161 | 104 | 814 |
| 2006 | 1,283 | 192 | 1,960 | 268 | 2,766 | 6,357 | 257 | 297 | 159 |

| Year | Naso unicornis | Siganus spinus | Lethrinus harak | Chlorurus frontalis | Epinephelus fasciatus | Caranx melampygus | Lethrinus olivaceus | Lutjanus fulvus | Scarus rubroviolaceus |
|--------------|-------------------|-------------------|--------------------|------------------------|-----------------------|----------------------|------------------------|--------------------|--------------------------|
| 2007 | 4,848 | 18 | 1,354 | 98 | 2,616 | 1,365 | 799 | 616 | 4,175 |
| 2008 | 10,882 | 1,341 | 1,023 | 1,915 | 1,894 | 5,349 | 179 | 424 | 375 |
| 2009 | 6,588 | 101 | 6,741 | 1,165 | 2,003 | 3,134 | 1,870 | 694 | 0 |
| 2010 | 4,291 | 0 | 4,164 | 847 | 2,061 | 1,751 | 1,454 | 495 | 178 |
| 2011 | 2,341 | 0 | 6,954 | 0 | 2,246 | 1,218 | 1,319 | 1,018 | 0 |
| 2012 | 93 | 15 | 4,781 | 431 | 1,073 | 1,000 | 414 | 791 | 0 |
| 2013 | 3,269 | 158 | 7,195 | 551 | 1,962 | 9,524 | 113 | 324 | 785 |
| 2014 | 5,950 | 344 | 8,231 | 115 | 1,590 | 5,394 | 2,729 | 773 | 0 |
| 2015 | 2,064 | 235 | 2,550 | 0 | 1,917 | 371 | 741 | 324 | 0 |
| 2016 | 2,226 | 614 | 2,132 | 332 | 1,114 | 3,669 | 375 | 144 | 453 |
| 2017 | 711 | 79 | 2,289 | 32 | 1,632 | 2,162 | 356 | 793 | 0 |
| 2018 | 4,578 | 0 | 503 | 1,752 | 672 | 855 | 756 | 134 | 30 |
| 2019 | 5,375 | 418 | 1,909 | 178 | 756 | 1,654 | 905 | 367 | 0 |
| 2020 | 940 | 1,614 | 804 | 1,787 | 1,335 | 187 | 882 | 198 | 0 |
| 2021 | 9,013 | 3,667 | 634 | 349 | 515 | 968 | 1,024 | 273 | 145 |
| 2022 | 14,046 | 415 | 1,227 | 3,955 | 716 | 447 | 380 | 55 | 1,207 |
| 10-year avg. | 4,817 | 754 | 2,747 | 905 | 1,221 | 2,523 | 826 | 339 | 262 |
| 10-year SD | 4,123 | 1,120 | 2,724 | 1,260 | 541 | 2,964 | 731 | 254 | 423 |
| 20-year avg. | 5,898 | 502 | 3,484 | 834 | 2,096 | 2,943 | 952 | 428 | 458 |
| 20-year SD | 4,984 | 869 | 2,502 | 1,004 | 2,017 | 2,499 | 754 | 273 | 948 |

Calculations: Catch tallied from commercial purchase data for species identified as priority ECS (Appendix A). From the prioritized ECS list, only *Siganus spinus* is included because there are no specific species codes for the other eight prioritized species in the Guam commercial coding system, which tends to aggregate data into larger groups such as taxonomic family.

Table 30b. Catch (lb) from commercial purchase data for Siganus spinus in Guam

| Year | Siganus |
|------|---------|
| | spinus |
| 1982 | n.d. |
| 1983 | n.d. |
| 1984 | 32 |
| 1985 | n.d. |
| 1986 | n.d. |
| 1987 | n.d. |
| 1988 | n.d. |
| 1989 | n.d. |
| 1990 | 419 |
| 1991 | n.d. |
| 1992 | n.d. |
| 1993 | n.d. |
| 1994 | n.d. |
| 1995 | n.d. |
| 1996 | 131 |
| 1997 | 84 |
| 1998 | 1,895 |
| 1999 | 3,450 |
| 2000 | 0 |
| 2001 | 15 |
| 2002 | n.d. |
| 2003 | n.d. |
| 2004 | n.d. |
| 2005 | n.d. |
| 2006 | n.d. |
| 2007 | n.d. |
| 2008 | n.d. |
| 2009 | n.d. |
| 2010 | n.d. |
| 2011 | n.d. |
| 2012 | n.d. |
| 2013 | n.d. |
| 2014 | n.d. |
| 2015 | n.d. |
| • | • |

| Year | Siganus spinus |
|--------------|-------------------|
| 2016 | n.d. |
| 2017 | 10,941 |
| 2018 | 6,262 |
| 2019 | n.d. |
| 2020 | n.d. |
| 2021 | n.d. |
| 2022 | n.d. |
| 10-year avg. | 8,602 |
| 10-year SD | 3,309 |
| 20-year avg. | 8,602 |
| 20-year SD | 3,309 |

"n.d." = Confidential (less than three dealers and/or vendors)

1.2.7 Catch-per-Unit-Effort (CPUE) Statistics

This section summarizes the estimates for CPUE in the boat-based fisheries both for all species and for BMUS only. The boat-based fisheries include the bottomfish fishing (handline gear), spearfishing (snorkel), and spearfishing (SCUBA). CPUE is reported as both pounds per gear hour and pounds per fishing trip in the boat-based fishery.

Calculations: CPUE is calculated from interview data by gear type using \sum catch $/\sum$ (number of gears used*number of hours fished) or \sum catch $/\sum$ trips for boat-based data. If the value is blank (i.e., zero), then there was no interview collected for that method. Landings from interviews without fishing hours or number of gears are excluded from the calculations.

All - lb/trip: All catch and trips are tallied from landings by gear level, including non-BMUS species.

All - lb/gr-hr.: All catch and trips are tallied from trips with data on the number of gears used and numbers of hours fished, including non-BMUS species.

BMUS - lb/trip: Only BMUS catch and trips that landed BMUS species are tallied from landings by gear level.

BMUS - lb/gr-hr.: Only BMUS catch and trips that landed BMUS are tallied from trips with data on the number of gears used and numbers of hours fished.

Table 31. CPUE (lb/gear hour and lb/trip) for bottomfish fishing gears in the Guam boat-based fishery for all species and BMUS only

| | | Botto | mfish | | | Spearfish | (Snorkel) | | | Spearfish | (SCUBA) | |
|------|---------|----------|---------|----------|---------|-----------|-----------|------------|---------|-----------|---------|------------|
| Year | A | All | BI | MUS | I | All | BN | IUS | A | All | BN | IUS |
| | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr |
| 1982 | 27 | 2.98 | 17 | 1.77 | 7 | 2.46 | 0 | 0.00 | 0 | 0.00 | 0 | 0.00 |
| 1983 | 23 | 2.95 | 20 | 2.33 | 7 | 1.67 | 0 | 0.00 | 18 | 5.89 | 0 | 0.00 |
| 1984 | 28 | 3.11 | 17 | 2.03 | 39 | 2.32 | 8 | 0.67 | 24 | 4.97 | 1 | 0.33 |
| 1985 | 27 | 2.41 | 17 | 1.49 | 48 | 4.53 | 6 | 0.52 | 25 | 6.58 | 2 | 0.56 |
| 1986 | 23 | 2.32 | 24 | 1.78 | 43 | 4.15 | 1 | 0.20 | 20 | 4.35 | 3 | 0.50 |
| 1987 | 23 | 2.55 | 18 | 1.71 | 28 | 5.46 | 4 | 0.85 | 30 | 6.66 | 3 | 0.53 |
| 1988 | 21 | 2.05 | 13 | 1.11 | 35 | 6.05 | 34 | 8.50 | 20 | 7.44 | 2 | 0.80 |
| 1989 | 20 | 2.10 | 15 | 1.50 | 26 | 3.07 | 1 | 0.19 | 31 | 5.98 | 1 | 0.29 |
| 1990 | 21 | 1.97 | 16 | 1.45 | 22 | 3.66 | 0 | 0.00 | 46 | 11.30 | 6 | 1.00 |
| 1991 | 19 | 2.17 | 16 | 1.76 | 24 | 4.45 | 1 | 0.13 | 47 | 14.43 | 5 | 0.97 |
| 1992 | 17 | 1.88 | 11 | 1.08 | 24 | 3.52 | 3 | 0.50 | 24 | 8.07 | 10 | 2.13 |
| 1993 | 19 | 1.84 | 18 | 1.69 | 21 | 3.37 | 0 | 0.00 | 58 | 19.11 | 5 | 1.27 |
| 1994 | 26 | 2.41 | 21 | 1.73 | 25 | 3.62 | 0 | 0.00 | 55 | 15.06 | 4 | 0.87 |
| 1995 | 13 | 1.00 | 11 | 0.85 | 31 | 3.74 | 3 | 0.25 | 89 | 17.29 | 10 | 1.49 |
| 1996 | 18 | 1.16 | 16 | 1.22 | 33 | 4.21 | 3 | 1.00 | 76 | 11.19 | 7 | 0.46 |
| 1997 | 14 | 0.95 | 11 | 0.72 | 25 | 3.09 | 10 | 4.00 | 81 | 14.57 | 4 | 0.54 |
| 1998 | 14 | 1.01 | 10 | 0.79 | 21 | 2.93 | 5 | 0.32 | 98 | 15.88 | 2 | 0.28 |
| 1999 | 16 | 1.10 | 17 | 1.21 | 17 | 2.08 | 7 | 3.50 | 100 | 14.81 | 2 | 0.31 |
| 2000 | 18 | 1.34 | 19 | 1.27 | 21 | 2.72 | 24 | 24.00 | 90 | 13.98 | 4 | 0.44 |
| 2001 | 20 | 1.65 | 15 | 1.26 | 56 | 4.69 | 21 | 1.31 | 69 | 10.98 | 4 | 0.39 |
| 2002 | 17 | 1.37 | 14 | 1.16 | 21 | 3.01 | 1 | 0.08 | 58 | 6.96 | 12 | 1.28 |
| 2003 | 21 | 1.56 | 16 | 0.95 | 40 | 5.05 | | | 108 | 13.20 | 3 | 0.22 |
| 2004 | 24 | 1.91 | 20 | 1.47 | 28 | 3.42 | 2 | 0.11 | 81 | 9.14 | 11 | 1.03 |
| 2005 | 27 | 2.18 | 31 | 2.23 | 20 | 2.56 | 6 | 1.10 | 61 | 5.55 | 13 | 0.52 |

| | | Botto | nfish | | | Spearfish | (Snorkel) | | | Spearfish | (SCUBA) | |
|---------------|---------|----------|---------|----------|---------|-----------|-----------|------------|---------|-----------|---------|------------|
| Year | A | All | BN | MUS | I | All | BM | 1US | A | All | BN | IUS |
| | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr | lb/trip | lb/gr-hr |
| 2006 | 31 | 2.15 | 26 | 1.43 | 24 | 2.30 | 16 | 1.02 | 13 | 2.69 | 0 | 0.00 |
| 2007 | 30 | 2.22 | 16 | 1.17 | 31 | 3.29 | 4 | 0.42 | 100 | 8.00 | 25 | 1.56 |
| 2008 | 21 | 1.75 | 17 | 1.25 | 38 | 3.05 | 2 | 0.18 | 35 | 4.49 | 0 | 0.00 |
| 2009 | 29 | 2.13 | 25 | 1.85 | 23 | 2.71 | 2 | 0.16 | 63 | 7.00 | 0 | 0.00 |
| 2010 | 17 | 1.21 | 13 | 0.83 | 19 | 2.42 | 1 | 0.20 | 2 | 0.44 | 0 | 0.00 |
| 2011 | 37 | 2.71 | 29 | 2.14 | 41 | 5.17 | 0 | 0.00 | 140 | 11.51 | 1 | 0.17 |
| 2012 | 21 | 2.06 | 18 | 1.62 | 58 | 7.62 | 0 | 0.00 | 70 | 10.00 | 0 | 0.00 |
| 2013 | 19 | 1.53 | 16 | 1.12 | 28 | 2.28 | 0 | 0.00 | 10 | 3.53 | 0 | 0.00 |
| 2014 | 24 | 1.33 | 13 | 0.91 | 35 | 2.39 | 4 | 0.50 | 33 | 8.61 | 0 | 0.00 |
| 2015 | 16 | 1.29 | 15 | 1.14 | 33 | 3.02 | 0 | 0.00 | 58 | 2.70 | 0 | 0.00 |
| 2016 | 21 | 1.49 | 17 | 1.15 | 27 | 2.76 | 4 | 0.29 | 68 | 4.79 | 0 | 0.00 |
| 2017 | 19 | 1.37 | 11 | 0.70 | 16 | 1.92 | 2 | 0.16 | 43 | 5.34 | 0 | 0.00 |
| 2018 | 26 | 0.51 | 21 | 0.37 | 41 | 3.66 | 3 | 0.11 | 97 | 7.18 | 29 | 1.80 |
| 2019 | 20 | 1.67 | 19 | 1.45 | 17 | 1.45 | 1 | 0.13 | 45 | 2.99 | 0 | 0.00 |
| 2020 | 14 | 1.18 | 13 | 0.85 | 9 | 1.07 | 1 | 0.50 | 76 | 4.78 | 0 | 0.00 |
| 2021 | 25 | 1.60 | 27 | 1.50 | 23 | 1.88 | 2 | 0.26 | 0 | 0.00 | 0 | 0.00 |
| 2022 | 22 | 1.61 | 21 | 1.40 | 34 | 2.36 | 8 | 0.44 | 0 | 0.00 | 0 | 0.00 |
| 10-yr avg | 21 | 1.36 | 17 | 1.06 | 26 | 2.28 | 3 | 0.24 | 43 | 3.99 | 3 | 0.18 |
| 10-yr SD | 4 | 0.32 | 5 | 0.34 | 9 | 0.72 | 2 | 0.18 | 31 | 2.63 | 9 | 0.54 |
| 20-yr avg. | 23 | 1.67 | 19 | 1.28 | 29 | 3.02 | 3 | 0.29 | 55 | 5.60 | 4 | 0.26 |
| 20-yr SD | 5 | 0.48 | 6 | 0.45 | 11 | 1.45 | 4 | 0.31 | 38 | 3.63 | 8 | 0.53 |

1.2.8 Effort Statistics

This section summarizes the effort trends in the Guam bottomfish fishery. Fishing effort trends provide insights on the level of fishing pressure through time. Effort information is provided for the top boat-based fishing methods that comprise most of the annual catch.

Calculations: Effort estimates (in both trips and gear hours) are calculated from boat-based interview data. Trips are tallied according the interview data in boat-based creel surveys. Gear hours are generated by summing the data on number of gears used*number of hours fished collected from interviews by gear type. For the boat-based estimates, data collection started in 1982.

All - Trips: All trips tallied by gear type.

All - Gear-hr: Gear hours tallied by gear type.

BMUS - Trips: Trips that landed BMUS tallied by gear type.

BMUS - Gear-hr: Gear hours tallied by gear type for trips landed BMUS with data on both number of gears used and numbers of hours fished.

Table 32. Effort (trips and gear hours) for bottomfish fishing gears in the Guam boat-based fishery for all species and BMUS only

| | | Bott | omfish | | | Spearfish (| Snorkel) | | | Spearfish | (SCUBA) | |
|------|-------|-------|--------|-------|-------|-------------|----------|------------|-------|-----------|---------|------------|
| Year | A | All | BM | US | A | 11 | BN | AUS | | All | BN | AUS |
| | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr |
| 1982 | 97 | 869 | 74 | 715 | 5 | 15 | 0 | 0 | 1 | 1 | 0 | 0 |
| 1983 | 89 | 683 | 66 | 566 | 6 | 24 | 0 | 0 | 13 | 40 | 0 | 0 |
| 1984 | 124 | 1,118 | 39 | 328 | 20 | 336 | 1 | 12 | 12 | 57 | 1 | 3 |
| 1985 | 217 | 2,391 | 139 | 1,635 | 19 | 203 | 4 | 42 | 36 | 139 | 3 | 9 |
| 1986 | 103 | 1,024 | 41 | 543 | 14 | 145 | 1 | 5 | 8 | 38 | 1 | 6 |
| 1987 | 114 | 1,041 | 72 | 758 | 20 | 101 | 3 | 13 | 11 | 50 | 3 | 15 |
| 1988 | 173 | 1,776 | 137 | 1,542 | 33 | 190 | 2 | 8 | 25 | 67 | 2 | 5 |
| 1989 | 187 | 1,790 | 127 | 1,307 | 24 | 204 | 3 | 16 | 24 | 123 | 1 | 4 |
| 1990 | 157 | 1,660 | 108 | 1,219 | 18 | 107 | 0 | 0 | 17 | 70 | 1 | 6 |
| 1991 | 152 | 1,316 | 92 | 852 | 20 | 109 | 2 | 16 | 27 | 89 | 5 | 24 |
| 1992 | 152 | 1,368 | 98 | 1,013 | 30 | 205 | 1 | 6 | 48 | 146 | 3 | 14 |
| 1993 | 164 | 1,700 | 81 | 842 | 38 | 242 | 0 | 0 | 29 | 87 | 4 | 15 |
| 1994 | 185 | 2,028 | 105 | 1,282 | 37 | 251 | 0 | 0 | 32 | 116 | 5 | 21 |
| 1995 | 302 | 3,860 | 127 | 1,613 | 56 | 464 | 1 | 12 | 56 | 287 | 8 | 56 |
| 1996 | 277 | 4,173 | 97 | 1,284 | 62 | 482 | 2 | 6 | 48 | 327 | 5 | 75 |
| 1997 | 238 | 3,554 | 75 | 1,183 | 41 | 328 | 1 | 3 | 27 | 150 | 2 | 13 |
| 1998 | 315 | 4,311 | 125 | 1,551 | 96 | 700 | 4 | 66 | 40 | 246 | 6 | 50 |
| 1999 | 285 | 4,039 | 112 | 1,549 | 51 | 428 | 1 | 2 | 43 | 290 | 9 | 65 |
| 2000 | 200 | 2,676 | 92 | 1,345 | 47 | 366 | 1 | 1 | 41 | 265 | 8 | 72 |
| 2001 | 197 | 2,337 | 95 | 1,161 | 22 | 261 | 1 | 16 | 29 | 182 | 4 | 38 |
| 2002 | 150 | 1,861 | 73 | 878 | 29 | 202 | 1 | 12 | 11 | 92 | 2 | 18 |
| 2003 | 107 | 1,411 | 55 | 905 | 22 | 175 | 0 | 0 | 13 | 106 | 2 | 23 |
| 2004 | 112 | 1,432 | 60 | 837 | 17 | 138 | 2 | 27 | 11 | 97 | 3 | 31 |
| 2005 | 121 | 1,510 | 69 | 946 | 24 | 186 | 2 | 10 | 7 | 76 | 1 | 25 |

| | | Bott | omfish | | | Spearfish (| Snorkel) | | Spearfish (SCUBA) | | | |
|--------------|-------|-------|--------|-------|-------|-------------|----------|------------|-------------------|-------|-------|------------|
| Year | A | All | BM | IUS | A | 11 | BN | IUS | | All | BN | IUS |
| | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr | Trips | Gr-hr |
| 2006 | 104 | 1,519 | 61 | 1,123 | 19 | 198 | 2 | 32 | 5 | 25 | 0 | 0 |
| 2007 | 84 | 1,126 | 55 | 745 | 13 | 121 | 2 | 19 | 2 | 25 | 1 | 16 |
| 2008 | 104 | 1,226 | 57 | 792 | 26 | 322 | 3 | 34 | 6 | 47 | 0 | 0 |
| 2009 | 146 | 1,979 | 76 | 1,019 | 28 | 233 | 4 | 43 | 3 | 27 | 0 | 0 |
| 2010 | 165 | 2,287 | 96 | 1,460 | 27 | 207 | 4 | 20 | 1 | 5 | 0 | 0 |
| 2011 | 101 | 1,373 | 62 | 840 | 15 | 118 | 0 | 0 | 4 | 49 | 1 | 6 |
| 2012 | 53 | 530 | 32 | 353 | 8 | 61 | 0 | 0 | 3 | 21 | 0 | 0 |
| 2013 | 60 | 763 | 31 | 437 | 12 | 148 | 0 | 0 | 3 | 9 | 0 | 0 |
| 2014 | 92 | 1,625 | 46 | 604 | 17 | 205 | 1 | 8 | 3 | 12 | 0 | 0 |
| 2015 | 73 | 887 | 34 | 432 | 17 | 184 | 0 | 0 | 4 | 86 | 0 | 0 |
| 2016 | 106 | 1,506 | 62 | 927 | 25 | 241 | 2 | 24 | 22 | 313 | 0 | 0 |
| 2017 | 115 | 1,573 | 69 | 1,073 | 31 | 256 | 2 | 19 | 4 | 32 | 0 | 0 |
| 2018 | 99 | 5,010 | 54 | 3,053 | 19 | 215 | 2 | 45 | 16 | 216 | 3 | 49 |
| 2019 | 127 | 1,525 | 76 | 1,016 | 20 | 217 | 1 | 8 | 6 | 91 | 0 | 0 |
| 2020 | 74 | 858 | 42 | 626 | 17 | 149 | 1 | 2 | 3 | 48 | 0 | 0 |
| 2021 | 151 | 2,390 | 90 | 1,628 | 44 | 532 | 4 | 35 | 0 | 0 | 0 | 0 |
| 2022 | 143 | 1,935 | 95 | 1,419 | 41 | 596 | 9 | 164 | 0 | 0 | 0 | 0 |
| 10-year avg. | 104 | 1,807 | 60 | 1,121 | 24 | 274 | 2 | 30 | 6 | 81 | 0 | 5 |
| 10-year SD | 29 | 1,171 | 21 | 747 | 10 | 149 | 3 | 47 | 7 | 99 | 1 | 15 |
| 20-year avg. | 107 | 1,623 | 61 | 1,012 | 22 | 225 | 2 | 24 | 6 | 64 | 1 | 7 |
| 20-year SD | 29 | 905 | 19 | 572 | 9 | 126 | 2 | 35 | 5 | 76 | 1 | 14 |

1.2.9 Participants

This section summarizes the estimated participation in each fishery. The information presented here can be used in the impact analysis of potential amendments in the FEPs associated with the bottomfish fisheries. The trend in participation can also be used as an indicator for fishing pressure.

Calculations: For boat-based data, the estimated number of unique vessels is calculated by tallying the number of vessels recorded in the interview data via vessel registration or name.

All: Total unique vessels by gear type.

BMUS: Unique vessels from trips that landed BMUS by gear type.

Table 33a. Estimated number of unique vessels for bottomfish fishing gears in the Guam boat-based fishery for all species and BMUS only

| D 44 C 1 G C | | | | | | | | | | | |
|--|-----|--------|-----|-----------|-----|---------|--|--|--|--|--|
| Year | | omfish | _ | (Snorkel) | _ | (SCUBA) | | | | | |
| Tear | All | BMUS | All | BMUS | All | BMUS | | | | | |
| 1982 | 58 | 47 | 4 | 0 | 1 | 0 | | | | | |
| 1983 | 51 | 41 | 5 | 0 | 4 | 0 | | | | | |
| 1984 | 75 | 33 | 13 | 1 | 6 | 1 | | | | | |
| 1985 | 97 | 66 | 9 | 3 | 21 | 3 | | | | | |
| 1986 | 62 | 27 | 12 | 1 | 7 | 1 | | | | | |
| 1987 | 71 | 42 | 14 | 3 | 8 | 2 | | | | | |
| 1988 | 92 | 76 | 22 | 2 | 14 | 1 | | | | | |
| 1989 | 100 | 70 | 20 | 3 | 18 | 1 | | | | | |
| 1990 | 87 | 58 | 17 | 0 | 9 | 1 | | | | | |
| 1991 | 96 | 65 | 19 | 2 | 19 | 4 | | | | | |
| 1992 | 88 | 62 | 23 | 1 | 29 | 3 | | | | | |
| 1993 | 116 | 53 | 25 | 0 | 20 | 4 | | | | | |
| 1994 | 122 | 71 | 32 | 0 | 22 | 4 | | | | | |
| 1995 | 170 | 82 | 39 | 1 | 30 | 5 | | | | | |
| 1996 | 148 | 68 | 44 | 2 | 28 | 3 | | | | | |
| 1997 | 126 | 51 | 31 | 1 | 18 | 2 | | | | | |
| 1998 | 153 | 72 | 54 | 4 | 20 | 4 | | | | | |
| 1999 | 152 | 69 | 44 | 1 | 16 | 6 | | | | | |
| 2000 | 107 | 61 | 35 | 1 | 21 | 5 | | | | | |
| 2001 | 131 | 73 | 18 | 1 | 16 | 3 | | | | | |
| 2002 | 104 | 58 | 24 | 1 | 9 | 2 | | | | | |
| 2003 | 80 | 48 | 21 | 0 | 9 | 2 | | | | | |
| 2004 | 83 | 47 | 16 | 2 | 5 | 2 | | | | | |
| 2005 | 78 | 42 | 16 | 2 | 6 | 1 | | | | | |
| 2006 | 72 | 45 | 18 | 2 | 4 | 0 | | | | | |
| 2007 | 58 | 41 | 11 | 2 | 2 | 1 | | | | | |

| Year | Botto | mfish | Spearfish | (Snorkel) | Spearfish | (SCUBA) |
|--------------|-------|-------|-----------|-----------|-----------|---------|
| 1 ear | All | BMUS | All | BMUS | All | BMUS |
| 2008 | 78 | 44 | 19 | 3 | 3 | 0 |
| 2009 | 98 | 49 | 25 | 4 | 3 | 0 |
| 2010 | 103 | 61 | 22 | 4 | 1 | 0 |
| 2011 | 72 | 44 | 14 | 0 | 3 | 1 |
| 2012 | 46 | 29 | 8 | 0 | 2 | 0 |
| 2013 | 48 | 28 | 12 | 0 | 3 | 0 |
| 2014 | 69 | 39 | 12 | 1 | 3 | 0 |
| 2015 | 60 | 26 | 15 | 0 | 2 | 0 |
| 2016 | 75 | 41 | 18 | 2 | 10 | 0 |
| 2017 | 85 | 54 | 26 | 2 | 2 | 0 |
| 2018 | 67 | 37 | 16 | 2 | 7 | 3 |
| 2019 | 84 | 52 | 13 | 1 | 3 | 0 |
| 2020 | 63 | 35 | 14 | 1 | 3 | 0 |
| 2021 | 93 | 55 | 28 | 3 | 0 | 0 |
| 2022 | 92 | 63 | 29 | 9 | 0 | 0 |
| 10-year avg. | 74 | 43 | 18 | 2 | 3 | 0 |
| 10-year SD | 14 | 12 | 6 | 2 | 3 | 1 |
| 20-year avg. | 75 | 44 | 18 | 2 | 4 | 1 |
| 20-year SD | 15 | 10 | 6 | 2 | 3 | 1 |

Calculations: For boat-based data, the estimated number of fishermen per trip is calculated by filtering interviews that recorded the number of fishers, and then \sum fishers/ \sum trips. A blank cell indicates insufficient data to generate an estimate of average fishers.

All: Average fishers from all trips by gear type.

BMUS: Average fishers from trips that landed BMUS by gear type.

Table 33b. Estimated number of fishermen per trip for bottomfish fishing gears in the Guam boat-based fishery for all species and BMUS only

| Year | Botte | omfish | Spearfish (| (Snorkel) | Spearfish | n (SCUBA) |
|------|-------|--------|-------------|-----------|-----------|-----------|
| rear | All | BMUS | All | BMUS | All | BMUS |
| 1982 | 2 | 2 | 3 | 0 | 1 | 0 |
| 1983 | 2 | 2 | 2 | 0 | 1 | 0 |
| 1984 | 3 | 3 | 4 | 3 | 2 | 1 |
| 1985 | 3 | 3 | 4 | 3 | 2 | 1 |
| 1986 | 3 | 2 | 3 | 1 | 3 | 2 |
| 1987 | 2 | 2 | 2 | 1 | 2 | 2 |
| 1988 | 3 | 3 | 3 | 2 | 2 | 1 |
| 1989 | 3 | 3 | 3 | 2 | 3 | 3 |
| 1990 | 3 | 3 | 4 | 0 | 3 | 4 |

| Voor | Bott | omfish | Spearfish | (Snorkel) | Spearfisl | h (SCUBA) |
|--------------|------|--------|-----------|-----------|-----------|-----------|
| Year | All | BMUS | All | BMUS | All | BMUS |
| 1991 | 3 | 3 | 3 | 3 | 3 | 4 |
| 1992 | 3 | 3 | 4 | 1 | 3 | 3 |
| 1993 | 3 | 3 | 3 | 0 | 4 | 4 |
| 1994 | 3 | 3 | 3 | 0 | 4 | 4 |
| 1995 | 4 | 3 | 3 | 2 | 4 | 5 |
| 1996 | 5 | 3 | 3 | 1 | 4 | 6 |
| 1997 | 6 | 4 | 3 | 5 | 4 | 4 |
| 1998 | 4 | 3 | 3 | 4 | 4 | 5 |
| 1999 | 4 | 3 | 3 | 2 | 4 | 4 |
| 2000 | 4 | 3 | 3 | 2 | 4 | 4 |
| 2001 | 3 | 2 | 3 | 2 | 4 | 5 |
| 2002 | 3 | 2 | 3 | 2 | 4 | 4 |
| 2003 | 3 | 3 | 4 | 0 | 4 | 4 |
| 2004 | 4 | 3 | 3 | 6 | 4 | 4 |
| 2005 | 3 | 2 | 3 | 3 | 3 | 5 |
| 2006 | 3 | 2 | 3 | 3 | 3 | 0 |
| 2007 | 4 | 3 | 3 | 2 | 4 | 4 |
| 2008 | 3 | 2 | 3 | 3 | 3 | 0 |
| 2009 | 3 | 2 | 3 | 3 | 4 | 0 |
| 2010 | 3 | 3 | 3 | 3 | 3 | 0 |
| 2011 | 3 | 3 | 4 | 0 | 4 | 3 |
| 2012 | 3 | 3 | 3 | 0 | 5 | 0 |
| 2013 | 3 | 3 | 4 | 0 | 3 | 0 |
| 2014 | 3 | 3 | 4 | 4 | 3 | 0 |
| 2015 | 4 | 4 | 4 | 0 | 7 | 0 |
| 2016 | 3 | 3 | 3 | 2 | 5 | 0 |
| 2017 | 2 | 2 | 3 | 3 | 5 | 0 |
| 2018 | 4 | 3 | 4 | 4 | 5 | 3 |
| 2019 | 3 | 3 | 4 | 5 | 7 | 0 |
| 2020 | 3 | 3 | 4 | 6 | 6 | 0 |
| 2021 | 3 | 3 | 4 | 4 | 0 | 0 |
| 2022 | 3 | 2 | 4 | 5 | 0 | 0 |
| 10-year avg. | 3 | 3 | 4 | 3 | 4 | 0 |
| 10-year SD | 1 | 1 | 0 | 2 | 2 | 1 |
| 20-year avg. | 3 | 3 | 4 | 3 | 4 | 1 |
| 20-year SD | 0 | 1 | 0 | 2 | 2 | 2 |

1.2.10 Bycatch Estimates

This section focuses on MSA § 303(a)(11), which requires that all FMPs establish a standardized reporting methodology to assess the amount and type of bycatch occurring in the fishery, and include conservation and management measures that, to the extent practicable, minimize bycatch and bycatch mortality. The MSA § 303(a)(11) standardized reporting methodology is commonly referred to as a "Standardized Bycatch Reporting Methodology" (SBRM) and was added to the MSA by the Sustainable Fisheries Act of 1996 (SFA). The Council implemented omnibus amendments to FMPs in 2003 to address MSA bycatch provisions and establish SBRMs.

The following are recent bycatch estimates for the boat-based BMUS and non-BMUS fisheries. The bycatch estimates presented here are self-reported by fishers during creel survey interviews, and thus, the data are likely biased downward.

Calculations: The number caught is the sum of the total number of individuals found in the raw data including bycatch. The number discarded or released is number of individuals marked as bycatch. Percent bycatch is the sum of all released divided by the number caught.

Table 34. Time series of observed catch and bycatch in Guam boat-based fisheries

| | | BMUS | | ľ | Non-BMUS | S | BMU | JS + Non-B | MUS |
|------|-------------|----------------------|--------------|-------------|----------------------|--------------|-------------|----------------------|--------------|
| Year | # Caught | # Discard or Release | % Bycatch | # Caught | # Discard or Release | % Bycatch | # Caught | # Discard or Release | % Bycatch |
| 1982 | 1,062 | 0 | 0.00 | 535 | 0 | 0.00 | 1,597 | 0 | 0.00 |
| 1983 | 940 | 0 | 0.00 | 567 | 0 | 0.00 | 1,507 | 0 | 0.00 |
| 1984 | 590 | 0 | 0.00 | 2,757 | 0 | 0.00 | 3,347 | 0 | 0.00 |
| 1985 | 1,830 | 0 | 0.00 | 3,010 | 0 | 0.00 | 4,840 | 0 | 0.00 |
| 1986 | 546 | 0 | 0.00 | 1,078 | 0 | 0.00 | 1,624 | 0 | 0.00 |
| 1987 | 1,313 | 0 | 0.00 | 1,206 | 0 | 0.00 | 2,519 | 0 | 0.00 |
| 1988 | 1,399 | 0 | 0.00 | 1,603 | 0 | 0.00 | 3,002 | 0 | 0.00 |
| 1989 | 2,028 | 0 | 0.00 | 1,534 | 0 | 0.00 | 3,562 | 0 | 0.00 |
| 1990 | 1,542 | 0 | 0.00 | 1,328 | 0 | 0.00 | 2,870 | 0 | 0.00 |
| 1991 | 1,366 | 0 | 0.00 | 1,417 | 0 | 0.00 | 2,783 | 0 | 0.00 |
| 1992 | 1,046 | 0 | 0.00 | 1,481 | 0 | 0.00 | 2,527 | 0 | 0.00 |
| 1993 | 946 | 0 | 0.00 | 1,947 | 0 | 0.00 | 2,893 | 0 | 0.00 |
| 1994 | 1,663 | 0 | 0.00 | 2,067 | 0 | 0.00 | 3,730 | 0 | 0.00 |
| 1995 | 1,449 | 0 | 0.00 | 3,536 | 0 | 0.00 | 4,985 | 0 | 0.00 |
| 1996 | 1,281 | 0 | 0.00 | 3,963 | 0 | 0.00 | 5,244 | 0 | 0.00 |
| 1997 | 983 | 0 | 0.00 | 3,359 | 0 | 0.00 | 4,342 | 0 | 0.00 |
| 1998 | 993 | 0 | 0.00 | 4,145 | 0 | 0.00 | 5,138 | 0 | 0.00 |
| 1999 | 1,081 | 0 | 0.00 | 3,857 | 0 | 0.00 | 4,938 | 0 | 0.00 |
| 2000 | 1,090 | 6 | 0.55 | 2,815 | 526 | 18.69 | 3,905 | 532 | 13.62 |
| 2001 | 1,023 | 16 | 1.56 | 2,873 | 607 | 21.13 | 3,896 | 623 | 15.99 |
| 2002 | 629 | 2 | 0.32 | 1,875 | 351 | 18.72 | 2,504 | 353 | 14.10 |

| | | BMUS | | 1 | Non-BMUS | 8 | BMU | JS + Non-B | MUS |
|---------------|-------------|----------------------|--------------|-------------|----------------------|--------------|-------------|----------------------|--------------|
| Year | # Caught | # Discard or Release | % Bycatch | # Caught | # Discard or Release | % Bycatch | # Caught | # Discard or Release | % Bycatch |
| 2003 | 497 | 20 | 4.02 | 1,391 | 171 | 12.29 | 1,888 | 191 | 10.12 |
| 2004 | 586 | 0 | 0.00 | 1,218 | 122 | 10.02 | 1,804 | 122 | 6.76 |
| 2005 | 616 | 0 | 0.00 | 1,090 | 66 | 6.06 | 1,706 | 66 | 3.87 |
| 2006 | 1,140 | 27 | 2.37 | 1,048 | 118 | 11.26 | 2,188 | 145 | 6.63 |
| 2007 | 417 | 7 | 1.68 | 955 | 132 | 13.82 | 1,372 | 139 | 10.13 |
| 2008 | 572 | 3 | 0.52 | 1,085 | 118 | 10.88 | 1,657 | 121 | 7.30 |
| 2009 | 860 | 0 | 0.00 | 1,991 | 77 | 3.87 | 2,851 | 77 | 2.70 |
| 2010 | 890 | 0 | 0.00 | 1,698 | 29 | 1.71 | 2,588 | 29 | 1.12 |
| 2011 | 707 | 0 | 0.00 | 1,421 | 45 | 3.17 | 2,128 | 45 | 2.11 |
| 2012 | 309 | 0 | 0.00 | 615 | 37 | 6.02 | 924 | 37 | 4.00 |
| 2013 | 293 | 0 | 0.00 | 929 | 44 | 4.74 | 1,222 | 44 | 3.60 |
| 2014 | 658 | 6 | 0.91 | 1,794 | 163 | 9.09 | 2,452 | 169 | 6.89 |
| 2015 | 366 | 0 | 0.00 | 1,054 | 70 | 6.64 | 1,420 | 70 | 4.93 |
| 2016 | 641 | 2 | 0.31 | 1,033 | 45 | 4.36 | 1,674 | 47 | 2.81 |
| 2017 | 766 | 0 | 0.00 | 1,547 | 26 | 1.68 | 2,313 | 26 | 1.12 |
| 2018 | 406 | 2 | 0.49 | 1,115 | 27 | 2.42 | 1,521 | 29 | 1.91 |
| 2019 | 865 | 3 | 0.35 | 982 | 44 | 4.48 | 1,847 | 47 | 2.54 |
| 2020 | 302 | 0 | 0.00 | 525 | 16 | 3.05 | 827 | 16 | 1.93 |
| 2021 | 693 | 0 | 0.00 | 1,253 | 5 | 0.40 | 1,946 | 5 | 0.26 |
| 2022 | 1,248 | 1 | 0.08 | 708 | 15 | 2.12 | 1,956 | 16 | 0.82 |
| 10-yr avg. | 624 | 1 | 0.21 | 1,094 | 46 | 3.90 | 1,718 | 47 | 2.68 |
| 10-yr SD | 283 | 2 | 0.29 | 352 | 43 | 2.43 | 468 | 45 | 1.91 |
| 20-yr avg. | 642 | 4 | 0.54 | 1,173 | 69 | 5.90 | 1,814 | 72 | 4.08 |
| 20-yr SD | 260 | 7 | 1.01 | 371 | 50 | 3.89 | 511 | 54 | 2.90 |

1.2.11 Federal Logbook Data

1.2.11.1 Number of Federal Permit Holders

The CFR, Title 50, Part 665 requires the following federal permits for Guam fisheries in the EEZ under the Mariana Archipelago FEP. Regulations governing fisheries under this FEP are in the CFR, Title 50, Part 665

1.2.11.1.1 Guam Large Vessel Bottomfish Permit

Regulations require this permit for any large vessel (50 feet or longer in overall length) fishing for, landing, or transshipping BMUS or bottomfish ECS in the EEZ seaward of Guam.

1.2.11.1.2 Special Coral Reef Ecosystem Permit

Regulations require the coral reef ecosystem special permit for anyone fishing for coral reef ECS in a low-use marine protected area (MPA), fishing for species on the list of Potentially Harvested Coral Reef Taxa or using fishing gear not specifically allowed in the regulations. NMFS will make an exception to this permit requirement for any person issued a permit to fish under any FEP who incidentally catches Guam coral reef ECS while fishing for bottomfish MUS, crustacean ECS, western Pacific pelagic MUS, precious coral, or seamount groundfish. Regulations require a transshipment permit for any receiving vessel used to land or transship potentially harvested coral reef taxa, or any coral reef ecosystem ECS caught in a low-use MPA.

1.2.11.1.3 Western Pacific Precious Corals Permit

Regulations require this permit for anyone harvesting or landing black, bamboo, pink, red, or gold corals in the EEZ in the Western Pacific.

1.2.11.1.4 Western Pacific Crustaceans Permit (Lobster or Deepwater Shrimp)

Regulations require a permit by the owner of a U.S. fishing vessel used to fish for lobster or deep-water shrimp in the EEZ around American Samoa, Guam, CNMI, Hawaii, and the PRIA.

There is no record of special coral reef or precious coral fishery permits issued for the EEZ around Guam since 2007. Table 35 provides the number of permits issued for Guam fisheries between 2013 and 2022. Data are from the NMFS PIRO SFD permits program.

Table 35. Number of federal permits holders for the crustacean and bottomfish fisheries of Guam

| Guam Fisheries | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Bottomfish | 2 | 2 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
| Lobster | 0 | 0 | 0 | 1* | 0 | 1* | 0 | 0 | 0 | 0 |
| Shrimp | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |

Source: PIRO SFD unpublished data.

1.2.11.2 Summary of Catch and Effort for FEP Fisheries

The Marianas Archipelago FEP requires fishermen to obtain a federal permit to fish for certain MUS or ECS in federal waters and to report all catch and discards. While NMFS annually issues permits for various FEP fisheries, there is currently limited data available on the level of catch or effort made by federal non-longline permit holders. Determining the level of fishing activity through the required federal logbook reporting for each fishery helps establish the level of non-longline fishing occurring in federal waters to assess whether there is a continued need for active conservation and management measures (e.g., annual catch limits) for these fisheries. For each FEP fishery, the number of federal permits issued since the federal permit and logbook reporting requirements became effective as well as available catch and effort data are presented in Table 36 through Table 38.

NMFS has never issued a federal permit for precious coral or coral reef fishing in federal waters around Guam. Therefore, catch and effort data is not presented for these fisheries.

^{*}Permits apply to multiple areas and may include American Samoa, Guam, CNMI, and PRIA.

1.2.11.2.1 Large Vessel Bottomfish Fishery

Table 36. Summary of federal logbook data for the Guam large vessel bottomfish fishery

| Year | No. of Federal Bottomfish Permits Issued ¹ | Bottomiisn Pormits | No. of Trips in Guam | Total R Logbook | | Total Reported Logbook MUS Release/Discard (lb) | |
|-------|---|-----------------------|----------------------------|---|--------------------------------|---|--------------------------------|
| 1 car | | Permits Reporting | | Bottomfish MUS & ECS ² | Coral Reef ECS ² | Bottomfish MUS & ECS ² | Coral Reef ECS ² |
| 2006 | 0 | - | | | | | |
| 2007 | 1 | 0 | | | | | |
| 2008 | 2 | 0 | | | | | |
| 2009 | 1 | 0 | | | | | |
| 2010 | 6 | 0 | | | | | |
| 2011 | 6 | 0 | | | | | |
| 2012 | 2 | 0 | | | | | |
| 2013 | 2 | 0 | | | | | |
| 2014 | 2 | 0 | | | | | |
| 2015 | 1 | 0 | | | | | |
| 2016 | 1 | 0 | | | | | |
| 2017 | 1 | 0 | | | | | |
| 2018 | 1 | 0 | | | | | |
| 2019 | 0 | - | | | | | |
| 2020 | 0 | - | | | | | |
| 2021 | 0 | - | | | | | |
| 2022 | 0 | - | | | | | |

¹ Source: PIRO SFD unpublished data.

Note: Federal permit and reporting requirements for large vessels in Guam's bottomfish fishery became effective on December 4, 2006 (71 FR 69496, December 1, 2006).

1.2.11.2.2 Spiny and Slipper Lobster Fishery

Table 37. Summary of federal logbook data for Guam lobster fisheries

| Year | No. of Federal Lobster | No. of Federal Lobster Permits Reporting | No. of Trips in Guam | Total Reported Logbook 'Catch (lb) | | | ted Logbook Discard (lb) |
|------|--------------------------------|--|----------------------------|------------------------------------|-------------------------------------|-----------------------------------|-------------------------------------|
| | Permits Issued ¹ | Catch EEZ | | Spiny lobster ECS ² | Slipper lobster ECS ² | Spiny lobster ECS ² | Slipper lobster ECS ² |
| 2004 | 0 | - | | | | | |
| 2005 | 0 | - | | | | | |
| 2006 | 2 | 0 | | | | | |
| 2007 | 2 | 0 | | | | | |
| 2008 | 7 | 0 | | | | | |

² On February 8, 2019, NMFS published a final rule (84 FR 2767) to reclassify some BMUS and all CREMUS in the Mariana Archipelago as ECS.

| Year | No. of Federal Lobster | No. of Federal Lobster Permits Reporting | No. of Trips in Guam | Total Report | | | rted Logbook Discard (lb) |
|------|--------------------------------|--|----------------------------|--------------------------------|-------------------------|-----------------------------------|------------------------------|
| | Permits Issued ¹ | Cotch | | Spiny lobster ECS ² | Slipper lobster ECS² | Spiny lobster ECS ² | Slipper lobster ECS² |
| 2009 | 0 | - | | | | | |
| 2010 | 0 | - | | | | | |
| 2011 | 0 | - | | | | | |
| 2012 | 0 | - | | | | | |
| 2013 | 0 | - | | | | | |
| 2014 | 0 | - | | | | | |
| 2015 | 0 | - | | | | | |
| 2016 | 1* | 0 | | | | | |
| 2017 | 0 | - | | | | | |
| 2018 | 1* | 0 | | | | | |
| 2019 | 0 | - | | | | | |
| 2020 | 0 | - | | | | | |
| 2021 | 0 | - | | | | | |
| 2022 | 0 | - | | | | | |

¹ Source: PIRO SFD unpublished data.

1.2.11.2.3 Deepwater Shrimp Fishery

Table 38. Summary of federal logbook data for Guam deepwater shrimp fisheries

| Year | No. of Federal Shrimp Permits Issued ¹ | No. of Federal Shrimp Permits Reporting Catch | No. of Trips in Guam EEZ | Total Reported Logbook Shrimp ECS ² Catch (lb) | Total Reported Logbook Shrimp ECS ² Release/Discard (lb) |
|------|---|---|--------------------------------|---|---|
| 2009 | 0 | - | | | |
| 2010 | 2 | 0 | | | |
| 2011 | 2 | 0 | | | |
| 2012 | 0 | - | | | |
| 2013 | 0 | - | | | |
| 2014 | 0 | - | | | |
| 2015 | 1 | 0 | | | |
| 2016 | 1 | 0 | | | |
| 2017 | 0 | - | | | |
| 2018 | 0 | - | | | |

² On February 8, 2019, NMFS published a final rule (84 FR 2767) to reclassify all CMUS in the Mariana Archipelago as ECS.

^{*} Permits apply to multiple areas and may include American Samoa, Guam, CNMI, and PRIA. Note: Federal permit and reporting requirements for CNMI lobster fisheries became effective on December 6, 2006 (71 FR 69496, December 1, 2006).

| Year | No. of Federal Shrimp Permits Issued ¹ | No. of Federal Shrimp Permits Reporting Catch | No. of Trips in Guam EEZ | Total Reported Logbook Shrimp ECS ² Catch (lb) | Total Reported Logbook Shrimp ECS ² Release/Discard (lb) |
|------|---|---|--------------------------------|---|---|
| 2019 | 0 | - | | | |
| 2020 | 0 | - | | | |
| 2021 | 0 | - | | | |
| 2022 | 0 | - | | | |

¹ Source: PIRO SFD unpublished data

Note: Federal permit and reporting requirements for deepwater shrimp fisheries became effective on June 29, 2009 (74 FR 25650, May 29, 2009).

1.2.12 Status Determination Criteria

1.2.12.1 Bottomfish Fishery

Overfishing criteria and control rules are specified and applied to individual species within the multi-species stock whenever possible. When this is not possible, they are based on an indicator species for the multi-species stock. It is important to recognize that individual species would be affected differently based on this type of control rule, and it is important that for any given species fishing, mortality does not currently exceed a level that would result in excessive depletion of that species. No indicator species are being used for the bottomfish multi-species stock complex. Instead, the control rules are applied to each stock complex as a whole.

The MSY control rule is used as the maximum fishing mortality threshold (MFMT). The MFMT and minimum stock size threshold (MSST) are specified based on recommendations in Restrepo et al. (1998) and both are dependent on the natural mortality rate (M; Table 39). The value of M used to determine the reference point values is not specified in this section. The latest estimate, published annually in the SAFE report, is used and the value is occasionally re-estimated using the best available information. The range of M among species within a stock complex is taken into consideration when estimating and choosing the M to be used for the purpose of computing the reference point values. In addition to the thresholds MFMT and MSST, a warning reference point, B_{FLAG}, is specified at some point above the MSST to provide a trigger for consideration of management action prior to B reaching the threshold.

Table 39. Overfishing threshold specifications for Guam BMUS

| MFMT | MSST | $\mathbf{B}_{	ext{FLAG}}$ | | | | |
|--|---------------------------------|-------------------------------|--|--|--|--|
| $F(B) = \frac{F_{MSY}B}{c B_{MSY}} \text{for } B \le c B_{MSY}$ | $c~\mathrm{B}_{	ext{	iny MSY}}$ | $\mathbf{B}_{	ext{	iny MSY}}$ | | | | |
| $F(B) = F_{MSY}$ for $B > c B_{MSY}$ | | | | | | |
| where $c = \max(1-M, 0.5)$ | | | | | | |

Standardized values of fishing effort (E) and CPUE can be used as proxies for F and B, respectively, so E_{MSY} , $CPUE_{MSY}$, and $CPUE_{FLAG}$ can be used as proxies for F_{MSY} , B_{MSY} , and B_{FLAG} , respectively.

² On February 8, 2019, NMFS published a final rule (84 FR 2767) to reclassify all CMUS in the Mariana Archipelago as ECS.

In cases where reliable estimates of CPUE_{MSY} and E_{MSY} are not available, they can be estimated from catch and effort times series, standardized for all identifiable biases. CPUEMSY would be calculated as half of a multi-year average reference CPUE, called CPUE_{REF}. The multi-year reference window would be objectively positioned in time to maximize the value of CPUE_{REF}. E_{MSY} would be calculated using the same approach or, following Restrepo et al. (1998), by setting E_{MSY} equal to E_{AVE}, where E_{AVE} represents the long-term average effort prior to declines in CPUE. When multiple estimates are available, the more precautionary one is used.

Since the MSY control rule specified here applies to multi-species stock complexes, it is important to ensure that no particular species within the complex has a mortality rate that leads to excessive depletion. In order to accomplish this, a secondary set of reference points is specified to evaluate stock status with respect to recruitment overfishing. A secondary "recruitment overfishing" control rule is specified to control fishing mortality with respect to that status. The rule applies only to those component stocks (species) for which adequate data are available. The ratio of a current spawning stock biomass proxy (SSBP_t) to a given reference level (SSBP_{REF}) is used to determine if individual stocks are experiencing recruitment overfishing. SSBP is CPUE scaled by percent mature fish in the catch. When the ratio SSBP_t/SSBP_{REF}, or the "SSBP ratio" (SSBPR) for any species drops below a certain limit (SSBPR_{MIN}), that species is considered to be recruitment overfished and management measures will be implemented to reduce fishing mortality on that species. The rule applies only when the SSBPR drops below the SSBPR_{MIN}, but it will continue to apply until the ratio achieves the "SSBP ratio recovery target" (SSBPR_{TARGET}), which is set at a level no less than SSBPR_{MIN}. These two reference points and their associated recruitment overfishing control rule, which prescribe a target fishing mortality rate (F_{RO-REBUILD}) as a function of the SSBPR, are specified as indicated in Table 40. Again, E_{MSY} is used as a proxy for F_{MSY}.

 $\mathbf{F}_{\text{RO-REBUILD}}$ **SSBPR**_{MIN} $SSBPR_{TARGET}$ for SSBPR ≤ 0.10 $F(SSBPR) = 0.2 F_{MSY}$ for $0.10 < SSBPR \le SSBPR_{MIN}$ 0.20 0.30

Table 40. Rebuilding control rules for Guam BMUS

1.2.12.2 Current Stock Status

 $F(SSBPR) = 0.5 F_{MSY}$ for $SSBPR_{MIN} < SSBPR \le SSBPR_{TARGET}$

1.2.12.2.1 Bottomfish

F(SSBPR) = 0

Biological and other fishery data are poor for all bottomfish species in the Mariana Archipelago. Generally, data are only available on commercial landings by species and CPUE for the multispecies complexes as a whole. At this time, it is not possible to partition these effort measures among the various BMUS. The most recent stock assessment (Langseth et al. 2019) for the Guam BMUS complex (comprised of 11 species of shallow and deep species of snapper, grouper, jacks, and emperors) was based on estimate of total catch, an abundance index derived from the nominal CPUE generated from the creel surveys. The assessments used a state-space Bayesian surplus production model within the modeling framework Just Another Bayesian Biomass Assessment (JABBA), which included biological information and fishery-dependent data through 2017. Determinations of overfishing and overfished status can then be made by comparing current biomass and harvest rates to MSY level reference points. To date, the Guam BMUS is in an overfished state but not undergoing overfishing.

Table 41. Stock assessment parameters for the Guam BMUS complex (from Langseth et al. 2019)

| Parameter | Parameter Value Notes | | Status |
|--------------------|-----------------------|-------------------------------|--------------------------|
| MSY | 42.1 (29.3-65.5) | Expressed in 1000 lb (with | |
| IVIS I | 42.1 (29.3-03.3) | 95% confidence interval) | |
| H ₂₀₁₇ | 0.11 | Expressed in percentage | |
| тт | 0.17 (0.071 – 0.382) | Expressed in percentage (with | |
| H_{CR} | 0.17 (0.071 – 0.362) | 95% confidence interval) | |
| H/H _{CR} | 0.81 | | No overfishing occurring |
| B ₂₀₁₇ | 143.0 | Expressed in 1000 lb | |
| D | 249 9 (107 1 626 9) | Expressed in 1000 lb (with | |
| B _{MSY} | 248.8 (107.1-636.8) | 95% confidence interval) | |
| B/B _{MSY} | 0.57 | | Overfished |

1.2.13 Overfishing Limit, Acceptable Biological Catch, and Annual Catch Limits

1.2.13.1 Brief Description of the ACL Process

The Council developed a tiered system of control rules to guide the specification of ACLs and Accountability Measures (AMs; WPRFMC 2011). The process starts with the use of the best scientific information available (BSIA) in the form of, but not limited to, stock assessments, published papers, reports, and/or available data. These data are categorized into the different tiers in the control rule ranging from Tier 1 (i.e., most information available, typically a stock assessment) to Tier 5 (i.e., catch-only information). The control rules are applied to the BSIA. Tiers 1 to 3 involve conducting a Risk of Overfishing Analysis (denoted by P*) to quantify the scientific uncertainties associated with the assessment to specify the Acceptable Biological Catch (ABC), lowering the MSY-based OFL to the ABC. A Social, Ecological, Economic, and Management (SEEM) Uncertainty Analysis is performed to quantify the uncertainties associated with the SEEM factors, and a buffer is used to lower the ABC to an ACL. For Tier 4, which is comprised of stocks with MSY estimates but no active fisheries, the control rule is 91 percent of MSY. For Tier 5, which has catch-only information, the control rule is a one-third reduction in the median catch depending on a qualitative evaluation of stock status via expert opinion. ACL specification can choose from a variety of methods including the above mentioned SEEM analysis or a percentage buffer (i.e., percent reduction from ABC based on expert opinion) or the use of an Annual Catch Target (ACT). Specifications are done on an annual basis, but the Council normally produces a multi-year specification.

The AM for Guam bottomfish fisheries is an overage adjustment. The next ACL is downward adjusted with the amount of overage from the previous ACL based on a three-year running average.

1.2.13.2 Current OFL, ABC, ACL, and Recent Catch

On February 18, 2022, NMFS the rebuilding plan for the Guam bottomfish stock complex that specified a 31,000 lb ACL (87 FR 9271). The catch shown in Table 42 takes the average of the most recent three years as recommended by the Council at its 160th meeting to avoid large fluctuations in catch due to high interannual variability in creel survey estimates.

Table 42. Guam 2022 ACL table with three-year average catch (lb)

| Fishery | MUS | OFL | ABC | ACL | Catch |
|------------|----------------------------------|-----|-----|--------|---------|
| Bottomfish | Bottomfish multi-species complex | - | - | 31,000 | 37,593* |

^{*} Value is preliminary and undergoing data validation.

1.2.14 Best Scientific Information Available

1.2.14.1 Bottomfish fishery

1.2.14.1.1 Stock Assessment Benchmark

The benchmark stock assessment for the Territory Bottomfish Management Unit Species complex was developed and finalized by Langseth et al. (2019). The assessments used a state-space Bayesian surplus production model within the modeling framework Just Another Bayesian Biomass Assessment (JABBA). Estimates of harvest rate (H), annual biomass (H), the harvest rate associated with overfishing as determined by the harvest control rule (H_{CR}), maximum sustainable yield (H_{CR}), and the biomass at maximum sustainable yield (H_{CR}) allowed for determination of stock status relative to reference points determining overfishing (H_{CR}) and overfished (H_{CR}) status. Stock projections were conducted for 2020-2025 for a range of hypothetical 6-year catches, and the corresponding risk of overfishing was calculated.

1.2.14.1.2 Stock Assessment Updates

Updates to the 2007 benchmark were done in 2012 (Brodziak et al. 2012) and 2015 (Yau et al. 2016). These included a three-year stock projection table used for selecting the level of risk the fishery will be managed under ACLs. Yau et al. (2016) is considered the BSIA for the Guam BMUS complex after undergoing a Western Pacific Stock Assessment Review (WPSAR) Tier 3 panel review (Franklin et al. 2015) prior to the Langseth et al. (2019) benchmark stock assessment. This was the basis for the P* and SEEM analyses that previously determined the risk levels to specify past ABCs and ACLs.

1.2.14.1.3 Other Information Available

Approximately every five years PIFSC administers a socioeconomic survey to small boat fishermen in Guam. This survey consists of about 60 questions regarding a variety of topics, including fishing experiences, market participation, vessels and gear, demographics and household income, and fishermen perspectives. The survey requests participants to identify which MUS they primarily targeted during the previous 12 months, by percentage of trips. Full reports of these surveys can be found at the <u>PIFSC Socioeconomics webpage</u>.

1.2.15 Harvest Capacity and Extent

The MSA defines the term "optimum," with respect to the yield from a fishery, as the amount of fish that:

- Will provide the greatest overall benefit to the Nation, particularly with respect to food production and recreational opportunities, and taking into account the protection of marine ecosystems.
- Is prescribed on the basis of the MSY from the fishery, as reduced by any relevant social, economic, or ecological factor.

• In the case of an overfished fishery, provides for rebuilding to a level consistent with producing the MSY in such a fishery [50 CFR §600.310(f)(1)(i)].

Optimum yield (OY) in the bottomfish fisheries is prescribed based on the MSY from the stock assessment and the best available scientific information. In the process of specifying ACLs, social, economic, and ecological factors were considered and the uncertainties around those factors defined the management uncertainty buffer between the ABC and ACL. OY for the bottomfish MUS complex is defined to be the level of harvest equal to the ACL consistent with the goals and objectives of the FEPs and used by the Council to manage the stock.

The Council recognizes that MSY and OY are long-term values whereas the ACLs are yearly snapshots based on the level of fishing mortality at MSY (F_{MSY}). There are situations when the long-term means around MSY are lower than ACLs especially if the stock is known to be productive or relatively pristine or lightly fished. A stock can have catch levels and catch rates exceeding that of MSY over the short-term to lower the biomass to a level around the estimated MSY and still not jeopardize the stock.

The harvest extent, in this case, is defined as the level of catch harvested in a fishing year relative to the ACL or OY. The harvest capacity is the level of catch remaining in the annual catch limit that can potentially be used for the TALFF.

Table 43 summarizes the harvest extent and harvest capacity information for Guam tracking the annual catch against the most recently implemented ACL (86 FR 24511, May 7, 2021).

Table 43. Guam ACL proportion of harvest capacity and extent in 2022

| Fishery | MUS | ACL | Catch (lb) | Harvest extent (%) | Harvest capacity (%) |
|------------|----------------------------------|--------|------------|--------------------|----------------------|
| Bottomfish | Bottomfish multi-species complex | 31,000 | 45,071 | 145.4 | 0.0 |

^{*} Value is preliminary and undergoing data validation.

1.2.16 Other Relevant Ocean-Uses and Fishery-Related Information

1.2.16.1 Territorial Marine Preserves

Guam has five locally managed MPAs: Achang Reef Flat in Merizo, Sasa Bay in Piti, Piti Bombholes in Piti, Tumon Bay in Tumon, and Pati Point in Yigo. A total of 11.8 percent of Guam's coastline is located within these MPAs.

1.2.16.2 Local Environmental Co-Variates

In early 2010, the U.S. military began exercises in an area south and southeast of Guam designated W-517. W-517 is a special use airspace (SUA) (approximately 14,000 nm²) that overlays deep open ocean approximately 50 miles south-southwest of Guam. Exercises in W-517 generally involve live fire and/or pyrotechnics When W-517 or other areas are in use, a notice to mariners (NTM) is issued, and vessels attempting to use the area are advised to be cautious of objects in the water and other small vessels. This discourages access to virtually all banks south of Guam, including Galvez, Santa Rosa, White Tuna, and other popular fishing areas. From 1982-2015, DAWR surveys recorded more than 2,930 trolling and bottom fishing trips to these southern banks, an average of more than 83 trips per year. Available data from 2022 indicate that NTMs were associated with 126 warning days for W-517. Additional information and data can be found in Table 68 in Section 2.8.5.

1.2.17 Administrative and Regulatory Actions

This summary describes management actions NMFS implemented for insular fisheries in Guam during calendar year 2022.

On February 18, 2022, NMFS published a final rule to implement Amendment 6 to the Fishery Ecosystem Plan for the Mariana Archipelago and the rebuilding plan for the Guam bottomfish stock complex (87 FR 9271). The rule included a 31,000 lb ACL. As an in-season AM, if NMFS projects that the fishery will reach the ACL in any fishing year, NMFS will close the fishery in Federal waters for the remainder of that year. As a post-season AM, if the total annual catch exceeds the ACL during a fishing year, NMFS will close the fishery in Federal waters until NMFS and the Guam government implement a coordinated management regime that allows the stock to rebuild. This action was necessary to rebuild the overfished stock consistent with the requirements of the Magnuson-Stevens Act and became effective March 21, 2022. NMFS and the Council will review the rebuilding plan every two years and modify it, as necessary, per section 304(e)(7) of the Magnuson-Stevens Act.

2 ECOSYSTEM CONSIDERATIONS

2.1 FISHER OBSERVATIONS

Hawaii fishermen Clay Tam and Roy Morioka started the fisher observations initiative in 2020 to add traditional and local ecological knowledge, and on-the-water observations to fishery-dependent data sources in the annual SAFE reports. Fisher observations from 2021 can be found in the pelagic and the respective archipelagic reports (WPRFMC 2022a; WPRFMC 2022b; WPRFMC 2022c; WPRFMC 2022d).

During 2022, the Council collected archipelagic fisher observations during quarterly advisory panel meetings for Guam and the CNMI. Input collected by fishers during these meetings was limited to Advisory Panel members. The Council also convened a meeting dedicated to Guam and CNMI fisher observations on February 7, 2023. This meeting included Guam and CNMI Advisory Panel members, but also included other individuals from their respective fishing communities. The full results from these fisher observation meetings is available as a PIFSC data report. The Marianas archipelagic fisher observations will begin with a summary quarterly advisory panel meetings from the 2022 calendar year, separated by island area, followed by a summary of 2022 archipelagic fisher observations data collected from the February 2023 meeting.

2.1.1 Information from Advisory Panel Meetings

2.1.1.1 CNMI – Saipan

During April to June, Talaya [throw net] fishers caught ti'ao [juvenile goatfish] on Saipan and small groups of ti'ao were also observed on Mañagaha along with Mañahak [juvenile rabbitfish]. An AP member reported crowding at boat ramps on Saipan, especially on weekends. They felt that they crowding underscores the need for another boat ramp for fishing. During the summer, an AP member reported that the atulai run was the best it's been in 10 years. But, spearfishing for Laggua [parrotfish] has become difficult. Divers need to go deeper to find them. AP members also reported that some shoreline fishing access has been cut off due to private or government property closures.

2.1.1.2 CNMI - Tinian

From April to June, warm water and an abundance of sharks made fishing difficult. Sharks were even hitting fishing vessels. Unleaded fuel prices on Tinian reached \$8.50 per gallon with Diesel at \$9-10 per gallon. Military activity continued to increase on Tinian, but the main issue affecting fishing was the fuel prices. Another AP member noted that fuel prices continue to increase while fish prices remain the same. They are still waiting on assistance from the Public Assistance Office to complete the floating marina. They held a groundbreaking ceremony for the new Tinian marina, which will be paved. During the summer, the good atulai run continued along with an ie'e run. Fishers noted that it has been a good season. AP members reported good fishing around debris off Tinian.

2.1.1.3 CNMI - Rota

During the spring a foreign purse seine vessel was observed off of Rota. Fuel prices decreased in the summer to around \$6.90 per gallon, but that price is still very high.

2.1.1.4 Guam

From January to March, AP members noted small runs of mañahak and ti'ao. Fuel prices for regular unleaded gasoline were \$6.39 per gallon, premium was \$6.79 per gallon, and Diesel increased to \$7.25 per gallon. From March to June, AP members reported few fish being caught. From August to September, many hook-and-line atulai fishers were out under the new moon. Although the atulai were abundant, heavy rains and hot sun prevented many of them from being caught. Strange weather with strong winds affected fishing trips.

2.1.2 Information from the Annual Summit

On February 7, 2023 from 6:00-8:00pm Chamorro Standard Time, the Council convened a fisher observations meeting with 5 advisory panel members from Guam and 6 from the CNMI, along with other 8 members of the fishing community (4 each from Guam and the CNMI). Hawai'i fishermen Clay Tam and Roy Morioka convened and facilitated the meeting and it was also attended remotely by Council staff, and one PIFSC staff member. Like the 2022 meeting that collected 2021 fisher observations, the focus of the meeting was to describe notable fishery events, changes in timing of fisheries events, issues the council should pay attention to, and consideration for causes behind any changes. Discussions were based upon a streamlined interview guide developed by Roy Morioka and Council staff member Zach Yamada. Although the interview guide was streamlined from the previous year, it did not substantially change participant responses. When necessary, attendees were asked follow up questions related to different social, economic, ecological, and management (SEEM) aspects of the fishery to facilitate their use in fisheries science and management. These four SEEM categories comprise a qualitative construct which have been used to complement the quantitative P* construct and process, and provide additional guidance when setting annual catch limits (Hospital et al. 2019).

The Guam and CNMI fisher observations meeting was not recorded, but PIFSC staff along with Council staff took detailed notes during the meeting and captured attendee quotes as close to verbatim as possible and captured all main ideas. Main ideas were categorized topically using the SEEM categories, then into additional sub-categories to provide further detail on fisher observations from Guam and CNMI fishers in 2022. Below, their observations of archipelagic fisheries are separated and described using the SEEM categories.

2.1.2.1 Social

In Guam, fishers described up to 30 new boats entering the fishery from the CNMI, crowding and thefts at boat ramps, and customary exchange of fish in the community.

In the CNMI, fishers noted a marina upgrade, and about 30 boats exiting the fishery, which were delivered to Guam. They also described ongoing military exercises that interfered with fishing activity and reduced fishing effort.

2.1.2.2 Economic

Guam fishers reported high fuel costs and lower fish prices due to excess supply of fish, although one fisher reported that markets for deep bottomfish species were more resilient.

Market conditions remain challenging in the CNMI, as has been noted in previous social research in fishing community profiles (Allen and Amesbury 2012; Ayers 2018), but fishers still

found some markets for their catch by making door-to-door sales. Fuel costs, normally high in the Marianas, reached as high as \$7.29/gallon in 2022.

2.1.2.3 Ecological (Biological and Physical/Oceanographic)

Most Guam fishers reported good fishing in 2022, both in terms of amount and size of fish. Guam fishers also reported larger sizes of bottomfish and ongoing shark depredation. Guam fishers noted stronger winds and weather, which inhibited fishing trips. Abnormal currents made bottomfishing difficult. Fishers also recorded cooler water temperatures throughout the year.

CNMI fishers reported larger numbers of ecosystem component species around coral reefs. They also described seeing larger numbers of sea turtles in lagoon areas and continued shark depredation. 2022 was a good year for atulai and fishers reported larger numbers of other forage items. Rougher water was reported due to prevailing weather and wind patterns, along with cooler water temperatures, and stronger currents. One fisher noted a lot of fishing debris in the water around the islands.

2.1.2.4 Management Uncertainty

There were no comments from fishers pertaining to management.

2.2 CORAL REEF ECOSYSTEM PARAMETERS

2.2.1 Regional Reef Fish Biomass and Habitat Condition

Description: 'Reef fish biomass' is mean biomass of reef fishes per unit area derived from visual survey data between 2010 and 2022. Hard coral cover is mean cover derived from visual estimates by divers of sites where reef fish surveys occurred. No surveys occurred in 2020 or 2021 due to COVID-19, but surveys were conducted in 2022.

Rationale: Reef fish biomass has been widely used as an indicator of relative ecosystem status and has repeatedly been shown to be sensitive to changes in fishing pressure, habitat quality, and oceanographic regime. Hard coral cover is an indicator of relative status of the organisms that build coral reef habitat and has been shown to be sensitive to changes in oceanographic regime, and a range of direct and indirect anthropogenic impacts. Most fundamentally, cover of hard corals has been increasingly impacted by temperature stress as a result of global heating.

<u>Data Category</u>: Fishery-independent

Timeframe: Triennial

<u>Jurisdiction</u>: American Samoa, Guam, the Commonwealth of the Northern Mariana Islands (CNMI), Main Hawaiian Islands (MHI), Northwestern Hawaiian Islands (NWHI), and Pacific Remote Island Areas (PRIA)

Spatial Scale: Regional

<u>Data Source</u>: Data used to generate cover and biomass estimates come from visual surveys conducted by the National Marine Fisheries Service (NMFS) Pacific Island Fisheries Science Center (PIFSC) Ecosystem Sciences Division (ESD) and their partners as part of the Coral Reef Conservation Program's (CRCP) National Coral Reef Monitoring Program (<u>NCRMP</u>). Survey methods are described in detail in Ayotte et al. (2015). In brief, they involve teams of divers conducting stationary point count cylinder (SPC) surveys within a target domain of < 30 meter hard-bottom habitat at each island, stratified by depth zone and, for larger islands, by section of coastline. For consistency among islands, only data from forereef habitats are used. At each SPC, divers record the number, size, and species of all fishes within or passing through paired 15 meter-diameter cylinders over the course of a standard count procedure.

Fish sizes and abundance are converted to biomass using standard length-to-weight conversion parameters, taken largely from FishBase and converted to biomass per unit area by dividing by the area sampled per survey. Site-level data were pooled into island-scale values by first calculating mean and variance within strata, and then calculating weighted island-scale mean and variance using the formulas given in Smith et al. (2011) with strata weighted by their respective sizes.

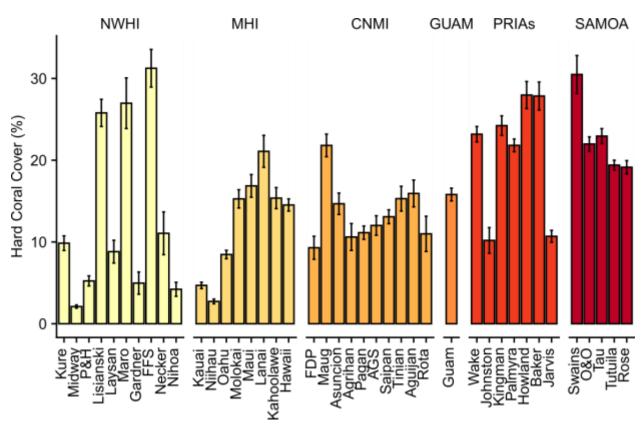


Figure 1. Mean coral cover (% \pm SEM) per U.S. Pacific Island averaged over the years 2010-2022 by latitude

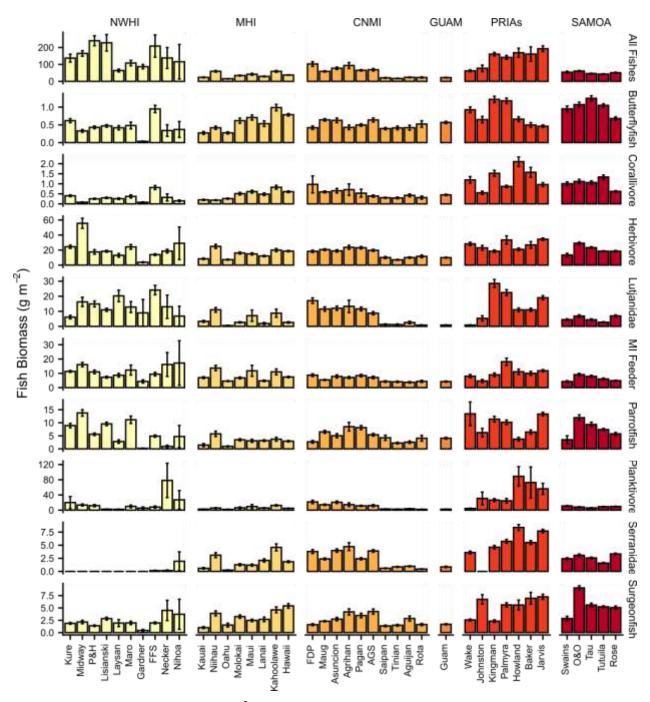


Figure 2. Mean fish biomass (g/m² ± SEM) of functional, taxonomic, and trophic groups by U.S. Pacific reef area from the years 2010-2022 by latitude. The group 'Serranidae' excludes planktivorous members of that family (i.e., anthias, which can by hyper-abundant in some regions). Similarly, the bumphead parrotfish, *Bolbometopon muricatum*, has been excluded from the corallivore group. The group 'MI Feeder' consists of fishes that primarily feed on mobile invertebrates, 'Butterflyfish' are non-planktivorous butterflyfish species, and 'Surgeonfish' are mid-large target surgeonfish species

2.2.2 CNMI Reef Fish Biomass and Habitat Condition

Description: 'Reef fish biomass' is mean biomass of reef fishes per unit area derived from visual survey data between 2010 and 2022. Hard Coral cover is mean cover derived from visual estimates by divers of sites where reef fish surveys occurred. No surveys occurred in 2020 or 2021 due to COVID-19, but surveys were conducted in 2022.

Rationale: Reef fish biomass has been widely used as an indicator of relative ecosystem status and has repeatedly been shown to be sensitive to changes in fishing pressure, habitat quality, and oceanographic regime. Hard coral cover is an indicator of relative status of the organisms that build coral reef habitat and has been shown to be sensitive to changes in oceanographic regime, and a range of direct and indirect anthropogenic impacts. Most fundamentally, cover of hard corals has been increasingly impacted by temperature stress as a result of global heating.

<u>Data Category</u>: Fishery-independent

<u>Timeframe</u>: Triennial <u>Jurisdiction</u>: CNMI **Spatial Scale**: Island

<u>Data Source</u>: Data used to generate biomass and cover estimates comes from visual surveys conducted by NMFS PIFSC ESD and partners, as part of the Pacific NCRMP Survey methods and sampling design, and methods to generate reef fish biomass are described in Section 2.2.1.

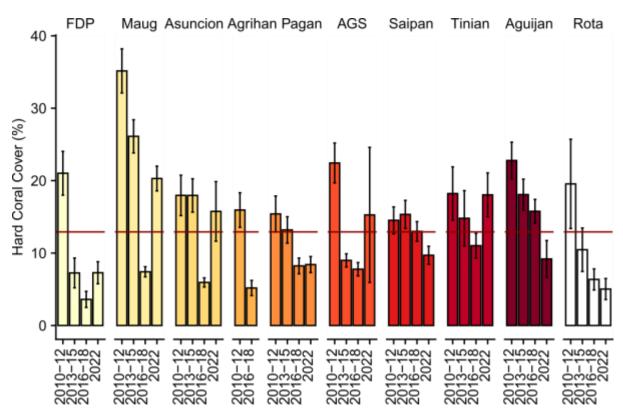


Figure 3. Mean coral cover ($\% \pm SEM$) per island over the years 2010-2022 by latitude with CNMI mean estimates plotted for reference (horizontal red line)

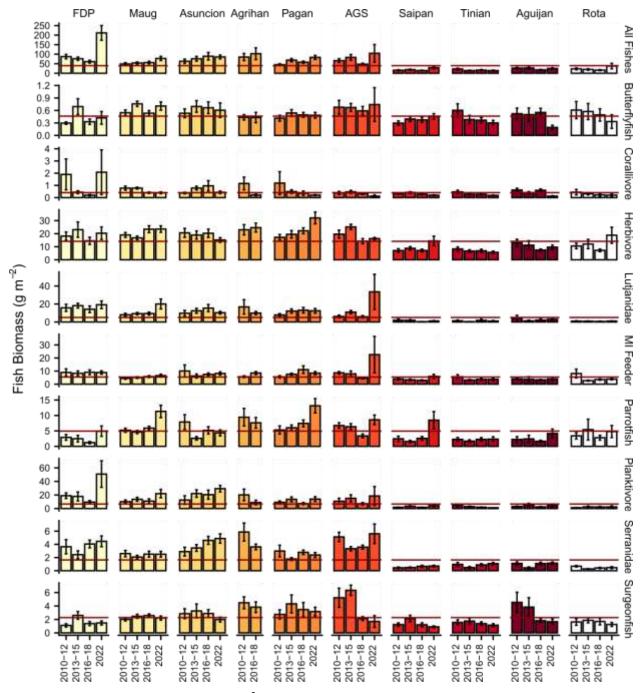


Figure 4. Mean fish biomass (g/m² ± SEM) of CNMI functional, taxonomic, and trophic groups over the years 2010-2022 by island. The group 'Serranidae' excludes planktivorous members of that family (i.e., anthias, which can by hyper-abundant in some regions). Similarly, the bumphead parrotfish, *Bolbometopon muricatum*, has been excluded from the corallivore group. The group 'MI Feeder' consists of fishes that primarily feed on mobile invertebrates, 'Butterflyfish' are non-planktivorous butterflyfish species, and 'Surgeonfish' are mid-large target surgeonfish species. Red horizontal lines are the region-wide mean estimates for reference

2.2.3 Guam Reef Fish Biomass and Habitat Condition

<u>Description</u>: 'Reef fish biomass' is mean biomass of reef fishes per unit area derived from visual survey data between 2010 and 2022. Hard Coral cover is mean cover derived from visual estimates by divers of sites where reef fish surveys occurred.

Rationale: Reef fish biomass has been widely used as an indicator of relative ecosystem status and has repeatedly been shown to be sensitive to changes in fishing pressure, habitat quality, and oceanographic regime. Hard coral cover is an indicator of relative status of the organisms that build coral reef habitat and has been shown to be sensitive to changes in oceanographic regime, and a range of direct and indirect anthropogenic impacts. Most fundamentally, cover of hard corals has been increasingly impacted by temperature stress as a result of global heating.

<u>Data Category</u>: Fishery-independent

<u>Timeframe</u>: Triennial <u>Jurisdiction</u>: Guam <u>Spatial Scale</u>: Island

<u>Data Source</u>: Data used to generate biomass and cover estimates comes from visual surveys conducted by NMFS PIFSC ESD and partners, as part of the Pacific NCRMP Survey methods and sampling design, and methods to generate reef fish biomass are described in Section 2.2.1.

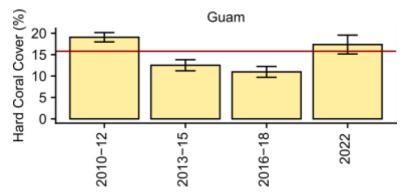


Figure 5. Mean coral cover ($\% \pm SEM$) over the years 2010-2022 by latitude with mean for the entire time period plotted for reference (horizontal red line)

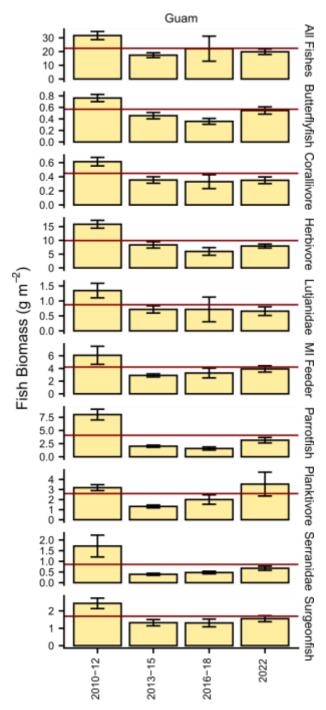


Figure 6. Mean fish biomass (g/m² ± SEM) of Guam functional, taxonomic, and trophic groups over the years 2010-2020. The group 'Serranidae' excludes planktivorous members of that family (i.e., anthias, which can by hyper-abundant in some regions). Similarly, the bumphead parrotfish, *Bolbometopon muricatum*, has been excluded from the corallivore group. The group 'MI Feeder' consists of fishes that primarily feed on mobile invertebrates, 'Butterflyfish' are non-planktivorous butterflyfish species, and 'Surgeonfish' are mid-large target surgeonfish species. Red horizontal lines are the Guam mean estimates for reference

2.3 LIFE HISTORY AND LENGTH DERIVED PARAMETERS

The annual stock assessment and fishery evaluation (SAFE) report will serve as the repository of available life history information for the Western Pacific region. Life history data particularly age, growth, reproduction, and mortality information inform stock assessments on fish productivity and population dynamics. Some assessments, particularly for data poor stocks, utilize information from other areas that introduces biases and increase uncertainties in the population estimates. An archipelago specific life history parameter ensures accuracy in the input parameters used in the assessment.

The NMFS PIFSC Bio-Sampling Program allows for the collection of life history samples like otoliths and gonads from priority species in the bottomfish and coral reef fisheries. A significant number of samples are also collected during research cruises. These life history samples, once processed and examined, will contribute to the body of scientific information for the two datapoor fisheries in the region (coral reef fish and bottomfish). The life history information available from the region will be monitored by the Fishery Ecosystem Plan Team and will be tracked through this section of the report.

This section will be divided into two fisheries: 1) prioritized coral reef ecosystem component species, and 2) management unit species (MUS). The prioritized coral reef species list was developed by the CNMI Department of Fish and Wildlife (DFW) and the Guam Division of Aquatic and Wildlife Resources (DAWR) in 2019. The MUS are the species that are listed in the federal ecosystem plan and are managed on a federal level. Within each fishery, the available life history information will be described under the age, growth, and reproductive maturity section. The section labelled fish length derived parameters summarizes available information derived from sampling the fish catch or the markets. Length-weight conversion coefficients provide areaspecific values to convert length from fishery-dependent and fishery-independent data collection to weight or biomass.

2.3.1 CNMI Coral Reef Ecosystem Components Life History

2.3.1.1 Age, Growth, and Reproductive Maturity

Description: Age determination is based on counts of yearly growth marks (annuli) and/or daily growth increments (DGIs) internally visible within transversely cut, thin sections of sagittal otoliths. Validated age determination is based on several methods including an environmental signal (bomb radiocarbon 14 C) produced during previous atmospheric thermonuclear testing in the Pacific and incorporated into the core regions of sagittal otolith and other aragonite-based calcified structures such as hermatypic corals. This technique relies on developing a regionally based aged coral core reference series for which the rise, peak, and decline of 14 C values is available over the known age series of the coral core. Estimates of fish age are determined by projecting the 14 C otolith core values back in time from its capture date to where it intersects with the known age 14 C coral reference series. Fish growth is estimated by fitting the length-atage data to a von Bertalanffy growth function (VBGF). This function typically uses three coefficients (L_{∞} , k, and t_0), which together characterize the shape of the length-at-age growth relationship.

Length-at-reproductive maturity is based on the histological analyses of small tissue samples of gonad material that are typically collected along with otoliths when a fish is processed for life history studies. The gonad tissue sample is preserved, cut into five-micron sections, stained, and

sealed onto a glass slide for subsequent examination. Based on standard cell structure features and developmental stages within ovaries and testes, the gender, developmental stage, and maturity status (immature or mature) is determined via microscopic evaluation. The percent of mature samples for a given length interval are assembled for each sex and these data are fitted to a three- or four-parameter logistic function to determine the best fit of these data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish have achieved reproductive maturity (L_{50}). For species that undergo sex reversal (primarily female to male in the tropical Pacific region) - such as groupers and deeperwater emperors among the bottomfishes, and for parrotfish, shallow-water emperors, and wrasses among the coral reef fishes - standard histological criteria are used to determine gender and reproductive developmental stages that indicate the transitioning or completed transition from one sex to another. These data are similarly analyzed using a three or four-parameter logistic function to determine the best fit of the data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish of a particular species have or are undergoing sex reversal ($L\Delta_{50}$).

Age at 50% maturity (A_{50}) and age at 50% sex reversal ($A\Delta_{50}$) is typically derived by referencing the VBGF for that species and using the corresponding L_{50} and $L\Delta_{50}$ values to obtain the corresponding age value from this growth function. In studies where both age & growth and reproductive maturity are concurrently determined, estimates of A_{50} and $A\Delta_{50}$ are derived directly by fitting the percent of mature samples for each age (one-year) interval to a three- or four-parameter logistic function using statistical analyses. The mid-point of this fitted logistic function provides a direct estimate of the age at which 50% of fish of a particular species have achieved reproductive maturity (A_{50}) and sex reversal ($A\Delta_{50}$).

Category: Biological

Timeframe: N/A

Jurisdiction: CNMI

Spatial Scale: Archipelagic

<u>Data Source</u>: Sources of data are directly derived from research cruises sampling and market samples collected by the CNMI contracted bio-sampling team which samples the catch of fishermen and local fish vendors. Laboratory analyses and data generated from these analyses reside with the PIFSC Life History Program (LHP). Refer to the "Reference" column in Table 44 for specific details on data sources by species.

Parameter Definitions:

 T_{max} (maximum age) – The maximum observed age revealed from an otolith-based age determination study. T_{max} values can be derived from ages determined by annuli counts of sagittal otolith sections and/or bomb radiocarbon (14 C) analysis of otolith core material. Units are years.

 L_{∞} (asymptotic length) – One of three coefficients of the VBGF that measures the mean maximum length at which the growth curve plateaus and no longer increases in length with increasing age. This coefficient reflects the estimated mean maximum length and not the observed maximum length. Units are centimeters.

k (growth coefficient) – One of three coefficients of the VBGF that measures the shape and steepness by which the initial portion of the growth function approaches its mean maximum length (L_{∞}) .

 t_{θ} (hypothetical age at length zero) – One of three coefficients of the VBGF whose measure is highly influenced by the other two VBGF coefficients (k and L_{∞}) and typically assumes a negative value when specimens representing early growth phases) are not available for age determination. This parameter can be fixed at 0. Units are years.

M (natural mortality) – This is a measure of the mortality rate for a fish stock and is considered to be directly related to stock productivity (i.e., high M indicates high productivity and low M indicates low stock productivity). M can be derived through use of various equations that link M to T_{max} and the VBGF coefficients (k and L_{∞}) or by calculating the value of the slope from a regression fit to a declining catch curve (regression of the natural logarithm of abundance versus age class) derived from fishing an unfished or lightly fished population.

 A_{50} (age at 50% maturity) – Age at which 50% of the sampled stock under study has attained reproductive maturity. This parameter is best determined based on studies that concurrently determine both age (otolith-based age data) and reproductive maturity status (logistic function fitted to percent mature by age class with maturity determined via microscopic analyses of gonad histology preparations). A more approximate means of estimating A_{50} is to use an existing L_{50} estimate to find the corresponding age (A_{50}) from an existing VBGF curve. Units are years.

 $A\Delta_{50}$ (age of sex switching) – Age at which 50% of the immature and adult females of the sampled stock under study is undergoing or has attained sex reversal. This parameter is best determined based on studies that concurrently determines both age (otolith-based age data) and reproductive sex reversal status (logistic function fitted to percent sex reversal by age class with sex reversal determined via microscopic analyses of gonad histology preparations). A more approximate means of estimating $A\Delta_{50}$ is to use an existing $L\Delta_{50}$ estimate to find the corresponding age $(A\Delta_{50})$ from the VBGF curve. Units are years.

 L_{50} (length at which 50% of a fish population are capable of spawning) – Length at which 50% of the females of a sampled stock under study has attained reproductive maturity; this is the length associated with A_{50} estimates. This parameter is derived using a logistic function to fit the percent mature data by length class with maturity status best determined via microscopic analyses of gonad histology preparations. L_{50} information is typically more available than A_{50} since L_{50} estimates do not require knowledge of age and growth. Units are centimeters.

 $L\Delta_{50}$ (length of sex switching) – Length at which 50% of the immature and adult females of the sampled stock under study is undergoing or has attained sex reversal; this is the length associated with $A\Delta_{50}$ estimates. This parameter is derived using a logistic function to fit the percent sex reversal data by length class with sex reversal status best determined via microscopic analyses of gonad histology preparations. $L\Delta_{50}$ information is typically more available than $A\Delta_{50}$ since $L\Delta_{50}$ estimates do not require knowledge of age and growth. Units are centimeters.

<u>Rationale</u>: These nine life history parameters provide basic biological information at the species level to evaluate the productivity of a stock - an indication of the capacity of a stock to recover once it has been depleted. These parameters are also used as direct inputs into stock assessments. Currently, the assessment of coral reef fish resources in CNMI is data limited. Knowledge of these life history parameters support current efforts to characterize the resilience of these

resources and provide important biological inputs for future stock assessment efforts and enhance our understanding of the species' likely role and status as a component of the overall ecosystem. Furthermore, knowledge of life histories across species at the taxonomic level of families or among different species that are ecologically or functionally similar can provide important information on the diversity of life histories and the extent to which species can be grouped (based on similar life histories) for future multi-species assessments.

Table 44. Available age, growth, reproductive maturity, and natural mortality information for prioritized coral reef ecosystem component species in CNMI

| C | | A | Age, growth | , and repro | ductive n | naturity pa | rameters | | | Reference |
|---------------------------------|--------------------------------------|--|--|--|-------------------|--|--|--|----------------|-----------------------|
| Species | T_{max} | L_{∞} | k | t_0 | M | A_{50} | $A\Delta_{50}$ | L_{50} | $L\Delta_{50}$ | Keierence |
| Acanthurus lineatus | | | | | | | | | | |
| Lethrinus harak | f=9 ^d m=9 ^d | f=37.2 ^d m=27.3 ^d | f=0.14 ^d m=0.38 ^d | f=-2.92 ^d m=-1.11 ^d | | f=2.6 ^d m=2.4 ^d | f=0.43 ^d m=0.44 ^d | f=19.6 ^d m=18.7 ^d | | Trianni (2016) |
| Mulloidichthys flavolineatus | f=5° M=4° | f=25.55° m=21.80° | f=1.24 ^c m=1.69 ^c | | | | | f=15.8 ^c m=16.1 ^c | | Reed et al. (2020) |
| Naso lituratus | | | | | | | | | NA | |
| Naso unicornis | | | | | | | | 238 ^b | NA | |
| Scarus rubroviolaceus | | | | | | | | | | |
| Scarus ghobban | | | | | | | | | | |
| Siganus argenteus | 7 ^d | 274 ^d | 0.9 ^d | -0.3 ^d | 0.56 ^d | 1.3 ^d | NA | 218 ^d | NA | Taylor et. al. (2016) |

^a signifies estimate pending further evaluation in an initiated and ongoing study.

Parameter estimates are for females unless otherwise noted (f=females, m=males). Parameters T_{max} , t_0 , A_{50} , and $A\Delta_{50}$ are in units of years; L_{∞} , L_{50} , and $L\Delta_{50}$ are in units of mm fork length (FL); k is in units of year⁻¹; X=parameter estimate too preliminary or Y=published age and growth parameter estimates based on DGI numerical integration technique and likely to be inaccurate; NA=not applicable. Superscript letters indicate status of parameter estimate (see footnotes below table). Published or in press publications (d) are denoted in the "Reference" column.

2.3.1.2 Fish Length Derived Parameters

Description: The NMFS Commercial Fishery Bio-sampling Program started in 2010. This program has two components: first is the Field/Market Sampling Program, and the second is the Lab Sampling Program, details of which are described in a separate section of this report. The goals of the Field/Market Sampling Program are:

- Broad scale look at commercial landings (by fisher/trip, gear, and area fished);
- Length and weight frequencies of whole commercial landings per fisher-trip (with an effort to also sample landings not sold commercially);
- Accurate species identification;
- Develop accurate local length-weight curves.

In CNMI, the Bio-sampling Program was focused on the commercial coral reef spear fishery with occasional sampling of the bottomfish fishery occurring locally and less frequently at the

^b signifies a preliminary estimate taken from ongoing analyses.

^c signifies an estimate documented in an unpublished report or draft manuscript.

^d signifies an estimate documented in a finalized report or published journal article (including in press).

northern islands. However, in 2020 the Program switched focus to the MUS. Sampling is conducted in partnership with the fish vendors and fishermen. The Market Sampling information includes (but not limited to): 1) fish length; 2) fish weight; 3) species identification; and 4) basic effort information. Specific for CNMI, the program collects Daily Vendor Logs for reef fish that includes basic catch and effort information.

Category: Biological

<u>Timeframe</u>: N/A **Jurisdiction:** CNMI

Spatial Scale: Archipelagic

Data Source: NMFS Bio-sampling Program

Parameter definitions:

n – sample size is the total number of fish sampled for length for each species recorded in the Bio-Sampling Program database.

 L_{max} – maximum fish length is the largest individual per species recorded in the Bio-Sampling Program database from the commercial spear fishery. This value is derived from measuring the length of individual samples for species occurring in the spear fishery. Units are centimeters.

 N_{L-W} – sample size for L-W regression is the number of samples used to generate the a and b coefficients.

a and b – length-weight coefficients are the coefficients derived from the regression line fitted to all length and weight measured by species in the commercial spear fishery. These values are used to convert length information to weight. Values are influenced by the life history characteristics of the species, geographic location, population status, and nature of the fisheries from which the species are harvested.

Rationale: Length derived information is an important component of fisheries monitoring and data poor stock assessment approaches. Maximum length (L_{max}) is used to derive missing species- and location-specific life history information (Nadon et al. 2015; Nadon and Ault 2016; Nadon 2019). The length-weight coefficients (a and b values) are used to convert length to weight for fishery-dependent and fishery-independent data collection where length is typically recorded but weight is the factor being used for management. This section of the report presents the best available information for the length derived variables for the CNMI coral reef ecosystem component fisheries.

Table 45. Available length derived information for prioritized coral reef ecosystem component species in CNMI

| Crasica | I | Length (| derived | parameter | S | Defenence |
|------------------------------|--------|-----------|-----------|-----------|-------|------------------------|
| Species | n | L_{max} | N_{L-W} | а | b | Reference |
| Acanthurus lineatus | 20,228 | 23.5 | 4927 | 0.03882 | 2.868 | Matthews et al. (2019) |
| Lethrinus harak | 2,697 | | | | | |
| Mulloidichthys flavolineatus | 12,516 | 31.4 | 2798 | 0.0138 | 3.05 | Matthews et al. (2019) |
| Naso lituratus | 28,507 | 30.1 | 3868 | 0.0163 | 3.103 | Matthews et al. (2019) |

| Charles | I | Length | derived | parameter | s | Reference | |
|-----------------------------|--------|-----------|-----------|-----------|-------|------------------------|--|
| Species | n | L_{max} | N_{L-W} | a | b | Keierence | |
| Naso unicornis | 12,481 | 53.6 | 4448 | 0.0269 | 2.908 | Matthews et al. (2019) | |
| Scarus ghobban ¹ | 7,612 | 38.1 | 1644 | 0.0129 | 3.12 | Matthews et al. (2019) | |
| Scarus rubroviolaceus | 4,032 | 52.6 | 1830 | 0.0089 | 3.24 | Matthews et al. (2019) | |
| Siganus argenteus | 14,614 | 34.1 | 3961 | 0.0129 | 3.112 | Matthews et al. (2019) | |

¹ Scarus ghobban did not have data to cover 30% of the total length range.

2.3.2 CNMI Management Unit Species Life History

2.3.2.1 Age, Growth, and Reproductive Maturity

<u>Description</u>: Age determination is based on counts of yearly growth marks (annuli) and/or DGIs internally visible within transversely cut, thin sections of sagittal otoliths. Validated age determination is based on several methods including an environmental signal (bomb radiocarbon 14 C) produced during previous atmospheric thermonuclear testing in the Pacific and incorporated into the core regions of sagittal otolith and other aragonite-based calcified structures such as hermatypic corals. This technique relies on developing a regionally based aged coral core reference series for which the rise, peak, and decline of 14 C values is available over the known age series of the coral core. Estimates of fish age are determined by projecting the 14 C otolith core values back in time from its capture date to where it intersects with the known age 14 C coral reference series. Fish growth is estimated by fitting the length-at-age data to a VBGF. This function typically uses three coefficients (L_{∞} , k, and t_0), which together characterize the shape of the length-at-age growth relationship.

Length-at-reproductive maturity is based on the histological analyses of small tissue samples of gonad material that are typically collected along with otoliths when a fish is processed for life history studies. The gonad tissue sample is preserved, cut into five-micron sections, stained, and sealed onto a glass slide for subsequent examination. Based on standard cell structure features and developmental stages within ovaries and testes, the gender, developmental stage, and maturity status (immature or mature) is determined via microscopic evaluation. The percent of mature samples for a given length interval are assembled for each sex, and these data are fitted to a three- or four-parameter logistic function to determine the best fit for the data based on statistical analyses. The mid-point of the fitted function provides an estimate of the length at which 50% of fish have achieved reproductive maturity (L_{50}). For species that undergo sex reversal (primarily female to male in the tropical Pacific region), such as groupers and deeperwater emperors among the bottomfishes, and for parrotfish, shallow-water emperors, and wrasses among the coral reef fishes, standard histological criteria are used to determine gender and reproductive developmental stages that indicate the transitioning or completed transition from one sex to another. These data are similarly analyzed using a three- or four-parameter logistic function to determine the best fit of the data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish of a particular species have or are undergoing sex reversal ($L\Delta_{50}$).

Age at 50% maturity (A_{50}) and age at 50% sex reversal ($A\Delta_{50}$) is typically derived by referencing the VBGF for that species and using the corresponding L_{50} and $L\Delta_{50}$ values to obtain the corresponding age value from this growth function. In studies where both age and growth and reproductive maturity are concurrently determined, estimates of A_{50} and $A\Delta_{50}$ are derived directly by fitting the percent of mature samples for each age (i.e., one-year) interval to a three- or four-

parameter logistic function using statistical analyses. The mid-point of this fitted logistic function provides a direct estimate of the age at which 50% of fish of a species have achieved reproductive maturity (A_{50}) and sex reversal ($A\Delta_{50}$).

Category: Biological

<u>Timeframe</u>: N/A **Jurisdiction**: CNMI

Spatial Scale: Archipelagic

<u>Data Source</u>: Sources of data are directly derived from research cruises sampling and market samples collected by the CNMI contracted bio-sampling team which samples the catch of fishermen and local fish vendors. Laboratory analyses and data generated from these analyses reside with the PIFSC LHP. Refer to the "Reference" column in Table 46 for specific details on data sources by species.

Parameter Definitions: Identical to Section 2.3.2.1

Rationale: These nine life-history parameters provide basic biological information at the species level to evaluate the productivity of a stock - an indication of the capacity of a stock to recover once it has been depleted. Currently, the assessment of coral reef fish resources in CNMI is data limited. Knowledge of these life-history parameters support current efforts to characterize the resilience of these resources, provide important biological inputs for future stock assessment efforts, and enhance our understanding of the species' likely role and status as a component of the overall ecosystem. Furthermore, knowledge of life histories across species at the taxonomic level of families or among different species that are ecologically or functionally similar can provide important information on the diversity of life histories and the extent to which species can be grouped (based on similar life histories) for future multi-species assessments.

Table 46. Available age, growth, reproductive maturity, and natural mortality information for MUS in CNMI

| | | Age | e, growtł | n, and re | producti | ve maturi | ty par | ameters | | |
|---|-----------------|-------------------|-------------------|--------------------|-------------------|--|---------------|--|-------------------|------------------------|
| Species | Tmax | L_{∞} | k | t_0 | M | A_{50} | <i>Α</i> Δ 50 | L_{50} | $L\Delta_{50}$ | Reference |
| Aphareus rutilans | | | | | | | NA | | NA | |
| Caranx ignobilis | | | | | | | | | | |
| Caranx lugubris | | | | | | | | | | |
| Etelis carbunculus ¹ | | | | | | | NA | | NA | |
| Etelis coruscans | | | | | | | NA | | NA | |
| Lethrinus rubrioperculatus | 8 ^d | 31.5 ^d | 0.80^{d} | -0.52 ^d | | | | 23.2 ^d | 29.0 ^d | Trianni (2011) |
| Lutjanus kasmira | | | | | | | NA | | NA | |
| Pristipomoides auricilla ² | 18 ^d | 32.5 ^d | 0.60^{d} | | 0.18 ^d | | NA | | NA | O'Malley et al. (2019) |
| Pristipomoides filamentosus ² | 31° | 54.6° | 0.19 ^c | | | f=5.0 ^d m=2.8 ^d | NA | f=41.2 ^d m=27.6 ^d | NA | Villagomez (2019) |
| Pristipomoides flavipinnis | | | | | | | NA | | NA | |
| Pristipomoides sieboldii | | | | | | | NA | | NA | |
| Pristipomoides | Xa | Xa | Xa | Xa | | | NA | | NA | LHP |

| Age, growth, and reproductive maturity parameters | | | | | | | | | | | |
|---|------|--------------|---|----------------|---|-----|---------------|-----|--------------|-----------|--|
| Species | Tmax | L_{∞} | k | t ₀ | М | A50 | <i>Α</i> Δ 50 | L50 | $L\Delta 50$ | Reference | |
| zonatus | | | | | | | | | | (in prep) | |
| Variola louti | | | | | | | | | | | |

¹ *E. carbunculus* is now known to be comprised of two distinct, non-interbreeding lineages (Andrews et al. 2016). Both species occur in the Mariana Archipelago and are likely both captured by fishermen but reported as one species.

Parameter estimates are for females unless otherwise noted (f=females, m=males). Parameters T_{max} , t_0 , A_{50} , and $A\Delta_{50}$ are in units of years; L_{∞} , L_{50} , and $L\Delta_{50}$ are in units of mm FL; k is in units of year⁻¹; X=parameter estimate too preliminary or Y=published age and growth parameter estimates based on DGI numerical integration technique and likely to be inaccurate; NA=not applicable. Superscript letters indicate status of parameter estimate (see footnotes below table). Published or in press publications (^d) are denoted in the "Reference" column.

2.3.2.2 Fish Length Derived Parameters

<u>Description</u>: The NMFS Commercial Fishery Bio-sampling Program started in 2010. This program has two components: first is the Field/Market Sampling Program and the second is the Lab Sampling Program, details of which are described in a separate section of this report. The goals of the Field/Market Sampling Program are:

- Broad scale look at commercial landings (by fisher/trip, gear, and area fished);
- Length and weight frequencies of whole commercial landings per fisher-trip (with an effort to also sample landings not sold commercially);
- Accurate species identification;
- Develop accurate local length-weight curves.

In CNMI, the Bio-sampling Program was focused on the commercial coral reef spear fishery with occasional sampling of the bottomfish fishery occurring locally and less frequently at the northern islands. However, in 2020 the Program switched focus to the MUS. Sampling is conducted in partnership with the fish vendors and fishermen. The Market Sampling information includes (but not limited to): 1) fish length; 2) fish weight; 3) species identification; and 4) basic effort information. Specific for CNMI, the program collects Daily Vendor Logs for reef fish that includes basic catch and effort information.

<u>Category</u>: Biological <u>Timeframe</u>: N/A <u>Jurisdiction</u>: CNMI

Spatial Scale: Island

Data Source: NMFS Bio-sampling Program

² Estimates are for the southern portion of the Mariana Archipelago.

^a signifies estimate pending further evaluation in an initiated and ongoing study.

^b signifies a preliminary estimate taken from ongoing analyses.

^c signifies an estimate documented in an unpublished report or draft manuscript.

d signifies an estimate documented in a finalized report or published journal article (including in press).

Parameter Definitions: Identical to Section 2.3.1.2

Rationale: Length derived information is an important component of fisheries monitoring and data poor stock assessment approaches. Maximum length (L_{max}) is used to derive missing species- and location-specific life history information (Nadon et al. 2015; Nadon and Ault 2016; Nadon 2019). The length-weight coefficients (a and b values) are used to convert length to weight for fishery-dependent and fishery-independent data collection where length is typically recorded but weight is the factor being used for management. This section of the report presents the best available information for the length derived variables for the CNMI MUS fisheries.

Table 47. Available length derived information for MUS species in CNMI

| Species | | Length | derived p | arameters | } | Reference |
|------------------------------------|------|-----------|-----------|-----------|--------|--|
| Species | n | L_{max} | N_{L-W} | a | b | |
| Aphareus rutilans | 120 | | | | | |
| Caranx ignobilis | 6 | | | | | |
| Caranx lugubris | 132 | 82.5 | 130 | 0.0313 | 2.87 | Matthews et al. (2019) |
| Etelis carbunculus ¹ | 746 | 53.5 | 685 | 0.0150 | 3.0430 | 2010-2015 CNMI Bio- Sampling Database |
| Etelis coruscans | 377 | 96.4 | 325 | 0.0716 | 2.6147 | 2010-2015 CNMI Bio- Sampling Database |
| Lethrinus rubrioperculatus | 1438 | 38.0 | 1353 | 0.0185 | 2.9897 | 2010-2015 CNMI Bio- Sampling Database |
| Lutjanus kasmira | 422 | 32.5 | 258 | 0.0087 | 3.2307 | 2010-2015 CNMI Bio- Sampling Database |
| Pristipomoides auricilla | 471 | 39.5 | 465 | 0.0189 | 3.0060 | 2010-2015 CNMI Bio- Sampling Database |
| Pristipomoides filamentosus | 123 | 58.5 | 123 | 0.0773 | 2.5914 | 2010-2015 CNMI Bio- Sampling Database |
| Pristipomoides flavipinnis | 179 | 51.5 | 168 | 0.0133 | 3.0762 | 2010-2015 CNMI Bio- Sampling Database |
| Pristipomoides sieboldii | 112 | | | | | |
| Pristipomoides zonatus | 404 | 45.4 | 371 | 0.0180 | 3.0411 | 2010-2015 CNMI Bio- Sampling Database |
| Variola louti | 6 | | | | | |

2.3.3 Guam Coral Reef Ecosystem Components Life History

2.3.3.1 Age, Growth, and Reproductive Maturity

<u>Description</u>: Age determination is based on counts of yearly growth marks (annuli) and/or DGIs internally visible within transversely cut, thin sections of sagittal otoliths. Validated age determination is based on several methods including an environmental signal (bomb radiocarbon 14 C) produced during previous atmospheric thermonuclear testing in the Pacific and incorporated into the core regions of sagittal otolith and other aragonite-based calcified structures such as hermatypic corals. This technique relies on developing a regionally based aged coral core reference series for which the rise, peak, and decline of 14 C values is available over the known age series of the coral core. Estimates of fish age are determined by projecting the 14 C otolith core values back in time from its capture date to where it intersects with the known age 14 C coral reference series. Fish growth is estimated by fitting the length-at-age data to a VBGF. This function typically uses three coefficients (L_{∞} , k, and t_0), which together characterize the shape of the length-at-age growth relationship.

Length-at-reproductive maturity is based on the histological analyses of small tissue samples of gonad material that are typically collected along with otoliths when a fish is processed for life history studies. The gonad tissue sample is preserved, cut into five-micron sections, stained, and sealed onto a glass slide for subsequent examination. Based on standard cell structure features and developmental stages within ovaries and testes, the gender, developmental stage, and maturity status (immature or mature) is determined via microscopic evaluation. The percent of mature samples for a given length interval are assembled for each sex and these data are fitted to a three- or four-parameter logistic function to determine the best fit of these data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish have achieved reproductive maturity (L_{50}) . For species that undergo sex reversal (primarily female to male in the tropical Pacific region) - such as groupers and deeperwater emperors among the bottomfishes, and for parrotfish, shallow-water emperors, and wrasses among the coral reef fishes - standard histological criteria are used to determine gender and reproductive developmental stages that indicate the transitioning or completed transition from one sex to another. These data are similarly analyzed using a three or four-parameter logistic function to determine the best fit of the data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish of a particular species have or are undergoing sex reversal ($L\Delta_{50}$).

Age at 50% maturity (A_{50}) and age at 50% sex reversal ($A\Delta_{50}$) is typically derived by referencing the VBGF for that species and using the corresponding L_{50} and $L\Delta_{50}$ values to obtain the corresponding age value from this growth function. In studies where both age & growth and reproductive maturity are concurrently determined, estimates of A_{50} and $A\Delta_{50}$ are derived directly by fitting the percent of mature samples for each age (one-year) interval to a three- or four-parameter logistic function using statistical analyses. The mid-point of this fitted logistic function provides a direct estimate of the age at which 50% of fish of a particular species have achieved reproductive maturity (A_{50}) and sex reversal ($A\Delta_{50}$).

Category: Biological

¹ *E. carbunculus* is now known to be comprised of two distinct, non-interbreeding lineages (Andrews et al. 2016). Both species occur in the Mariana Archipelago and are likely both captured by fishermen but reported as one species.

<u>Timeframe</u>: N/A <u>Jurisdiction</u>: Guam

Spatial Scale: Archipelagic

<u>Data Source</u>: Sources of data are directly derived from research cruises sampling and market samples collected by the Guam contracted bio-sampling team which samples the catch of fishermen and local fish vendors. Laboratory analyses and data generated from these analyses reside with the PIFSC LHP. Refer to the "Reference" column in Table 48 for specific details on data sources by species.

Parameter Definitions:

 T_{max} (maximum age) – The maximum observed age revealed from an otolith-based age determination study. T_{max} values can be derived from ages determined by annuli counts of sagittal otolith sections and/or bomb radiocarbon (14 C) analysis of otolith core material. Units are years.

 L_{∞} (asymptotic length) – One of three coefficients of the VBGF that measures the mean maximum length at which the growth curve plateaus and no longer increases in length with increasing age. This coefficient reflects the estimated mean maximum length and not the observed maximum length. Units are centimeters.

k (growth coefficient) – One of three coefficients of the VBGF that measures the shape and steepness by which the initial portion of the growth function approaches its mean maximum length (L_{∞}) .

 t_0 (hypothetical age at length zero) – One of three coefficients of the VBGF whose measure is highly influenced by the other two VBGF coefficients (k and L_∞) and typically assumes a negative value when specimens representing early growth phases) are not available for age determination. This parameter can be fixed at 0. Units are years.

M (natural mortality) – This is a measure of the mortality rate for a fish stock and is considered to be directly related to stock productivity (i.e., high M indicates high productivity and low M indicates low stock productivity). M can be derived through use of various equations that link M to T_{max} and the VBGF coefficients (k and L_{∞}) or by calculating the value of the slope from a regression fit to a declining catch curve (regression of the natural logarithm of abundance versus age class) derived from fishing an unfished or lightly fished population.

 A_{50} (age at 50% maturity) – Age at which 50% of the sampled stock under study has attained reproductive maturity. This parameter is best determined based on studies that concurrently determine both age (otolith-based age data) and reproductive maturity status (logistic function fitted to percent mature by age class with maturity determined via microscopic analyses of gonad histology preparations). A more approximate means of estimating A_{50} is to use an existing L_{50} estimate to find the corresponding age (A_{50}) from an existing VBGF curve. Units are years.

 $A\Delta_{50}$ (age of sex switching) – Age at which 50% of the immature and adult females of the sampled stock under study is undergoing or has attained sex reversal. This parameter is best determined based on studies that concurrently determines both age (otolith-based age data) and reproductive sex reversal status (logistic function fitted to percent sex reversal by age class with sex reversal determined via microscopic analyses of gonad histology preparations). A more

approximate means of estimating $A\Delta_{50}$ is to use an existing $L\Delta_{50}$ estimate to find the corresponding age $(A\Delta_{50})$ from the VBGF curve. Units are years.

 L_{50} (length at which 50% of a fish population are capable of spawning) – Length at which 50% of the females of a sampled stock under study has attained reproductive maturity; this is the length associated with A_{50} estimates. This parameter is derived using a logistic function to fit the percent mature data by length class with maturity status best determined via microscopic analyses of gonad histology preparations. L_{50} information is typically more available than A_{50} since L_{50} estimates do not require knowledge of age and growth. Units are centimeters.

 $L\Delta_{50}$ (length of sex switching) – Length at which 50% of the immature and adult females of the sampled stock under study is undergoing or has attained sex reversal; this is the length associated with $A\Delta_{50}$ estimates. This parameter is derived using a logistic function to fit the percent sex reversal data by length class with sex reversal status best determined via microscopic analyses of gonad histology preparations. $L\Delta_{50}$ information is typically more available than $A\Delta_{50}$ since $L\Delta_{50}$ estimates do not require knowledge of age and growth. Units are centimeters.

Rationale: These nine life history parameters provide basic biological information at the species level to evaluate the productivity of a stock - an indication of the capacity of a stock to recover once it has been depleted. These parameters are also used as direct inputs into stock assessments. Currently, the assessment of coral reef fish resources in Guam is data-limited. Knowledge of these life history parameters support current efforts to characterize the resilience of these resources and provide important biological inputs for future stock assessment efforts and enhance our understanding of the species' likely role and status as a component of the overall ecosystem. Furthermore, knowledge of life histories across species at the taxonomic level of families or among different species that are ecologically or functionally similar can provide important information on the diversity of life histories and the extent to which species can be grouped (based on similar life histories) for future multi-species assessments.

Table 48. Available age, growth, reproductive maturity, and natural mortality information for prioritized coral reef ecosystem component species in Guam

| Cm a si a s | Age | , growth | , and r | eproducti | ve matur | ity parame | eters | Defenence |
|--------------------------|-----------------|-------------------|-------------------|---------------------|-------------------|-------------------|-------------------|----------------------------|
| Species | Tmax | L_{∞} | k | t ₀ | A_{50} | L_{50} | $L\Delta_{50}$ | Reference |
| Caranx melampygus | X ^a | Xª | Xª | Xª | X ^a | Xª | Xª | LHP (in progress) |
| Chlorurus frontalis | 11 ^d | 37.2 ^d | 0.71 ^d | -0.058 ^d | 1.55 ^d | 24.0 ^d | 34.3 ^d | Taylor and Choat (2014) |
| Epinephelus fasciatus | | | | | | | | |
| Lethrinus harak | | | | | | | | |
| Lethrinus olivaceus | | | | | | | | |

| Lutjanus fulvus | | | | | | | | |
|--------------------------|-----------------|-------------------|-------------------|---------------------|--|--|-------------------|----------------------------|
| Naso unicornis | 23 ^d | 49.3 ^d | 0.22 ^d | -0.048 ^d | f=4.0 ^d m=3.2 ^d | f=29.2 ^d m=27.1 ^d | | Taylor et al. (2014) |
| Scarus rubroviolaceus | 6 ^d | 37.6 ^d | 0.66 ^d | -0.062 ^d | 1.91 ^d | 27.1 ^d | 32.9 ^d | Taylor and Choat (2014) |
| Siganus spinus | | | | | | | | |

^a signifies estimate pending further evaluation in an initiated and ongoing study.

Parameter estimates are for females unless otherwise noted (f=females, m=males). Parameters T_{max} , t_0 , A_{50} , and $A\Delta_{50}$ are in units of years; L_{∞} , L_{50} , and $L\Delta_{50}$ are in units of mm FL; k is in units of year⁻¹; X=parameter estimate too preliminary or Y=published age and growth parameter estimates based on DGI numerical integration technique and likely to be inaccurate; NA=not applicable. Superscript letters indicate status of parameter estimate (see footnotes below table). Published or in press publications (^d) are denoted in the "Reference" column.

2.3.3.2 Fish Length Derived Parameters

<u>Description</u>: The NMFS Commercial Fishery Bio-sampling Program started in 2009. This program has two components: first is the Field/Market Sampling Program, and the second is the Lab Sampling Program, details of which are described in a separate section of this report. The goals of the Field/Market Sampling Program are:

- Broad scale look at commercial landings (by fisher/trip, gear, and area fished);
- Length and weight frequencies of whole commercial landings per fisher-trip (with an effort to also sample landings not sold commercially);
- Accurate species identification;
- Develop accurate local length-weight curves.

In Guam, the Bio-sampling Program was focused on the commercial coral reef spear fishery with occasional sampling of the bottomfish fishery occurring locally and less frequently at the northern islands. However, in 2020 the Program switched focus to the MUS. Sampling is conducted in partnership with the fish vendors and fishermen. The Market Sampling information includes (but not limited to): 1) fish length; 2) fish weight; 3) species identification; and 4) basic effort information.

Category: Biological

Timeframe: N/A

Jurisdiction: Guam

^b signifies a preliminary estimate taken from ongoing analyses.

^c signifies an estimate documented in an unpublished report or draft manuscript.

^d signifies an estimate documented in a finalized report or published journal article (including in press).

Spatial Scale: Archipelagic

Data Source: NMFS Bio-sampling Program

Parameter Definitions:

n – *sample size* is the total number of fish sampled for length for each species recorded in the Bio-Sampling Program database.

 L_{max} – maximum fish length is the largest individual per species recorded in the Bio-Sampling Program database from the commercial spear fishery. This value is derived from measuring the length of individual samples for species occurring in the spear fishery. Units are centimeters.

 N_{L-W} – sample size for L-W regression is the number of samples used to generate the a and b coefficients.

a and b – length-weight coefficients are the coefficients derived from the regression line fitted to all length and weight measured by species in the commercial spear fishery. These values are used to convert length information to weight. Values are influenced by the life history characteristics of the species, geographic location, population status, and nature of the fisheries from which the species are harvested.

Rationale: Length derived information is an important component of fisheries monitoring and data poor stock assessment approaches. Maximum length (L_{max}) is used to derive missing species- and location-specific life history information (Nadon et al. 2015, Nadon and Ault 2016, Nadon 2019). The length-weight coefficients (a and b values) are used to convert length to weight for fishery-dependent and fishery-independent data collection where length is typically recorded but weight is the factor being used for management. This section of the report presents the best available information for the length derived variables for the Guam coral reef fisheries.

Table 49. Available length derived information for prioritized coral reef ecosystem component species in Guam

| Smaaiga | I | ength d | erived pa | rameters | | Deference |
|--------------------------|--------|-----------|-----------|----------|------|------------------------|
| Species | n | L_{max} | N_{L-W} | а | b | Reference |
| Caranx melampygus | 1,157 | 69.8 | 551 | 0.0228 | 2.95 | Kamikawa et al. (2015) |
| Chlorurus frontalis | 534 | 48.5 | 238 | 0.0172 | 3.08 | Kamikawa et al. (2015) |
| Epinephelus fasciatus | 4,223 | 57.0 | 1701 | 0.0118 | 3.08 | Kamikawa et al. (2015) |
| Lethrinus harak | 886 | 29.9 | 258 | 0.0281 | 2.89 | Kamikawa et al. (2015) |
| Lethrinus olivaceus | 751 | 71.7 | 272 | 0.0200 | 2.93 | Kamikawa et al. (2015) |
| Lutjanus fulvus | 426 | 29.6 | 91 | 0.0134 | 3.12 | Kamikawa et al. (2015) |
| Naso unicornis | 20,618 | 57.2 | 7790 | 0.0267 | 2.92 | Kamikawa et al. (2015) |
| Scarus rubroviolaceus | 2,563 | 47.8 | 1713 | 0.0114 | 3.18 | Kamikawa et al. (2015) |

| Chasins | I | Length do | erived pa | Deference | | |
|----------------|-------|-----------|-----------|-----------|------|------------------------|
| Species | n | L_{max} | N_{L-W} | a | b | Reference |
| Siganus spinus | 5,475 | 27.0 | 890 | 0.0284 | 2.87 | Kamikawa et al. (2015) |

2.3.4 Guam Management Unit Species Life History

2.3.4.1 Age, Growth, and Reproductive Maturity

Description: Age determination is based on counts of yearly growth marks (annuli) and/or DGIs internally visible within transversely cut, thin sections of sagittal otoliths. Validated age determination is based on several methods including an environmental signal (bomb radiocarbon 14 C) produced during previous atmospheric thermonuclear testing in the Pacific and incorporated into the core regions of sagittal otolith and other aragonite-based calcified structures such as hermatypic corals. This technique relies on developing a regionally based aged coral core reference series for which the rise, peak, and decline of 14 C values is available over the known age series of the coral core. Estimates of fish age are determined by projecting the 14 C otolith core values back in time from its capture date to where it intersects with the known age 14 C coral reference series. Fish growth is estimated by fitting the length-at-age data to a VBGF. This function typically uses three coefficients (L_{∞} , k, and t_0), which together characterize the shape of the length-at-age growth relationship.

Length-at-reproductive maturity is based on the histological analyses of small tissue samples of gonad material that are typically collected along with otoliths when a fish is processed for life history studies. The gonad tissue sample is preserved, cut into five-micron sections, stained, and sealed onto a glass slide for subsequent examination. Based on standard cell structure features and developmental stages within ovaries and testes, the gender, developmental stage, and maturity status (immature or mature) is determined via microscopic evaluation. The percent of mature samples for a given length interval are assembled for each sex and these data are fitted to a three- or four-parameter logistic function to determine the best fit of these data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish have achieved reproductive maturity (L_{50}) . For species that undergo sex reversal (primarily female to male in the tropical Pacific region) - such as groupers and deeperwater emperors among the bottomfishes, and for parrotfish, shallow-water emperors, and wrasses among the coral reef fishes - standard histological criteria are used to determine gender and reproductive developmental stages that indicate the transitioning or completed transition from one sex to another. These data are similarly analyzed using a three or four-parameter logistic function to determine the best fit of the data based on statistical analyses. The mid-point of this fitted function provides an estimate of the length at which 50% of fish of a particular species have or are undergoing sex reversal ($L\Delta_{50}$).

Age at 50% maturity (A_{50}) and age at 50% sex reversal ($A\Delta_{50}$) is typically derived by referencing the VBGF for that species and using the corresponding L_{50} and $L\Delta_{50}$ values to obtain the corresponding age value from this growth function. In studies where both age & growth and reproductive maturity are concurrently determined, estimates of A_{50} and $A\Delta_{50}$ are derived directly by fitting the percent of mature samples for each age (one-year) interval to a three- or four-parameter logistic function using statistical analyses. The mid-point of this fitted logistic function provides a direct estimate of the age at which 50% of fish of a particular species have achieved reproductive maturity (A_{50}) and sex reversal ($A\Delta_{50}$).

Category: Biological

<u>Timeframe</u>: N/A <u>Jurisdiction</u>: Guam

Spatial Scale: Archipelagic

<u>Data Source</u>: Sources of data are directly derived from research cruises sampling and market samples collected by the Guam-contracted bio-sampling team which samples the catch of fishermen and local fish vendors. Laboratory analyses and data generated from these analyses reside with the PIFSC LHP. Refer to the "Reference" column in Table 50 for specific details on data sources by species.

Parameter Definitions: Identical to Section 2.3.3.1

Rationale: These nine life history parameters provide basic biological information at the species level to evaluate the productivity of a stock - an indication of the capacity of a stock to recover once it has been depleted. Currently, the assessment of coral reef fish resources in Guam is datalimited. Knowledge of these life history parameters support current efforts to characterize the resilience of these resources and also provide important biological inputs for future stock assessment efforts and enhance our understanding of the species' likely role and status as a component of the overall ecosystem. Furthermore, knowledge of life histories across species at the taxonomic level of families or among different species that are ecologically or functionally similar can provide important information on the diversity of life histories and the extent to which species can be grouped (based on similar life histories) for future multi-species assessments.

Parameter estimates are for females unless otherwise noted (f=females, m=males). Parameters T_{max} , t_0 , A_{50} , and $A\Delta_{50}$ are in units of years; L_{∞} , L_{50} , and $L\Delta_{50}$ are in units of mm FL; k is in units of year⁻¹; X=parameter estimate too preliminary or Y=published age and growth parameter estimates based on DGI numerical integration technique and likely to be inaccurate; NA=not applicable. Superscript letters indicate status of parameter estimate (see footnotes below table). Published or in press publications (^d) are denoted in the "Reference" column.

Table 50. Available age, growth, reproductive maturity, and natural mortality information for MUS in Guam

| g · | | | Age, grow | th, and rep | roductive | e maturity p | arametei | :s | | D. C. |
|---|--|--|--|--|-------------------|--|----------------|---|----------------|----------------------------|
| Species | T_{max} | L_{∞} | k | t_0 | M | A_{50} | $A\Delta_{50}$ | L_{50} | $L\Delta_{50}$ | Reference |
| Aphareus rutilans | | | | | | | NA | | NA | |
| Caranx ignobilis | | | | | | | NA | | NA | |
| Caranx lugubris | | | | | | | NA | | NA | |
| Etelis carbunculus ¹ | | | | | | | NA | | NA | |
| Etelis coruscans | | | | | | | NA | | NA | |
| Lethrinus rubrioperculatus | | | | | | | NA | | NA | |
| Lutjanus kasmira | | | | | | | NA | | NA | |
| Pristipomoides auricilla ² | 18 ^d | 32.5 ^d | 0.60 ^d | | 0.18 ^d | | NA | | NA | O'Malley et al. (2019) |
| Pristipomoides filamentosus ² | 31° | 54.6° | 0.19 ^c | | | f=5.0° m=2.8° | NA | f=41.2° m=27.6° | NA | Villagomez (2019) |
| Pristipomoides flavipinnis | | | | | | | NA | | NA | |
| Pristipomoides sieboldii | | | | | | | NA | | NA | |
| Pristipomoides zonatus | f=19 ^d m=30 ^d | f=35.3 ^d m=38.3 ^d | f=0.27 ^d m=0.29 ^d | f=-2.03 ^d m=-1.54 ^d | 0.22 ^d | f=1.54 ^d m=1.79 ^d | NA | f=22.5 ^d m=24.12 ^d | NA | Schemmel et al. (2021) |
| Variola louti | f=11 ^c m=14 ^c | 43.5° | 0.26° | -1.1° | 0.20° | 2.2° | 5.9° | 26.0° | 35.5° | Schemmel et al. (in press) |

¹ E. carbunculus is now known to be comprised of two distinct, non-interbreeding lineages (Andrews et al. 2016). Both species occur in the Samoa Archipelago and were likely both captured by fishermen in the 1980s but reported as one species.

² Estimates are for the southern portion of the Mariana Archipelago.

^a signifies estimate pending further evaluation in an initiated and ongoing study.

^b signifies a preliminary estimate taken from ongoing analyses.

^c signifies an estimate documented in an unpublished report or draft manuscript.

d signifies an estimate documented in a finalized report or published journal article (+ in press).

2.3.4.2 Fish Length Derived Parameters

Description: The NMFS Commercial Fishery Bio-sampling Program started in 2009. This program has two components: first is the Field/Market Sampling Program and the second is the LHP, details of which are described in a separate section of this report. The goals of the Field/Market Sampling Program are:

- Broad scale look at commercial landings (by fisher/trip, gear, and area fished);
- Length and weight frequencies of whole commercial landings per fisher-trip (with an effort to also sample landings not sold commercially);
- Accurate species identification;
- Develop accurate local length-weight curves.

In Guam, the Bio-sampling Program was focused on the commercial coral reef spear fishery with occasional sampling of the bottomfish fishery occurring locally and less frequently at the northern islands. However, in 2020 the Program switched focus to the MUS. Sampling is conducted in partnership with the fish vendors and fishermen. The Market Sampling information includes (but not limited to): 1) fish length; 2) fish weight; 3) species identification; and 4) basic effort information.

Category: Biological

<u>Timeframe:</u> N/A <u>Jurisdiction:</u> Guam

Spatial Scale: Island

<u>**Data Source:**</u> NMFS Bio-sampling Program

Parameter definition: Identical to Section 2.3.3.2

Rationale: Length derived information is an important component of fisheries monitoring and data poor stock assessment approaches. Maximum length (L_{max}) is used to derive missing species- and location-specific life history information (Nadon et al. 2015; Nadon and Ault 2016; Nadon 2019). The length-weight coefficients (a and b values) are used to convert length to weight for fishery-dependent and fishery-independent data collection where length is typically recorded but weight is the factor being used for management. This section of the report presents the best available information for the length derived variables for the Guam MUS fisheries.

Table 51. Available length derived information for MUS in Guam

| Smaaiaa | I | Length o | lerived | paramete | rs | Deference |
|---------------------------------|------|-----------|-----------|----------|------|------------------------|
| Species | n | L_{max} | N_{L-W} | a | b | Reference |
| Aphareus rutilans | 184 | 90.5 | 86 | 0.0343 | 2.77 | Kamikawa et al. (2015) |
| Caranx ignobilis | 371 | | | | | |
| Caranx lugubris | 309 | 80.8 | 58 | 0.0250 | 2.94 | Kamikawa et al. (2015) |
| Etelis carbunculus ¹ | 888 | 63.4 | 575 | 0.0159 | 3.03 | Kamikawa et al. (2015) |
| Etelis coruscans | 476 | 95.0 | 255 | 0.0425 | 2.75 | Kamikawa et al. (2015) |
| Lethrinus rubrioperculatus | 7681 | 46.6 | 2196 | 0.0228 | 2.94 | Kamikawa et al. (2015) |
| Lutjanus kasmira | 1395 | 30.3 | 460 | 0.0128 | 3.12 | Kamikawa et al. (2015) |

| Charles | Length derived parameters | | | | | D -f | |
|-----------------------------|---------------------------|-----------|-----------|--------|------|------------------------|--|
| Species | n | L_{max} | N_{L-W} | а | b | Reference | |
| Pristipomoides auricilla | 3345 | 39.0 | 1210 | 0.0135 | 3.11 | Kamikawa et al. (2015) | |
| Pristipomoides filamentosus | 277 | 67.4 | 114 | 0.0225 | 2.93 | Kamikawa et al. (2015) | |
| Pristipomoides flavipinnis | 657 | 59.4^2 | 223 | 0.0210 | 2.95 | Kamikawa et al. (2015) | |
| Pristipomoides sieboldii | 411 | 63.2 | 130 | 0.0243 | 2.91 | Kamikawa et al. (2015) | |
| Pristipomoides zonatus | 925 | 57.5 | 329 | 0.0180 | 3.04 | Kamikawa et al. (2015) | |
| Variola louti | 1149 | 49.0 | 716 | 0.0130 | 3.09 | Kamikawa et al. (2015) | |

¹ E. carbunculus is now known to be comprised of two distinct, non-interbreeding lineages (Andrews et al. 2016). Both species occur in the Samoa Archipelago and were likely both captured by fishermen in the 1980s but reported as one species.

² The value in Kamikawa et al. (2015) is suspiciously high (76.6 cm). Guam Bio-Sampling database L_{max} is more reasonable, albeit still high.

2.4 SOCIOECONOMICS

This section outlines the pertinent economic, social, and community information available for assessing the successes and impacts of management measures or the achievements of the Fishery Ecosystem Plan for the Marianas Archipelago (WPRFMC 2009). It meets the objective "Support Fishing Communities" adopted at the 165th Council meeting; specifically, it identifies the various social and economic groups within the region's fishing communities and their interconnections. The section begins with an overview of the socioeconomic context for the region, and then provides a summary of relevant studies and data for CNMI and Guam, followed by summaries of relevant studies and data for each fishery in CNMI and Guam.

In 1996, the Magnuson-Stevens Fishery Conservation and Management Act's National Standard 8 (NS8) specified that conservation and management measures take into account the importance of fishery resources to fishing communities, to provide for their sustained participation in fisheries and to minimize adverse economic impacts, provided that these considerations do not compromise the achievement of conservation. Unlike other regions of the U.S., the settlement of the Western Pacific region was intimately tied to the sea (Figure 7), which is reflected in local culture, customs, and traditions.

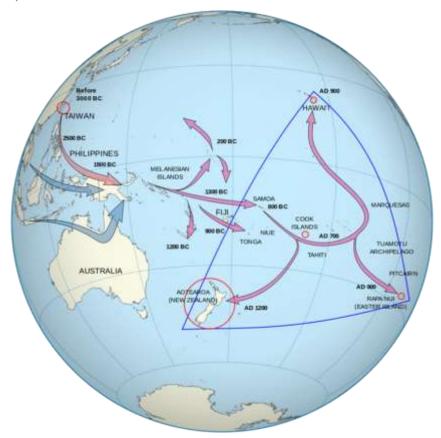


Figure 7. Settlement of the Pacific Islands, courtesy of Wikimedia Commons, https://commons.wikimedia.org/wiki/File:Polynesian_Migration.svg

Polynesian voyagers relied on the ocean and marine resources on their long voyages in search of new islands, as well as in sustaining established island communities. Today, the population of

the region also represents many Asian cultures from Pacific Rim countries, which reflect similar importance of marine resources. Thus, fishing and seafood are integral local community ways of life. This is reflected in the amount of seafood eaten in the region relative to the rest of the United States, as well as the language, customs, ceremonies, and community events. Because fishing is such an integral part of the culture, it is difficult to discern commercial from non-commercial fishing as most trips involving multiple motivations and multiple uses of the fish caught. While economics are an important consideration, fishermen report other motivations, such as customary exchange, as being equally important. Due to changing economies and westernization, recruitment of younger fishermen has become a concern for the sustainability of fishing and fishing traditions in the region.

The Marianas Archipelago consists of the Commonwealth of the Northern Mariana Islands (CNMI) at the northern end and Guam, the southernmost island. These are typically treated as two jurisdictions, which will be presented separately in the rest of this section despite being grouped under one FEP.

2.4.1 Response to Previous Council Recommendations

CNMI in March 2022, the Council directed staff to work with the Advisory Panels on restructuring fishermen's observation meetings and reports. PIFSC social scientists continued to assist the AP chairs, participating in organizational sessions, serving as note-takers, and providing synthesis reports.

At its 192nd meeting held via web conference and in Honolulu, HI in September 2022, the Council directed staff to incorporate scenario planning for extreme environmental events into EBFM-related planning. PIFSC, PIRO, and Council staff initiated a contract and began coordinating a training to build scenario planning capacity, which would be held in early 2023.

2.4.2 **CNMI**

2.4.2.1 Introduction

An overview of CNMI history, culture, geography, and relationship with the U.S. is described in the Fishery Ecosystem Plan for the Mariana Archipelago (WPRFMC 2009). Over the past decade, a number of studies have synthesized more specifics about the role of fishing and marine resources across CNMI, as well as information about the people who engage in the fisheries or use fishery resources.

The ancestors of the indigenous Chamorro first arrived in the Marianas around 3,500 years ago and relied on seafood as their principal source of protein (Allen and Amesbury 2012, Grace McCaskey 2014). Similar to other archipelagos in the Western Pacific, fish and marine resources have played a central role in shaping the social, cultural, and economic fabric of the CNMI that continues today. They fished for both reef and pelagic species, collected mollusks and other invertebrates, and caught sea turtles. The occupation of CNMI by foreign nations dramatically changed the island's ecosystems, reshaped communities, and disrupted fishing traditions. In the 17th and 18th centuries, Spanish colonizers destroyed the Chamorro's seagoing canoes, suppressed offshore fishing practices, and relocated populations from their traditional home. The CNMI was briefly occupied by Germany from 1899 to the beginning of WWII. During WWII, the CNMI was occupied by the Japanese military, and then was captured by the United States. Throughout this time, fishing remained an important activity. Later immigrants to the islands from East and Southeast Asia also possessed a strong fishing tradition. Today, only Saipan, Rota,

and Tinian are permanently inhabited, with 90% of the population living on the island of Saipan. Although the CNMI has transitioned to a tourism-based economy, fishing still plays an important cultural role and serves as a reliable source of local food (Ayers 2018).

Examination of the seascape of compliance across the US Pacific Island region found, that while the literature highlights the importance of enforcement, local experts emphasized barriers of capacity, governance process, and the lack of data. This suggests that non-instrumental and governance approaches can complement enforcement and should be part of an integrated compliance approach both in the region (Ayers and Leong 2020).

2.4.2.2 Equity and Environmental Justice

NOAA Fisheries equity and environmental justice (EEJ) goals are to 1) Prioritize identification, equitable treatment, and meaningful involvement of underserved communities, 2) Provide equitable delivery of services and 3) Prioritize EEJ in our mandated and mission work with demonstrable progress.

NOAA Fisheries commitment to EEJ is particularly relevant to the Pacific Islands Region. While every community is a fishing community in the Pacific Islands Region, there are specific features of these communities that can create barriers to EEJ. While some are shared across the region such as comparatively smaller populations and geographic isolation for NOAA Fisheries headquarters, others are specific to the cultural and political context of each archipelago, territory and commonwealth.

In this first year of adding EEJ to the SAFE report we will report a synthesis of feedback from partners and communities collected in informal listening sessions conducted in 2022. We have also included information from the NOAA Climate and Economic Justice Screening Tool Index of disadvantaged communities (https://coast.noaa.gov/digitalcoast/tools/cejst.html).

Going forward we will work to further develop this section to highlight the social and cultural impacts of fisheries science and management, and highlight the EEJ issues specific to archipelagic fisheries.

2.4.2.2.1 2022 Listening Sessions

With the support of NOAA Fisheries leadership, meetings relating to EEJ were held in person in Saipan, Tinian and Rota from August 14-20, 2022. The purpose of these meetings was to meet with key members of fishing communities, partners, and potentially underserved communities for feedback on the draft National Strategy for Equity and Environmental Justice (EEJ), and begin to build the foundation for developing the regional EEJ implementation plan. From these meetings PIFSC social scientists synthesized key EEJ barriers and issues.

Staff from the NOAA Pacific Islands Fisheries Science Center (PIFSC) met with over 60 individuals total in the CNMI, with representatives from: five fish markets; the fishing community; cultural practitioners; offices of the Governor, Lt. Governor, and the Mayors of Saipan, Northern Islands, Tinian, and Rota; and territorial agencies including the Division of Fish and Wildlife, Office of Planning and Development, and Division of Coastal Resources Management. Public meetings were held on Saipan, Tinian, and Rota, yielding participation from over 40 community members.

The Saipan meeting was held at the Hyatt, and included representation from such organizations as the United Carolinian Association, Micronesian Environmental Services, Tasi to Table, 500 Sails, and Council Advisory Panel. Collaborators identified through the Advisory Panel helped to organize meetings on Tinian and Rota, held at the Tinian Conference Center and Teteto beach pavilion, respectively.

Prior to travel, plan teams composed of on-island consultants met virtually to identify relevant stakeholders and discuss engagement practices. CNMI's plan team included individuals on contract with the Western Pacific Fishery Management Council, and staff from the NOAA Pacific Islands Regional Office (PIRO) and NOAA Office for Coastal Management. PIFSC staff also met with Honolulu-based NOAA Fisheries and Council staff conducting similar work in Hawaii and American Samoa for coordination, including PIFSC social scientists and PIRO communications staff.

All key themes were reviewed by partners prior to sharing.

2.4.2.2.2 Key EEJ themes

- NOAA Fisheries' work should be derived from people in the CNMI, to benefit the community, given the knowledge of and impact to these stakeholders:
- NOAA Fisheries should support the autonomy of territorial agencies and people instead of taking a top-down approach
- Seafood market development opportunities lacking
- Lack of feedback and communication to territorial agencies and people, despite NOAA
 Fisheries' research presence
- Federal regulation burdens locals, with apparent inability to regulate foreign fleets
- Improve engagement and communication with partners and communities
- Engage with more diverse groups
- Equity begins with return of ancestral lands by federal government
- Lack of continuity and communication within and across federal agencies, including high turnover within NOAA, makes progress slow and difficult

2.4.2.2.3 Index of Disadvantage

The NOAA Climate and Economic Justice Screening Tool has identified 72% of CNMI census tract communities (N=25) as disadvantaged

2.4.2.3 People who Fish

Allen and Amesbury (2012) summarized results of studies that demonstrated the sociocultural importance of fishing to Saipan residents. In a 2005 study, most of the active or commercial fishermen who responded to the survey had fished for more than 10 years. They most often participated in snorkel spearfishing at night (participated in by 73% of the fishermen) and snorkel spear fishing during daytime (58% of the fishermen), followed by hook-and-line less than 100 ft. deep (36%), trolling (21%), cast net (talaya; 14%), hook-and-line more than 100 ft. deep (9%), trapping (octopus, crabs, etc.; 19%), and foraging the reef (8%); 18% said they participated in one or more other techniques. Less than a third (~30%) said they owned a boat. The primary reasons for fishing were social, cultural, and nutrition; in addition to reporting that

they enjoy the activity itself (32%), many said they needed the fish to feed their family (23%), give to family and friends to strengthen social bonds (13%), that their family has always fished (12%), and that it strengthens bonds with their children/family (6%). Only 4% said they needed the money from the fish they sold. Other motivations included strengthening the bond with their fellow fishermen, fishing to catch fish for festivals and parties, and seasonal fishing for manahak, ti'ao, and i'e (2% each).

The fishermen reported fishing an average of 71 days per year, with 26% going once every two to three days, and 24% fishing once every two weeks. Those surveyed also reported a decrease in the amount of time they have spent fishing in the past decade, fishing 93 days per year on average. Saipan reef fish were the most frequently harvested species (caught by 54% of the fishermen), followed by shallow-water bottomfish (23%) and reef invertebrates such as octopus, shellfish and crabs (14%).

As in other parts of the region, much of the fisher's catch in the CNMI was consumed by themselves and their immediate family (70%), with another 20% consumed by extended family and friends. Only 8% of the catch was sold. There were 18 respondents that identified themselves as commercial fishermen. They reported a median monthly income of \$200 from fishing, with average monthly income of just over \$1,000. Costs exceeded sales for almost every income category for fishermen, suggesting that fishing is not a business for most, but that catch is simply sold to cover some of the cost.

While fish remain an important part of the local diet and an integral part of the people's history and culture, adaptation to and integration with a more westernized lifestyle appears to have changed people's dietary preferences on Saipan. Nearly half (45%) of the survey respondents reported eating "somewhat less fish" than they did a decade ago, although the majority still ate fish between one and three times a week. The majority also purchased their fish from a store or restaurant (40%), while 31% purchased fish from roadside vendors. Less common was acquiring fish from an extended relative/friend (13%) or their own catch (11%). Most of the fish consumed came from the U.S. mainland (41%), with other important sources coming from Saipan's coral reefs (31%), deepwater or pelagic fish caught off of Saipan (23%), or fish imported from other Pacific islands (e.g., Chuuk; 10%).

Few other surveys have been conducted on fishing in the CNMI. A household survey conducted in 2012 found that 37% of households had at least one individual that self-identified as a fisherman (Kotowicz and Allen 2015). Respondents from fishing households tended to be younger, possess lower education levels, and have a higher rate of unemployment than respondents from non-fishing households.

While proportionally few residents own a boat, more than 400 vessels were registered in the CNMI small boat fleet between 2010 and 2011 (Allen and Amesbury 2012). More than 200 of the vessels were active and operating in CNMI waters at that time, and more than 100 of the vessels were involved in fishing activities. The active small boat fleet targeted tunas, other small pelagics (through trolling), and bottomfish; with the increase in gas prices, however, pelagic fishing has waned. When caught, these fish are marketed locally, given away to family and friends, or used for ceremonial purposes such as parties, culturally significant fiestas, and the patron saint's days for each village.

On Saipan, fisheries managers estimated the active small boat fleet at approximately 100 vessels from 2010 to 2011. Full-time commercial fishing is primarily conducted by ethnic

nonindigenous minorities, namely Filipino residents that fish primarily as independent owners and/or operators and recent immigrants from the Federated States of Micronesia that fish for income. Chamorro and Carolinians, in contrast, primarily fish for recreational and subsistence purposes, typically only selling catch to recoup costs. A few vessel owner operators are considered "pescadors", a term used to refer to fishermen who provide fish for important community and familial events. Pescadors customarily provide 100-200 lb of reef fish for cooked dishes and pelagic species for kelaguen (a raw fish dish) used in community and family celebrations. The system of seafood distribution underwent significant changes from approximately the turn of the century with the establishment of large seafood vendors. In contrast to individual fishermen/vendors who only market their own catch, large vendors typically own and operate a number of vessels and purchase catch from independent fishermen to sell. This trend has reportedly caused prices to decline. In addition, increases in fuel prices, low market prices for fish, and downturns in the domestic economy have led to a general decline in participation in this fishery since 2000 in numbers of fishermen, trips, landings, and seafood purchasers. The Saipan Fishermen's Association (SFA) is a nonprofit organization established in 1985 that holds annual fishing derbies and participated in community involvement projects, such as beach cleanup.

On Tinian, estimates of fleet size range from 15 to 20 vessels in 2010-2011. An estimated one to three fishermen fished consistently with the primary intent of selling fish. Respondents suggested that fishing and eating of fish was more habitual, rather than geared toward a particular event. Increasing fuel prices have reportedly led to the decline in number of active fishermen, and fishermen frequently have sold fish to cover fuel costs. Three restaurants and two stores in Tinian purchase fish, although fishermen have also resorted to selling house-to-house; the fishermen commonly have an established clientele. A few charter boats serve tourist clientele; however, they do not land much catch, and even trolling trips serve more as photo opportunities. Charter boats are reportedly owned by non-local residents and target tourists by their country of origin (e.g., Japan, China, or Korea).

On Rota, fishermen target pelagic species when in season and bottomfish the rest of the year. Like on the other islands, the number and activity of fishermen have declined as a result of increased fuel prices. Family members will often make requests for certain kinds of fish, but they will also contribute money to purchase fuel for a fishing trip. In addition, fishermen will often check demand with local restaurants. In 2010 and 2011, fishermen sold catch to three separate restaurants or to neighbors and friends within the community (door-to-door or from a cooler on the roadside). One general store sold fish caught by a family member, who fished specifically to sell to that store. Rota holds a fishing derby in celebration of San Francisco, saint of the island.

A survey of the small boat fleet was also conducted in 2011 (Hospital and Beavers 2014). Respondents were 41 years old and had been boat fishing for 15 years on average, providing evidence of a deep tradition of boat fishing in the CNMI. They were more likely to identify themselves as Chamorro relative to the general population of the CNMI, although they were equally likely to have been born in the CNMI. In general, fishermen were more educated than the general population and of comparable affluence. Pelagic trolling was the most popular gear type, followed by deepwater bottomfish fishing, shallow-water bottomfish fishing, and spearfishing. Most fishermen (71%) reported fishing adjacent to a Fish Aggregating Device (FAD) at some point in the past 12 months and did so on nearly 22% of their fishing trips. A high degree of

seasonal fishing effort was reported across most fishing fleet subgroups, though fishermen on Tinian and Rota were more likely to fish year-round than those on Saipan.

A majority of fishermen (74%) reported selling at least a portion of their catch in the past year. However, less than half of survey respondents (43%) indicated that they could always sell any fish that they wanted. A significant percentage of fish caught was consumed at home (28%) or given away to relatives, friends, or for cultural events (38%); this reflects the strong family and social connections associated with fishing in the CNMI. Approximately 29% of fish catch was sold, with the remaining catch either released (2%) or exchanged for goods and services (3%). Even fishermen who regularly sold fish still retained approximately 22% of their catch for home consumption, participation in traditional fish-sharing networks, and customary exchange. Additionally, 91% of survey respondents considered the bottomfish they catch to be an important source of food, and 93% considered the reef fish to be similarly important. These findings validate the significance of fishing in building and maintaining social networks, perpetuating fishing traditions, and providing fish to local communities as a source of food security.

Fishing in the CNMI is a social activity; only 3% of fishermen reported to fish alone, but 70% reported that their boat is used without them on occasion. In addition, the majority of fishermen (57%) agreed that, as a fisherman, they are respected by the greater community. Nearly a third of respondents were neutral (27%) regarding this sentiment, while some were hesitant to express an opinion or simply did not know (13%). The study found that very few fishers (3%) felt that they were not respected by the community.

The designation of the Marianas Trench Marine National Monument (the Monument) in 2009 has resulted in concerns about loss of fishing access (Richmond and Kotowicz 2015; Kotowicz and Richmond 2013; Kotowicz and Allen 2015; and Kotowicz et al. 2017). Despite long distance, high cost, and inconvenience, travel to the areas now protected by the Monument were rare but culturally significant events, and fishing was an essential component. While CNMI residents generally supported designation of the monument, awareness was low regarding specific impacts (Kotowicz et al. 2017). In addition, fishing households showed higher awareness of the Monument, but were less likely to strongly support it.

Overall, the CNMI small boat fisheries are a mix of subsistence, cultural, recreational, and quasicommercial fishermen whose fishing behaviors provide evidence of the importance of fishing to the people of the CNMI. For nearly all fishery participants, the social and cultural motivations for fishing far outweigh economic prospects. Nearly all fishermen supplement their income with other jobs and are predominantly subsistence fishermen.

During the COVID-19 pandemic, the diversification of CNMI fisheries and ability to adapt to shift from a national and global economy to a local one played a vital role in supporting local food systems, nutrition, food security, and community social cohesion (Kleiber et al. 2022, Smith et al. 2022).

2.4.2.3.1 CNMI Bottomfish

Bottomfish was one of the gear types included in the 2011 Small Boat Survey (Hospital and Beavers 2014). Overall fisher demographics and catch disposition were summarized in the previous section. Approximately 68% of respondents reported fishing for deepwater bottomfish and 65% for shallow-water bottomfish; additionally, 41% identified deepwater bottomfish as their primary target, and 49% identified shallow-water bottomfish as their primary target.

Approximately 37% of trips included some form of bottomfish fishing. In general, deepwater bottomfish fishing appeared to be associated with more commercially-motivated fishermen. Fishers who primarily targeted bottomfish sold over half of their catch (52%) to friends, neighbors, and co-workers. Some self-identified primarily as subsistence fishers (58% selected this category) and recreational expense fishers (41%), although respondents spanned all response categories (full-time commercial, part-time commercial, recreational expense, purely recreational, subsistence, and cultural). Nearly half identified multiple motivations (49%).

2.4.2.3.2 CNMI Reef Fish

Coral reef fish were also included in the 2011 small boat survey (Hospital and Beavers 2014). Unsurprisingly, fishermen targeting reef fish, on average, were slightly younger than others, likely due to the physical requirements of reef fishing. Approximately 54% of respondents reported atulai fishing, 50% reported spearfishing, and 12% reported net fishing. Atulai was identified as the primary choice by 46% of fishermen, while 38% indicated spearfishing was preferable, and 14% net fishing as their primary gear type. Fishers who primarily targeted reef fish sold almost half of their catch (45%) to friends, neighbors, and co-workers. They self-identified primarily as subsistence fishers (44%) and cultural fishers (38%), although respondents spanned all response categories (full-time commercial, part-time commercial, recreational expense, purely recreational, subsistence, and cultural). Over one-third identified multiple motivations (38%).

In addition to playing an important role in subsistence and cultural fishing, coral reef ecosystems of Saipan only have been estimated at a value of \$61 million, 70% of which is accounted for by tourism (Grace McCaskey 2014).

2.4.2.3.3 CNMI Crustaceans

There are currently no socioeconomics data specific to the crustacean fishery. Future reports will include new information as resources allow.

2.4.2.3.4 CNMI Precious Corals

There are currently no socioeconomic data specific to the precious coral fishery. Future reports will include new information as resources allow.

2.4.2.4 CNMI Economic Performance

2.4.2.4.1 CNMI Bottomfish Commercial Participation, Landings, Revenue, Prices

This section will describe trends in commercial pounds sold, revenues and prices, for the CNMI bottomfish fishery. Figure 8 presents the trends of commercial pounds sold and revenues of bottomfish fishery (BMUS and boat-base creel surveys only) during 2012-2022 and Figure 9 presents the trend of fish price of bottomfish sold for the same period. Supporting data for Figure 8 and Figure 9 are shown in Table 52. The table also includes the percentage of pounds sold relative to estimates of total pounds landed for the bottomfish fishery. Both nominal and adjusted values are included.

As shown in Figure 8, the commercial landings of CNMI bottomfish were quite stable except for the recent two years (2021 and 2022). Commercial landings and revenue in 2021 were at a historical high (at 38,947 pounds landed valued at \$208,904 in nominal term), considering the period of 2003-2022, and slightly lower in 2022 at 32,106 pounds landed valued at \$180,672. The pounds sold were 68%% of the total landings in 2022. The 2021 and 2022 increase in

commercial landings are likely due to the improvement in reporting from more vendors. A mandatory reporting requirement was implemented in 2019 and outreach efforts have likely resulted in more vendors reporting. In total 37 out of an estimated 42 vendors reported in 2021. Fish prices dropped in 2020 from but it went up to \$5.36 in 2021 and continued increasing in 2022 to \$5.62.

It is worth noting that the data for pounds caught and pounds sold are collected by two different data collection methods. The data for pounds sold are collected through "Commercial Sales Receipt Books" Program, while the data for pounds caught are collected through Boat- and Shore-based creel surveys (only "Boat-based Creel Survey" information are included in the SAFE report since the majority of the BMUS were caught by boat fishing trips). Both data series are generated from an expansion algorithm built on a non-census data collection program, and the survey coverage rates of two data collection methods may change independently across individual years. Therefore, the two time series may not move coherently with each other. For example, the low percentage of pounds sold compared to pounds caught could be due to low coverage of dealers participating in the Commercial Receipt Books Program, or ratios exceeding 100% could reflect differences between commercially important species present in commercial markets that may not be encountered often in creel surveys. In 2014, the ratio of pounds sold to pound caught of BMUS was particularly high, 210%, while the total pounds sold in 2014 were similar to the figures in previous years and the estimated pounds caught were particularly low for 2014. Similarly, there is a very large discrepancy in 2018. It seems that there could be data quality concerns for the pounds landed estimations in some years.

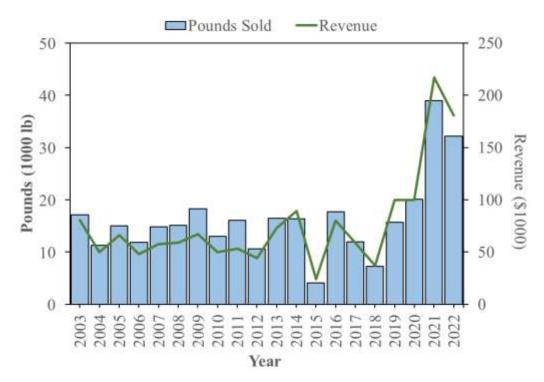


Figure 8. The commercial landings and revenues of BMUS, for the CNMI bottomfish fishery (adjusted to 2022 dollars*)

*Note: CPI information for CNMI has not been available since 2016, so this report assumed no CPI changes for the recent five years.

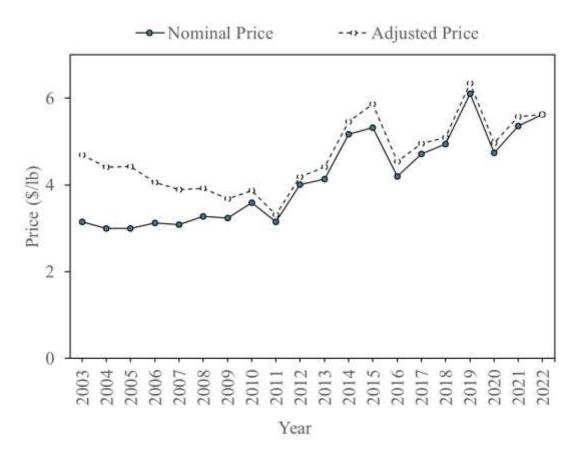


Figure 9. Price of BMUS for the CNMI bottomfish fishery

Estimated Estimated Estimated % of pounds pounds sold Estimated revenue (\$ pounds Fish price Fish price CPI Year adjustor caught (lb) (lb) revenue (\$) adj.) sold (\$) (\$ adj.) 2003 12,725 17,143 3.15 4.69 1.490 54,064 80,555 135% 2004 30,408 11,292 33,755 37% 2.99 4.42 1.477 49,856 2005 44% 4.42 1.474 34,313 15,026 45,086 66,457 3.00 2006 35,277 11,838 36,981 48,075 34% 3.12 4.06 1.300 2007 54,258 14,805 45,795 57,656 27% 3.09 3.89 1.259 2008 21,121 15,096 49,346 59,117 71% 3.27 3.92 1.198 2009 65,268 18,313 59,247 67,364 28% 3.24 3.68 1.137 2010 56,007 12,970 46,595 50,229 23% 3.59 3.87 1.078 2011 25,799 16,115 50,757 1.053 53,447 62% 3.15 3.32 2012 137,496 10,591 42,471 44,255 8% 4.01 4.18 1.042 2013 1.069 20,392 16,500 68,211 72,918 81% 4.13 4.41 2014 7,740 16,334 84,508 89,325 5.17 5.46 1.057 211% 2015 10,386 4,122 21,917 24,153 40% 5.32 5.86 1.102 2016 54,334 74,445 4.54 1.080 17,717 80,401 33% 4.20 2017 48,007 11,925 56,241 58,997 25% 4.72 4.95 1.049 2018 650 7,260 35,840 36,951 1117% 4.94 5.09 1.031 2019 21,012 15,699 95,801 99,729 75% 6.10 6.35 1.041 2020 45,547 20,071 95,197 99,862 44% 4.74 4.97 1.049

Table 52. Commercial landings and revenue information of CNMI bottomfish fishery

Data source: PIFSC FRMD-WPacFIN boat-based creel surveys (* CPI information for CNMI were not available since 2016, so this report assumed no CPI changes for the five years).

217,051

180,672

53%

68%

5.36

5.62

5.57

5.62

1.039

208,904

180,672

2.4.2.4.2 CNMI Bottomfish Costs of Fishing

38,947

32,160

73,863

47,418

2021

2022

Since 2009, PIFSC economists have maintained a continuous economic data collection program for small boat fisheries in Saipan through collaboration with PIFSC FRMD-WPacFIN (Chan and Pan 2019). The economic data collection program gathers fishing expenditure data for boat-based reef fish, bottomfish, and pelagic fishing trips on an ongoing basis. Data for fishing trip expenses include gallons of fuel used, price per gallon of fuel, cost of ice used, cost of bait and chum used, cost of fishing gear lost, and the engine type of the boat. These economic data are collected from same subset of fishing trips as the boat-based creel survey carried out by the local fisheries management agencies and PIFSC. Metadata for these data are available online (PIFSC 2016). Island-specific (Saipan, Tinian, and Rota) trip cost estimates for bottomfish fishing trips are available only for 2011 in Hospital and Beavers (2014). Other relevant cost information in Hospital and Beavers (2014) include estimates of annual fishing expenditures (fixed costs) and levels of investment in the fishery.

The trip cost data presented in this section were collected through the continuous economic data collection program on Saipan through collaboration with PIFSC. Figure 10 shows the trend of

average trip costs for CNMI bottomfish trips during 2012–2022 (adjusted to 2022 dollars). Supporting data of Figure 10 are presented in Table 53. The trip costs seem to have substantial inter-annual variability. The average cost for a bottomfish trip was \$43 in 2022, lower than the trip costs in 2021, due to lower fuel usage per trip. The cost data summaries were generated by excluding outliners (cases with >10 gallons/hours fished).

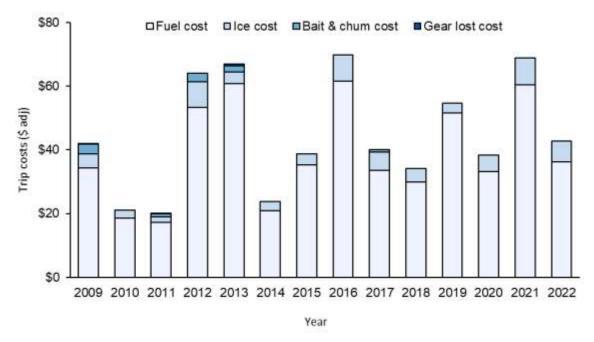


Figure 10. Average costs for CNMI bottomfish trips (adjusted to 2022 dollars)*

^{*} CPI information for CNMI were not available since 2016 so this report assumed no CPI changes for the four years. Data source: PIFSC Continuous Cost Data Collection Program (Chan and Pan 2019).

| | | Total trip | | | | Gear | Fuel price | |
|------|------------|------------|-----------|-----------|-----------|-----------|-------------|----------|
| | Total trip | cost adj. | Fuel cost | Ice cost | Bait cost | losted | adj. | CPI |
| Year | costs (\$) | (\$) | adj. (\$) | adj. (\$) | adj. (\$) | adj. (\$) | (\$/gallon) | Adjustor |
| 2009 | 37 | 42 | 34 | 4 | 3 | 0 | 3.74 | 1.137 |
| 2010 | 20 | 21 | 19 | 2 | 0 | 0 | 4.15 | 1.078 |
| 2011 | 19 | 20 | 17 | 2 | 1 | 0 | 4.94 | 1.053 |
| 2012 | 61 | 64 | 53 | 8 | 3 | 0 | 5.12 | 1.042 |
| 2013 | 63 | 67 | 61 | 4 | 2 | 1 | 5.35 | 1.069 |
| 2014 | 22 | 24 | 21 | 3 | 0 | 0 | 5.21 | 1.057 |
| 2015 | 35 | 39 | 35 | 3 | 0 | 0 | 4.50 | 1.102 |
| 2016 | 65 | 70 | 62 | 8 | 0 | 0 | 3.87 | 1.080 |
| 2017 | 38 | 40 | 34 | 6 | 1 | 0 | 4.06 | 1.049 |
| 2018 | 33 | 34 | 30 | 4 | 0 | 0 | 4.30 | 1.031 |
| 2019 | 53 | 55 | 52 | 3 | 0 | 0 | 4.10 | 1.041 |
| 2020 | 37 | 38 | 33 | 5 | 0 | 0 | 4.09 | 1.049 |
| 2021 | 66 | 69 | 60 | 8 | 0 | 0 | 4.98 | 1.039 |
| 2022 | 43 | 43 | 36 | 6 | 0 | 0 | 5.55 | 1 |

Table 53. Average trip costs for CNMI bottomfish trips (adjusted to 2022 dollars)*

2.4.2.4.3 CNMI Ecosystem Component Species

Based on new guidelines for the archipelagic SAFE report from the Council, this section highlights the top 10 Ecosystem Component Species (ECS) (sorted by landings) and the priority ECS (selected species recommended by the local fishery management agency) caught by small boats or shoreline fishing. Please note the top 10 species list and the priority species list reported in the socioeconomic module may not be consistent with the lists reported in the fishery module in the previous sections. The inconsistences result from several factors: 1) differences in data sources, 2) differences in level of species groupings, 3) differences in commercial landing vs. total landings.

Firstly, the data for pounds caught and pounds sold are collected by two different data collection methods, as mentioned in the earlier section. The data for "pounds sold" (commercial landings) reported in this socioeconomics module were collected through "Commercial Sales Receipt Books" Program, while the data of pounds caught were collected through "Boat-based Creel Survey". The survey coverage rates of two data collection methods may change independently in individual years. Secondly, the species groups used in the two data collection programs were different, as the species in the commercial receipt books are usually lumped into family levels or species groups while the species reported in the Creel Survey are often more detailed at the species level. Third, fish species with higher total pounds caught may not necessarily lead to higher pounds sold in the markets. Therefore, the two series may not move coherently with each other.

^{*} CPI information for CNMI were not available for 2016 and the year after. Data source: PIFSC Continuous Cost Data Collection Program (Chan and Pan 2019).

Table 54 shows the commercial landings and revenue of the top 10 ECS in CNMI. The total pounds sold of the top 10 species/species groups was 67,301 lb (valued at \$244,011) in 2022, slightly higher than the values of 2021.

Table 55 shows the ECS priority species. Eight fish species were suggested as priority species (species of interests) for the area. Four of the eight priority species showed up in the commercial receipt books while seven of them showed up with landings in 2022, similar to 2021. It seems the total commercial landings (pounds sold) reported in the commercial receipt books were higher than that total landings reported by creel surveys for both top 10 and priority species. However, comparison between commercial landings and total landings might not be straight forward since two data series were collected under different data collection systems and the species and species groupings might not be consistent between the two.

Table 54. Top 10 ECS Commercial landings, revenue, and price, 2021 and 2022

| Top 10 landings (Boat based) | Pounds kept | Top 10 commercial landings (receipt books) | Pounds sold | Revenue (\$) | Price (\$/lb) |
|------------------------------|----------------|--|-------------|--------------|------------------|
| Bigeye scad | 12,497 | Parrot (misc) /palakse/la | 15,342 | 69,082 | 4.50 |
| Tan-faced parrotfish | 7,464 | Bigeye scad | 20,295 | 57,769 | 2.85 |
| Orangespine unicornfish | 4,787 | Surgeonfish (misc.) | 7,429 | 24,015 | 3.23 |
| Bluespine unicornfish | 4,192 | Spiny lobster | 2,183 | 20,948 | 9.60 |
| Yellowtail kalikali | 4,176 | Goatfish/satmoneti | 6,325 | 20,738 | 3.28 |
| Yellowband parrotfish | 3,843 | Emperor (mafute/misc.) | 4,824 | 15,896 | 3.30 |
| Bluefin trevally | 3,300 | Rudderfish/guili | 4,123 | 13,173 | 3.20 |
| Longnose emperor | 2,275 | Squirrelfish/sagamelon | 2,399 | 7,870 | 3.28 |
| Blackspot emperor | 1,801 | Unicornfish/tataga | 2,368 | 7,563 | 3.19 |
| Stareye parrotfish | 1,513 | Jacks (misc.) | 2,013 | 6,957 | 3.46 |
| Sum 2022 | 45,848 | | 67,301 | 244,011 | 3.63 |
| Sum 2021 | 34,380 | | 53,196 | 163,636 | 3.08 |

Data source: PIFSC FRMD-WPacFIN boat-based creel survey, commercial receipt books.

Table 55. Priority ECS commercial landings, revenue, and price 2021 and 2022

| | | Landings | Com. | | |
|------|------------------------------|-------------------|----------|-----------|---------|
| | | Boat-Based | landings | Estimated | Price |
| Year | Common Name | (lbs) | (lbs) | revenue | (\$/lb) |
| 2022 | Forktail rabbitfish | 548 | 5,949 | 26,544 | 4.46 |
| 2022 | Orangespine unicornfish | 4,787 | 7,594 | 24,521 | 3.23 |
| 2022 | Bluespine unicornfish | 4,192 | 5,340 | 17,942 | 3.36 |
| 2022 | Bluebanded surgeonfish | 808 | 498 | 1,527 | 3.07 |
| 2022 | Redlip parrotfish | 13 | | | |
| 2022 | Thumbprint/blackspot emperor | 1,801 | | | |
| 2022 | Yellowstripe goatfish | 204 | | | |
| | Sum | 10,335 | 19,381 | 70,534 | 3.64 |
| 2021 | Forktail rabbitfish | 284 | 5,820 | 23,912 | 4.11 |
| 2021 | Orangespine unicornfish | 5,735 | 6,465 | 19,063 | 2.95 |
| 2021 | Bluespine unicornfish | 1,051 | 2,962 | 9,118 | 3.08 |
| 2021 | Bluebanded surgeonfish | 3,363 | 33 | 99 | 3.00 |
| 2021 | Redlip parrotfish | 80 | | | |
| 2021 | Blue-barred parrotfish | 16 | | | |
| 2021 | Thumbprint/blackspot emperor | 3,297 | | | |
| | Sum | 10,433 | 15,280 | 52,192 | 3.42 |

Data source: PIFSC FRMD, commercial receipt books.

2.4.3 Guam

2.4.3.1 Introduction

An overview of Guam's history, culture, geography, and relationship with the U.S. is described in the Fishery Ecosystem Plan for the Mariana Archipelago (WPRFMC 2009). Guam is the largest and southernmost island of the Mariana Archipelago and is also the largest and most heavily populated island in Micronesia. Over the past decade, a number of studies have synthesized more details about the role of fishing and marine resources for residents of Guam, as well as information about the people who engage in the fisheries and/or utilize fishery resources.

The ancestors of the indigenous Chamorro first arrived in the Marianas around 3,500 years ago, and were expert fishermen and seafarers, relying on seafood as their principal source of protein (Allen and Bartram 2008; Grace McCaskey 2014; Hospital and Beavers 2012). They fished on the high seas in large sailing canoes (proas) and used numerous methods to catch reef and bottomfish from boats. Similar to other archipelagos in the Western Pacific, fish and marine resources have played a central role in shaping the social, cultural, and economic fabric of Guam that continues today. Chamorro fished for both reef and pelagic species, collected mollusks and other invertebrates, and caught sea turtles.

The occupation of Guam by foreign nations dramatically changed the island's ecosystems, reshaped communities, and disrupted fishing traditions. In the 17th and 18th centuries, Spanish colonizers destroyed the Chamorro' seagoing canoes, suppressed offshore fishing practices, and relocated populations from their traditional home. Following the Spanish-American War in 1898,

the U.S. Navy took control of Guam until it was occupied by Japan from 1941-1944. Guam became a U.S. territory in 1950, and the U.S. military is currently in the process of building up an even greater presence on the island. Throughout this time, fishing has remained an important activity, although by the time Guam became and American territory, the indigenous inhabitants had lost many of their seafaring skills, fishing skills, and even the native names of many of the offshore species. Later immigrants to the islands from East and Southeast Asia also possessed a strong fishing tradition. In 2000, 37% of Guam's population that identified as a single ethnicity were Chamorro, followed by 32% Asian (about 80% of whom were Filipino), 17% other Pacific Islander, 7% white, and 1% black. Despite rapid socioeconomic change, households still reflect the traditional pattern of extended families with multigenerational clustering of relatives, especially in Guam's southern villages. Social occasions such as neighborhood parties, wedding and baptismal parties, wakes and funerals, and especially village fiestas that follow the religious celebrations of village patron saints all require large quantities of fish and other traditional foods, reflecting the role of fish in maintaining social ties and cultural identities. Sometimes fish are also sold to earn money to buy gifts for friends and relatives on important Catholic religious occasions such as novenas, births and christenings, and other holidays.

Since the late 1970s, Guam's most important role in commercial fisheries activity has been as a major regional fish transshipment center and resupply base for domestic and foreign tuna fishing fleets. Services provided include fueling, provisioning, unloading, air and sea transshipment, net and vessel repair, crew repatriation, medical care, and warehousing. Among Guam's advantages as a home port are well-developed and highly efficient port facilities in Apra Harbor, an availability of relatively low-cost vessel fuel, a well-established marine supply/repair industry. and recreational amenities for crew shore leave. In addition, the Territory is exempt from the Nicholson Act, which prohibits foreign ships from landing their catches in U.S. ports. Initially, the majority of vessels calling in Apra Harbor to discharge frozen tuna for transshipment were Japanese purse seine boats and carrier vessels. In the late 1980s, Guam became an important port for Japanese and Taiwanese longline fleets, but port calls have steadily declined and the transshipment volume has declined accordingly. By the early 1990s, an air transshipment operation had also been established in Guam. Fresh tuna was flown into Guam from the Federated States of Micronesia and elsewhere on air cargo planes and out of Guam to the Japanese market on wide-body passenger planes. Further, vessels from Japan and Taiwan also landed directly into Guam, where their fish were packed and transshipped by air to Japan. A second air transshipment operation began in the mid-1990s that was transporting fish to Europe that did not meet Japanese sashimi market standards, but this has since ceased. Moreover, the entire transshipment industry has contracted markedly with only a few operators still making transshipments to Japan. Annual volumes of tuna transshipped of between 2007 and 2011 averaged about 3,400 mt, with a 2012 estimate of 2,222 mt, compared to over 12,000 mt at the peak of operations between 1995 and 2001. As early as 2006, it was noted that the Port of Guam had lost much of its competitive advantage compared to alternative transshipment locations in the western Pacific and elsewhere, a trend that may not be reversible.

Otherwise, commercial fisheries have a relatively minor contribution to Guam's economy; the social and cultural importance of fisheries in Guam dwarfs their commercial value. Nearly all Guam domestic fishermen hold jobs outside the fishery, with fishing typically supplementing family subsistence. High value is placed on sharing one's fish catch with relatives and friends, and this social obligation extends to part-time and full-time commercial fishermen alike. A survey of Guam households in 2005 found that nearly one-quarter (24%) of fish consumed were

caught by the respondent or an immediate family member, and an additional 14% were caught by a friend or extended family member (Allen and Bartram 2008). However, a little more than half (51%) of the fish consumed were purchased at a store or restaurant, and 9% were purchased at a flea market or from a roadside stand. The same study found that annual seafood consumption in Guam is estimated to be about 60 lbs. per capita, with approximately 43% imported from the U.S.

The westernization of Guam, particularly since World War II, has not only resulted in a transition from a subsistence to wage-based economy, but has also contributed to dramatic changes in eating patterns, including lower seafood consumption. Indeed, recent years have seen steady declines in the market demand for fresh local fish across Guam (Hospital and Beavers 2012). While some families continue to supplement their diet by fishing and farming, no existing communities are completely dependent on local fishing as a source of food. A household survey conducted in 2016 found that only 29% of respondents participate in fishing (NCRMP 2016).

Allen and Bartram (2008) reviewed the history of shoreline and inshore fishing in Guam. They noted that the number of people engaged in shore fishing in the 1970s was surprisingly large, given that about 90% of the food consumed on the island was imported. A study conducted in 1975 found that 65% of households reported some participation in fishing, which was presumably shore-fishing as a result of the low level of boat ownership at the time. Creel surveys conducted by the Guam DAWR indicated that CPUE in Guam's shore-based fisheries for reef fish (pole, spear, cast net, surround net, and gill net) declined sharply in the 1980s and had not recovered by 2008. Offshore (boat-based) catches of reef-associated fish were relatively constant between 1992 and 2008, whereas inshore catches that accounted for the majority of the reef fish harvest during the 1990s comprised a minority of the total harvest by 2008. Much of the traditional harvest targets seasonal runs of juvenile rabbitfish, goatfish, bigeye scad (atulai, Selar crumenophthalmus), and jacks (i'e, family Carangidae). A study in 2007 estimated that Guam's coral reef resources were valued at close to \$127 million annually, primarily driven by the island's important tourism industry (Grace McCaskey 2014). Nearly 1.2 million people visited Guam in 2010, many of them attracted by reef-related activities, such as snorkeling and scuba diving.

As recently as the early 1970s, relatively few people from Guam fished offshore because boats and deep-sea fishing equipment were prohibitively expensive (Allen and Bartram 2008). During the economic boom from the late-1980s through most of the 1990s, Guam developed a small boat fishery that conducted trolling and bottomfish fishing mostly within 30 miles of shore.

The Guam Fishermen's Cooperative Association (GFCA) plays an important role in preserving important fishing traditions. It began operations in 1976 and was incorporated in 1977. In 2006, its membership included 164 full- and part-time fishermen from every district in Guam, and it processed and marketed approximately 80% of the local commercial catch. In addition, it plays a role in fisheries data collection, marine education and training, and fisheries conservation and management. The GFCA strives to provide benefits not just to fishermen but to residents throughout Guam, benefitting the broader Guam community. It utilizes a Hazard Analysis and Critical Control Point (HACCP) system to ensure safe seafood, and tests fish for potential toxins or whenever requested by the Guam Department of Health and Sanitation. It has also become a focal point for community activities, such as the Guam Marianas International Fishing Derby, cooking competitions, the Guam Fishermen's Festival, dissemination of educational materials on marine resources, vessel safety, seafood preparation, public meetings on resource management

issues, and communications via radio base to relay information and coordinate rescues. It also has adopted a policy of purchasing local origin products that benefits 40 small businesses in Guam, regularly donates seafood for village functions and charitable activities, and provides assistance to victims of periodic typhoons with emergency supplies of ice and fuel. In addition, the GFCA has become a voice for Guam fishermen in the policy arena to ensure that concerns of fishermen are incorporated into relevant issues, including the military buildup and loss of fishing grounds due to establishment of Marine Preserve Areas.

Fishing in Guam continues to be important not only in contributing to the subsistence needs of the Chamorro and other residents, but also in preserving their histories and identities. Knowledge of how fish are distributed and consumed locally is crucial to understanding the social and cultural significance of fishing in Guam.

2.4.3.2 Equity and Environmental Justice

NOAA Fisheries equity and environmental justice (EEJ) goals are to 1) Prioritize identification, equitable treatment, and meaningful involvement of underserved communities, 2) Provide equitable delivery of services and 3) Prioritize EEJ in our mandated and mission work with demonstrable progress.

NOAA Fisheries commitment to EEJ is particularly relevant to the Pacific Islands Region. While every community is a fishing community in the Pacific Islands Region, there are specific features of these communities that can create barriers to EEJ. While some are shared across the region such as comparatively smaller populations and geographic isolation for NOAA Fisheries headquarters, others are specific to the cultural and political context of each archipelago, territory and commonwealth.

In this first year of adding EEJ to the SAFE report we will report a synthesis of feedback from partners and communities collected in informal listening sessions conducted in 2022. We have also included information from the NOAA Climate and Economic Justice Screening Tool Index of disadvantaged communities (https://coast.noaa.gov/digitalcoast/tools/cejst.html).

Going forward we will work to further develop this section to highlight the social and cultural impacts of fisheries science and management, and highlight the EEJ issues specific to archipelagic fisheries.

2.4.3.2.1 2022 Listening Sessions

With the support of NOAA Fisheries leadership, meetings relating to EEJ were held in person in Guam from August 20-26, 2022. The purpose of these meetings was to meet with key members of fishing communities, partners, and potentially underserved communities for feedback on the draft National Strategy for Equity and Environmental Justice (EEJ), and begin to build the foundation for developing the regional EEJ implementation plan. From these meetings PIFSC social scientists synthesized key EEJ barriers and issues.

Staff from the NOAA Pacific Islands Fisheries Science Center met with over 30 individuals in Guam, with representation from 10 local government agencies and elected offices, 6 fish markets, fisheries participants, fishing tournament organizers, the Council and its Advisory Panel, researchers, and cultural experts. Institutions represented within the aforementioned categories include the Guam Governor's office, Merizo Mayor's office, Department of Chamorro Affairs, Commission on Decolonization, Department of Agriculture, Division of

Aquatic and Wildlife Resources, Climate Change Resiliency Commission, Guam Sea Grant, Bureau of Statistics and Plans, Humanities Guåhan, University of Guam 4H program, Guam Fishermen's Cooperative Association, and Micronesian Association. Although we contacted three mayors' offices prior to our arrival in the hopes of organizing in collaboration with them, we were only able to facilitate one public meeting, which was hosted by the Merizo Mayor and attended by five community members. About a third of our sample consisted of fisheries participants, making them the most represented group by number of individuals.

Prior to travel, plan teams composed of on-island consultants met virtually to identify relevant stakeholders, initiate correspondence with formal offices, and schedule meetings. Guam's plan team included representatives from the NOAA Pacific Islands Fisheries Science Center, NOAA Office for Coastal Management and the Western Pacific Fishery Management Council island coordinator. PIFSC staff also met with Honolulu-based NOAA Fisheries and Council staff conducting similar work in Hawaii and American Samoa for coordination, including PIFSC social scientists and PIRO communications staff.

All key themes were reviewed by partners prior to sharing.

2.4.3.2.2 Key EEJ themes

- Relationship between U.S. government/military to Guam as a barrier to sustainability and protection of land, ocean, and culture
- Locals pay cultural and financial costs for federal regulation, with foreign fishers and military relatively unaffected.
- NOAA Fisheries should support the autonomy of territorial agencies and people
- There is an apparent lack of continuity and communication within and across federal agencies, including high turnover within NOAA.
- NOAA Fisheries should align its work with local needs and understanding through improved consultation and funding processes.
- Concerns about pollution (e.g., DOD dumping of nuclear waste, PCBs at Cocos island) compound impacts of local and federal marine closures
- Improve engagement and communication with partners and communities
- Engage with more diverse groups
- Frustration with closure of federal and territorial areas to fishing

2.4.3.2.3 Index of Disadvantage

The NOAA Climate and Economic Justice Screening Tool has identified 16% of Guam census tract communities (N=57) as disadvantaged.

2.4.3.3 People who Fish

Few studies have been conducted on fishing in Guam in general. A household survey conducted in 2012 found that 35% of respondents said that they or someone else in their household was a fisherman (Kotowicz and Allen 2015). Respondents from fishing households tended to have lower education levels and have a higher rate of unemployment than respondents from non-fishing households.

As described in Allen and Bartram (2008), in 1999, a detailed study of the inshore fishing behaviors and spatial patterns was conducted for the three largest resident fishing cultures in Guam: Chamorro, Micronesian, and Filipino. At that time, Chamorro comprised about 75% of the fishing parties encountered, while Micronesians constituted about 17% and Filipinos about 7%. A number of contemporary reef fishing methods in Guam were observed, including gleaning, hand line, rod and reel, talaya (cast net), tekken (gill net), chenchulu (surround net), and spearfishing. Explicit rules governing permanent marine ownership were not observed, but Chamorro fishermen maintained a strong identification with village and municipal space. This village relationship included the reef during the early part of the 20th century but that has since largely disappeared. Instead, a system of "pliant tenure" (a vestige of traditional marine tenure) was recognized; while any reef area is publicly accessible, fishermen act according to a system of temporary ownership or pliant tenure of reef area. These rules were understood and incorporated by Chamorro and immigrant fishers alike. Respondents voiced concern about the loss of fishing grounds through designation of marine reserves and tourist watercraft activities. They viewed reduced coastal access as threatening the perpetuation of cultural identity and practice by reducing ability to teach and practice traditions such as communal harvests and distribution of the catches, which reinforce family cohesion and communal identity. These practices have been further jeopardized by the build-up of U.S. military personnel and families in recent years.

In the mid-1980s Guam fisheries were characterized as including (1) a small number of true commercial fishermen, (2) subsistence/recreational fishermen who regularly sell part of their catch, (3) a large number of subsistence fishermen who rarely sell any of their catch, and (4) a substantial number of recreational fishermen. Approximately 60% of catch was non-commercial, with fish sales primarily used to generate revenue to pay for fuel costs. A similar pattern continues in recent years.

In 2011, a survey was conducted of the small boat fleet, which included questions about trolling, bottomfish fishing, and reef fishing. On average, fishermen responding to the survey were 44 years old and reported to have been boat fishing for an average of 20 years. Respondents were also more educated and more affluent than the general population. The majority of respondents described themselves as Chamorro (72%), followed by white (23%) with relatively small proportions of Filipinos (6%), Micronesians (6%), other ethnicities (5%), and Carolinians (1%) represented. There was considerable evidence of co-ownership and sharing of fishing vessels. In addition, fishermen reported the use of multiple gear types, with pelagic trolling as the most popular gear type followed by shallow-water bottomfish fishing and deepwater bottomfish fishing. Almost all (96%) fishermen reported fishing at a Fish Aggregating Device (FAD) during the past year and on nearly half (53%) of their fishing trips. Fishing for bottomfish and reef fish was highly seasonal compared to pelagics. Whereas over half of the survey respondents (54%) fished all year for pelagics, only 16% fished year-round for bottomfish and reef fish.

Approximately 70% of fishermen reported selling at least a portion of their catch, and 82% could always sell all the fish that they wanted to sell. However, nearly 30% reported that they had not sold any fish in the past year, and nobody reported selling all the fish they caught. Instead, cost recovery was cited as the primary motivation for the sale of fish, with fish sales contributing very little to personal income for the majority of respondents (59%). In fact, 64% of fishermen reporting the sale of fish earned fishing revenues of less than \$1,000, which would not cover

overall trip expenditures for a year. Sale of pelagic fish contributes to nearly 67% of fishing income, with 20% from bottomfish revenues and the rest from reef fish.

While respondents sold approximately 24% of their total catch, 29% was consumed at home, while 42% was given away. The remaining catch was either released (2%) or exchanged for goods and services (3%). This diversity of catch disposition extends to fishermen who regularly sell fish, as they still retain approximately 30% of their catch for home consumption and participation in traditional fish-sharing networks and customary exchange. Additionally, 78% consider the pelagic fish they catch to be an important source of food, 79% for bottomfish, and 85% for reef fish. These findings validate the importance of fishing in terms of building and maintaining social and community networks, perpetuating fishing traditions, and providing food security to local communities.

Like with CNMI, fishing in Guam is a social activity. Only 7% of fishermen reported fishing alone, and 45% reported that their boat is used without them on occasion. In addition, 61% reported to be a member of a fishing club, association, or group. The majority of fishermen (60%) also agreed that as a fisherman, they are respected by the Guam community. Very few felt that they were not respected by the community.

There was also an open-ended portion of the survey that asked for comments. The two most prevalent themes were that of a rising population and rising fuel costs. Many believed that the expanding population would increase the demand for fish and number of fishermen, yet at the same time, others noted that fuel costs and economic considerations could restrict fishing. In addition, there was concern about the designation of Marianas Trench Marine National Monument, especially since respondents felt that the Marine Preserve Areas established in 1997 had already displaced them from their traditional fishing grounds. Military exercises also affected fishing trips. Other studies have also documented concerns about fishing access related to the designation of the Monument (Richmond and Kotowicz 2015; Kotowicz and Richmond 2013; Kotowicz and Allen 2015). Despite long distance, high cost, and inconvenience, travel to the areas now protected by the Monument were rare but culturally significant events of which fishing was an essential component.

Similar to CNMI, Guam's small boat fisheries are a complex mix of subsistence, cultural, recreational, and quasi-commercial fishermen whose fishing behaviors provide evidence of the importance of fishing to the island of Guam. For nearly all fishery participants, the social and cultural motivations for fishing far outweigh any economic prospects. Nearly all fishermen supplement their income with other jobs and are predominantly subsistence fishermen, selling occasionally to recover trip expenses.

2.4.3.3.1 Guam Bottomfish

Allen and Bartram (2008) reviewed the history of the bottomfish fishery in Guam, which consists of both shallow- and deepwater aspects. They noted that during the 1980s and 1990s, bottomfish fishing was a highly seasonal, small-scale, commercial, subsistence, and recreational fishery. The majority of the participants operated vessels less than 25 ft. long and targeted the shallow-water bottomfish complex because of the lower expenditure and relative ease of fishing close to shore. The commercially-oriented vessels tended to be longer than 25 ft., concentrating effort on the deepwater bottomfish complex. Both deepwater and shallow-water bottomfish are also important target species of the charter fishing fleet, and charter trips accounted for about 15–

20% of all Guam bottomfish fishing trips from 1995 through 2000. In 1998, the charter fleet attracted approximately 3% of visitors to Guam and consisted of a dozen core boats.

Bottomfish was one of the gear types included in the 2011 small boat survey (Hospital and Beavers 2014). Overall fisher demographics and catch disposition were summarized in the previous section. Approximately 57% of respondents reported fishing for deepwater bottomfish and 59% for shallow-water bottomfish, with 52% identifying deepwater bottomfish as their primary target and 49% identifying shallow-water bottomfish as their primary target. Fishers who primarily targeted bottomfish allocated their catch mainly through the Guam Fisherman's Cooperative Association (55%), or to friends, neighbors, and co-workers (41%). For the most part, they self-identified as recreational expense fishers (40%), cultural fishers (35%), subsistence fishers (35%), purely recreational fishers (30%), though respondents spanned all response categories except full-time commercial (i.e., part-time commercial, recreational expense, purely recreational, subsistence, and cultural). Over half of the respondents identified multiple motivations (54%).

2.4.3.3.2 Guam Reef Fish

Coral reef fish were also included in the 2011 small boat survey (Hospital and Beavers 2014). Approximately 33% of respondents reported atulai fishing, 32% spearfishing, and 8% net fishing. Atulai was identified as the primary target by 31%, 20% indicated spearfishing, and 4% indicated net fishing as their primary gear type. Fishers who primarily targeted reef fish sold their catch mainly through the Guam Fisherman's Cooperative Association (37%) or to friends, neighbors, and co-workers (51%). For the most part, respondents self-identified as subsistence fishers (46%), purely recreational fishers (46%), cultural fishers (38.5%), and recreational expense fishers (31%) although respondents spanned all response categories except full-time commercial (i.e., part-time commercial, recreational expense, purely recreational, subsistence, and cultural). Over half of respondents identified multiple motivations (54%).

2.4.3.3.3 Guam Crustaceans

There are currently no socioeconomic data specific to the crustacean fishery. Future reports will include new information as resources allow.

2.4.3.3.4 Guam Precious Corals

There are currently no socioeconomic data specific to the precious coral fishery. Future reports will include new information as resources allow.

2.4.3.4 Guam Fishery Economic Performance

2.4.3.4.1 Guam Bottomfish Commercial Participations, Landings, Revenue, Prices

This section describes trends in commercial pounds sold, revenues and prices, for the Guam bottomfish fishery. Figure 11 presents the trends of commercial pounds sold and revenues of bottomfish fishery during 2012-2022 and Figure 12 presents the trend of total caught versus commercial landings pounds sold during 2012-2022 (for BMUS and boat-base surveys only). Supporting data for Figure 11 and Figure 12 are shown in Table 56. Table 56 also includes the percentage of pounds sold to the total pounds caught of the bottomfish fishery. Both nominal and adjusted values are included in the table. However, only two years (2017 and 2018) of data are presented in the charts and tables, because the data for most of the other years were confidential due to fewer than three participating vendors in the data.

As shown in Figure 11, only two years (2017 and 2018) of commercial landings and revenue are presented during the 20 year-time period, due to data confidentiality considerations for other years (fewer than 3 participating vendors). The total commercial landings and revenue were estimated/expanded based on the sample data provided by dealers who participated the receipt book data collection program. Trends in fish prices were not presented for the same data confidentiality concerns.

Commercial landings were approximately 4,000 pounds, valued at \$17,434 in 2017, and 3,000 pounds valued \$15,290 in 2018. The nominal price for bottomfish in 2018 was \$5.05 per pound and \$4.36 in 2017. Compared to total pounds landed, the commercial landings of BMUS were only small portion. On average, in the two years (2017 and 2018), the pounds sold were only 19% of total estimated pounds caught. Bottomfish prices have been steady in general, but there have been some variations in recent years.

It is worth noting that the data for pounds caught and pounds sold are collected by two different data collection methods. The data of pounds sold were collected through "Commercial Sales Receipt Books" Program, while the data of pounds caught were collected through Boat-based and Shore-based creel surveys. Both data series are generated from an expansion algorithm built on a non-census data collection program, and the survey coverage rates of two data collection methods may change independently across individual years. Therefore, the two time series may not move coherently with each other. For example, the low percentage of pounds sold compared to pounds caught could be due to the low coverage of dealer participation in the Commercial Receipt Books Program.

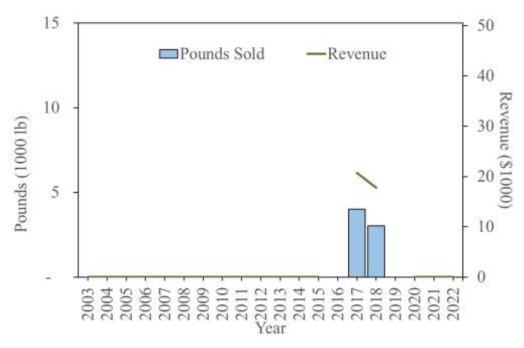


Figure 11. The pounds sold and revenues for the Guam bottomfish fishery (adjusted to 2022 dollars)

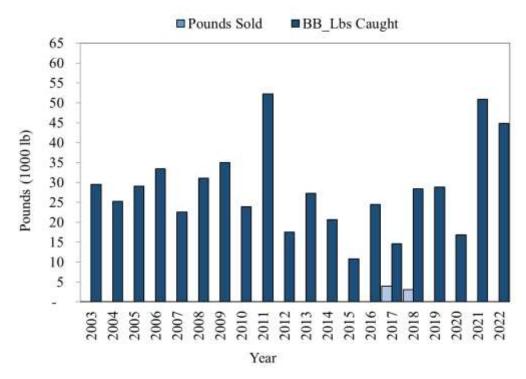


Figure 12. Pounds caught and pounds sold of BMUS for the Guam bottomfish fishery

| | T 4 1 | E 1 | | T 1 | 0/ 6 | | E' 1 ' | |
|------|-------------|-----------|--------------|-------------|--------|------------|------------|----------|
| | Estimated | Estimated | | Estimated | % of | | Fish price | CDI |
| | pounds | pounds | Estimated | revenue (\$ | pounds | Fish | (\$ | CPI |
| Year | caught (lb) | sold (lb) | revenue (\$) | adjusted) | sold | price (\$) | adjusted) | adjustor |
| 2003 | 29,505 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 2.076 |
| 2004 | 25,232 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.957 |
| 2005 | 29,087 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.817 |
| 2006 | 33,413 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.629 |
| 2007 | 22,577 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.526 |
| 2008 | 31,103 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.437 |
| 2009 | 35,029 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.413 |
| 2010 | 23,929 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.373 |
| 2011 | 52,230 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.329 |
| 2012 | 17,517 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.289 |
| 2013 | 27,276 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.289 |
| 2014 | 20,687 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.279 |
| 2015 | 10,783 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.291 |
| 2016 | 24,480 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.219 |
| 2017 | 14,652 | 4,002 | 17,434 | 20,729 | 27% | 4.36 | 5.18 | 1.189 |
| 2018 | 28,365 | 3,028 | 15,290 | 17,721 | 11% | 5.05 | 5.85 | 1.159 |
| 2019 | 28,849 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.139 |
| 2020 | 16,844 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.119 |
| 2021 | 50,865 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.078 |
| 2022 | 44,789 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1 |

Table 56. Commercial landings, revenue, and price information of Guam bottomfish fishery, 2012-2022

2.4.3.4.2 Guam Bottomfish Costs of Fishing

Since 2011, PIFSC economists have maintained a continuous economic data collection program for small boat fishing on Guam through collaboration with PIFSC FRMD (Chan and Pan 2019). The economic data collection gathers fishing expenditure data for boat-based reef fish, bottomfish, and pelagic fishing trips on an ongoing basis. Data for fishing trip expenses include gallons of fuel used, price per gallon of fuel, cost of ice used, cost of bait & chum used, cost of fishing gear lost, and the engine type of the boat. These economic data are collected from same subset of fishing trips as the boat-based creel survey carried out by the local fisheries management agencies and PIFSC. Metadata for these data are available online (PIFSC Socioeconomics Program 2016). Guam trip cost estimates from 2011 for bottomfish fishing trips and other relevant cost information (such as estimates of annual fixed costs) are available in a one-time survey (Hospital and Beavers 2012).

The time series of trip costs of Guam bottomfish fishing presented in Figure 13 are based on a continuous economic data collection program maintained by the PIFSC Socioeconomics Program through collaboration with PIFSC FRMD-WPacFIN. Trip cost data were not available prior to 2011, as this is when the trip data collection program began. The trip cost data for 2011 and 2022 were not presented due to confidentiality considerations (less than 3 trips reported).

^{*} Confidential (fewer than 3 participating vendors). Data source: PIFSC FRMD-WPacFIN.

The fishing costs of bottomfish represent a declining trend from 2012-2016, and then increased substantially in 2017. Trip costs in 2021 went up significantly compared to 2020, due to increasing costs across all cost items. Supporting data for are presented in Table 57. The cost data summaries were generated by excluding outliners (cases with >10 gallons/hours fished).

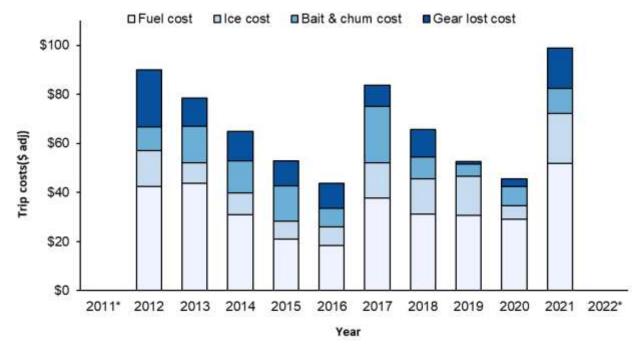


Figure 13. Average trip costs for Guam bottomfish fishing trips (adjusted to 2022 dollars)

Table 57. Average trip costs for Guam bottomfish fishing trips

| | | Total trip | | | | Gear | Fuel price | |
|------|------------|------------|-----------|-----------|-----------|-----------|-------------|--------------|
| | Total trip | cost adj. | Fuel cost | Ice cost | Bait cost | losted | adj. | |
| Year | costs (\$) | (\$) | adj. (\$) | adj. (\$) | adj. (\$) | adj. (\$) | (\$/gallon) | CPI Adjustor |
| 2011 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1.329 |
| 2012 | 70 | 90 | 42 | 15 | 10 | 23 | 5.95 | 1.289 |
| 2013 | 61 | 78 | 44 | 8 | 15 | 11 | 6.14 | 1.289 |
| 2014 | 51 | 65 | 31 | 9 | 13 | 12 | 6.04 | 1.279 |
| 2015 | 41 | 53 | 21 | 7 | 14 | 10 | 5.03 | 1.291 |
| 2016 | 36 | 44 | 18 | 8 | 7 | 10 | 4.15 | 1.219 |
| 2017 | 70 | 84 | 38 | 15 | 23 | 8 | 4.30 | 1.189 |
| 2018 | 57 | 66 | 31 | 14 | 9 | 11 | 4.53 | 1.159 |
| 2019 | 46 | 53 | 31 | 16 | 5 | 1 | 4.37 | 1.139 |
| 2020 | 41 | 46 | 29 | 5 | 8 | 3 | 3.73 | 1.119 |
| 2021 | 92 | 99 | 52 | 20 | 10 | 16 | 4.72 | 1.078 |
| 2022 | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | n.d. | 1 |

Data source: PIFSC Continuous Cost Data Collection Program (Chan and Pan 2019). Trip cost data are not available prior to 2011, as this is when the trip data collection program began.

2.4.3.4.3 Guam Ecosystem Component Species

Based on the new guideline for the archipelagic SAFE report from the Council, this section highlights the top 10 species (sorted by landings) and the priority species (recommended by the local fishery management agency) caught by small boats or shoreline fishing. Please note the top 10 species list and the priority species list reported in the socioeconomic module may not be consistent with the lists reported in the fishery module in the previous sections. The inconsistences result from several factors: 1) differences in data sources, 2) differences in level of species groupings, 3) differences in commercial landing vs. total landings.

Firstly, the data for pounds caught and pounds sold are collected by two different data collection methods, as mentioned in the earlier section. The data for "pounds sold" (commercial landings) reported in this socioeconomics module were collected through "Commercial Sales Receipt Books" Program, while the data of pounds caught were collected through "Boat-based Creel Survey". The survey coverage rates of two data collection methods may change independently in individual years. Secondly, the species groups used in the two data collection programs were different, as the species in the commercial receipt books are usually lumped into family levels or species groups while the species reported in the Creel Survey are often more detailed at the species level. In the case of the top 10 species in Guam, the sum of the top 10 commercial species landings is higher than the sum of the top 10 species landings. Third, fish species with higher total pounds caught may not necessarily lead to higher pounds sold in the markets. Therefore, the two series may not move coherently with each other.

Table 58 shows the commercial landings and revenue of the top 10 ECS in Guam for 2021 and 2022. The total pounds sold and revenue of the top 10 species/species groups in 2022, was lower than 2021. Regarding the nine priority species (species of interests) for the area, only one species showed up with commercial landings in 2022, but these values are not presented in this report due to confidentiality considerations.

Table 58. Top 10 ECS commercial landings, revenue, and price, 2022 and 2022

| Top 10 landings (Boat base) | Pounds kept | Top 10 commercial landings (receipt books) | Pounds sold | Revenue (\$) | Price (\$/lb) |
|-----------------------------|----------------|--|----------------|-----------------|------------------|
| Assorted reef fish | 12,942 | Reef fish | 20,011 | 63,064 | 3.15 |
| Bluespine unicornfish | 5,375 | Bigeye scad (atulai) | 12,218 | 36,753 | 3.01 |
| Orangespine unicornfish | 3,016 | Parrotfish | 5,531 | 19,405 | 3.51 |
| Bigeye scad | 2,059 | Unicornfish | 5,625 | 18,809 | 3.34 |
| Orange-striped emperor | 2,000 | Mafute (emperor) | 3,172 | 9,674 | 3.05 |
| Thumbprint emperor | 1,909 | Octopus | 1,118 | 4,049 | 3.62 |
| Jobfish | 1,735 | Grouper | 954 | 3,352 | 3.51 |
| Bluefin trevally | 1,654 | Jacks | 1,074 | 3,214 | 2.99 |
| Longface emperor | 905 | Rabbitfish (menahac) | 503 | 3,019 | 6.00 |
| 7-11 crab | 786 | Invertebrates | 1,632 | 1,632 | 1.00 |
| Sum 2019 | 32,381 | | 51,838 | 162,971 | 3.14 |
| Sum 2018 | 106,546 | | 94,000 | 279,512 | 2.97 |

^{*} Confidential (fewer than 3 participating vendors).

2.4.4 Ongoing Research and Information Collection

Each year, the PIFSC reports on the status of economic data collections for select regional commercial fisheries. This supports a national economic data monitoring effort known as the Commercial Fishing Economic Assessment Index (CFEAI). Details on the CFEAI and access to data from other regions is available at: https://www.st.nmfs.noaa.gov/data-and-tools/CFEAI-RFEAI/.

The table below represents the most recent data available for CFEAI metrics for select regional commercial fisheries for 2022. Entries for Marianas archipelagic fisheries are bolded in red. These values represent the most recent year of data for key economic data monitoring parameters (fishing revenues, operating costs, and fixed costs). The assessment column indicates the most recent publication year for specific economic assessments (returns above operating cost, profit), where available.

Table 59. Pacific Islands Region 2022 Commercial Fishing Economic Assessment Index

| | 2022 Projected CFEAI | | | | | |
|----------------------------|-------------------------------------|---------------------------------------|-----------------------------------|--|------------------------------------|--|
| | | 2022 Report | ing Year (e.g. 1 | /2022-12/2022) | | |
| | | Data | | Assessn | nent | |
| Pacific Islands Fisheries | Fishing Revenue Most Recent Year | Operating Cost Most Recent Year | Fixed Cost Most Recent Year | Returns Above Operating Costs (Quasi Rent) Assessment Most Recent Year | Profit Assessment Most Recent Year | |
| HI Longline | 2022 | 2022 | 2013 | 2022 | 2016 | |
| ASam Longline | 2022 | 2022 | 2016 | 2022 | 2019 | |
| HI Offshore Handline | 2022 | 2014 | 2014 | 2019 | 2019 | |
| HI Small Boat (pelagic) | 2022 | 2021 | 2021 | 2017 | 2019 | |
| HI Small Boat (bottomfish) | 2022 | 2021 | 2021 | 2017 | 2019 | |
| HI Small Boat (reef) | 2022 | 2021 | 2021 | 2017 | 2019 | |
| Guam Small boat | 2022 | 2022 | 2019 | 2019 | | |
| CNMI Small boat | 2022 | 2022 | 2019 | 2019 | | |
| ASam Small boat | 2022 | 2022 | 2021 | 2019 | | |

PIFSC maintained ongoing economic data collections in the CNMI and Guam for small boat fisheries (Chan and Pan 2019) during 2022.

PIFSC also generates projections for upcoming fiscal years, and the table below provides the projected CFEAI report for 2023 (*all projected activities and analyses are subject to funding*). Based on early projections, PIFSC intends to maintain ongoing economic data collections in the CNMI and Guam for small boat fisheries (Chan and Pan 2019) during 2022.

Table 60. Pacific Islands Region 2023 Commercial Fishing Economic Assessment Index

| | | 2023 CFEAI | | | | | | |
|----------------------------|-------------------------------------|---------------------------------------|-----------------------------------|--|---|--|--|--|
| | | 2023 Report | ing Year (e.g. 1 | /2023-12/2023) | | | | |
| | | Data | | Assessm | nent | | | |
| Pacific Islands Fisheries | Fishing Revenue Most Recent Year | Operating Cost Most Recent Year | Fixed Cost Most Recent Year | Returns Above Operating Costs (Quasi Rent) Assessment Most Recent Year | Profit Assessment Most Recent Year | | | |
| HI Longline | 2023 | 2023 | 2023 | 2023 | 2016 | | | |
| ASam Longline | 2023 | 2023 | 2016 | 2023 | 2019 | | | |
| HI Offshore Handline | 2023 | 2014 | 2014 | 2019 | 2019 | | | |
| HI Small Boat (pelagic) | 2023 | 2021 | 2021 | 2023 | 2023 | | | |
| HI Small Boat (bottomfish) | 2023 | 2021 | 2021 | 2023 | 2023 | | | |
| HI Small Boat (reef) | 2023 | 2021 | 2021 | 2023 | 2023 | | | |
| Guam Small boat | 2023 | 2023 | 2019 | 2019 | | | | |
| CNMI Small boat | 2023 | 2023 | 2019 | 2019 | | | | |
| ASam Small boat | 2023 | 2023 | 2021 | 2019 | | | | |

PIFSC completed a cost-earnings survey of small boat fisheries in Guam and the CNMI during 2018-2019, to serve as an update to the previous 2011 cost-earnings survey (Hospital and Beavers 2012, 2014). This 2018-2019 survey collected data on fishing revenues, operating costs, and fixed costs, as well as numerous elements related to fishing behavior, market participation, and fishery demographics. Efforts to complete the analysis of the 2018-2019 cost-earnings have been delayed due to staff departures coupled with COVID-19 monitoring requirements and PIFSC intends final survey results to be published during 2023.

Community social indicators have been generated for Guam and the CNMI (Kleiber et al. 2018) in accordance with a national project to describe and evaluate community well-being in terms of environmental justice, economic vulnerability, and gentrification pressure (https://www.fisheries.noaa.gov/national/socioeconomics/social-indicators-coastal-communities). However, these indicators rely on Census data, and cannot be updated until 2020 Census data becomes available, likely during 2023.

2.4.5 Relevant PIFSC Economics and Human Dimensions Publications: 2022

| Publication | MSRA priority |
|---|--------------------|
| Ayers A, Leong K, Hospital J, Tam C, Morioka R. 2022. Guam & CNMI fisher observations data summary and analysis. Pacific Islands Fisheries Science Center, PIFSC | HC3.1.1 HC3.1.3 |
| Data Report, DR-22-26, 17 p. https://doi.org/10.25923/wmv2-y197 | HC1.1.7 |
| Freitag A, Blake S, Clay PM, Haynie AC, Kelble C, Jepson M, Kasperski S, Leong KM, Moss JH, Regan SD. 2022. Scale matters - Relating Wetland Loss and Commercial Fishing Activity in Louisiana across Spatial Scales. Nature and Culture, 17(2), 144-169. https://doi.org/10.3167/nc.2022.170202 | HC2.1.2 HC3.1.3 |
| Kleiber D, Iwane M, Kamikawa K, Leong K, Hospital J. 2022. Pacific Islands Region Fisheries and COVID-19: Impacts and adaptations. U.S. Dept. of Commerce, NOAA Technical Memorandum NOAA-TM-NMFS-PIFSC-130, 36 p. https://doi.org/10.25923/2fpm-c128 | HC2.2.4 |
| Smith SL, Cook S, Golden A, Iwane MA, Kleiber D, Leong KM, Mastitski A, Richmond L, Szymkowiak M, Wise S. 2022. Review of adaptations of U.S. commercial fisheries in response to the COVID-19 pandemic using the Resist-Accept-Direct (RAD) framework. Fisheries Management and Ecology. 1-17. https://doi.org/10.1111/fme.12567 | HC2.2.4 |

2.5 PROTECTED SPECIES

This section of the report summarizes information on protected species interactions in fisheries managed under the Mariana FEP. Protected species covered in this report include sea turtles, seabirds, marine mammals, sharks, and corals. Most of these species are protected under the Endangered Species Act (ESA), the Marine Mammal Protection Act (MMPA), and/or the Migratory Bird Treaty Act (MBTA). A list of protected species found in or near Mariana Archipelago waters and a list of critical habitat designations in the Pacific Ocean are included in Appendix B.

2.5.1 Indicators for Monitoring Protected Species Interaction

This report monitors the status of protected species interactions in the Marianas FEP fisheries using proxy indicators such as fishing effort, and changes in gear types as these fisheries do not have observer coverage. Creel surveys and logbook programs are not expected to provide reliable data about protected species interactions. Discussion of protected species interactions is focused on fishing operations in federal waters and associated transit through territorial waters.

2.5.2 FEP Conservation Measures

Bottomfish, precious coral, coral reef and crustacean fisheries managed under this FEP have no specific regulations in place to mitigate protected species interactions. Destructive gear such as bottom trawls, bottom gillnets, explosives, and poisons are prohibited under this FEP, and these prohibitions benefit protected species by preventing potential interactions with non-selective fishing gear.

2.5.2.1 ESA Consultations

ESA consultations were conducted by NMFS and the U.S. Fish and Wildlife Service (USFWS; for species under their jurisdiction) to ensure ongoing fisheries operations managed under the Marianas FEP are not jeopardizing the continued existence of any ESA-listed species or adversely modifying critical habitat. The results of these consultations conducted under section 7 of the ESA are briefly described below and summarized in Table 61.

NMFS concluded in an informal consultation dated April 29, 2015 that all fisheries managed under the Mariana Archipelago FEP are not likely to adversely affect the Indo-West Pacific DPS of scalloped hammerhead shark or ESA-listed reef-building corals.

Table 61. Summary of ESA consultations for Mariana Archipelago FEP Fisheries

| Fishery | Consultation date | Consultation type ^a | Outcome ^b | Species |
|---|-------------------|--------------------------------|----------------------------|---|
| All fisheries | 4/29/2015 | LOC | NLAA | Reef-building corals, scalloped hammerhead shark (Indo-west Pacific DPS) |
| | 3/8/2008 | BiOp | NLAA | Loggerhead sea turtle |
| Bottomfish (CNMI & Guam) | 6/3/2008 | LOC | NLAA | Green sea turtle, olive ridley sea turtle, hawksbill sea turtle, leatherback sea turtle, blue whale, fin whale, humpback whale, sei whale sperm whale |
| Guani) | 8/26/2022 | BiOp | LAA, non- jeopardy | Oceanic whitetip shark |
| | 0,20,2022 | Бтор | NLAA | Sound without state |
| | 3/7/2002 | LOC | NLAA | Loggerhead sea turtle, leatherback sea turtle, olive ridley sea turtle, green sea turtle, hawksbill sea turtle, humpback whale, blue whale, fin whale, sei whale, sperm whale |
| Coral reef ecosystem (CNMI & Guam) | 5/22/2002 | LOC (USFWS) | NLAA | Green, hawksbill, leatherback, loggerhead and olive ridley turtles, Newell's shearwater, short-tailed albatross, Laysan duck, Laysan finch, Nihoa finch, Nihoa millerbird, Micronesian megapode, 6 terrestrial plants |
| | 6/3/2008 | LOC | NLAA | Green sea turtle, olive ridley sea turtle, hawksbill sea turtle, leatherback sea turtle, blue whale, fin whale, humpback whale, sei whale, sperm whale |
| | 9/18/2018 | No effect memo | No effect | Oceanic whitetip shark, giant manta ray |
| Crustaceans (CNMI & Guam) | 9/28/2007 | LOC | NLAA | Green sea turtle, loggerhead sea turtle, olive ridley sea turtle, hawksbill sea turtle, leatherback sea turtle, blue whale, humpback whale, sei whale, sperm whale |
| · | 9/18/2018 | No effect memo | No effect | Oceanic whitetip shark, giant manta ray |
| Precious corals (CNMI & | 10/4/1978 | BiOp | Does not constitute threat | Sperm whale, leatherback sea turtle |
| Guam) | 9/18/2018 | No effect memo | No effect | Oceanic whitetip shark, giant manta ray |

| Fishery | Consultation date | Consultation type ^a | Outcome ^b | Species |
|------------------------------|-------------------|--------------------------------|----------------------|--|
| Precious corals (Guam) | 12/20/2000 | LOC | NLAA | Humpback whale, green sea turtle, hawksbill sea turtle |

^a BiOp = Biological Opinion; LOC = Letter of Concurrence; BE = Biological Evaluation

2.5.2.1.1 Bottomfish Fishery

In a Biological Opinion issued on March 8, 2002, NMFS concluded that the ongoing operation of the Western Pacific Region's bottomfish and seamount fisheries was not likely to jeopardize the continued existence of any threatened or endangered species under NMFS's jurisdiction or destroy or adversely modify any critical habitat. In an informal consultation on June 3, 2008, NMFS concluded that Mariana Archipelago bottomfish fisheries are not likely to adversely affects four sea turtle species (leatherback, olive ridley, green, and hawksbill turtles) and five marine mammal species (humpback, blue, fin, sei, and sperm whales).

On August 26, 2022, NMFS completed a new BiOp that was initiated in response to the ESA listing of the oceanic whitetip shark, chambered nautilus, and giant manta ray. This BiOp did not re-evaluate species previously consulted on because NMFS determined that reinitiation has not been triggered for those species in a Biological Evaluation dated June 5, 2019. NMFS determined that both the Guam and CNMI bottomfish fishery are not likely to adversely affect giant manta rays or chambered nautilus. For oceanic whitetip sharks, NMFS determined that the continued operation of both the Guam and CNMI bottomfish activities would adversely affect the threatened sharks, but determined that the activities are not likely to jeopardize their continued existence. Both bottomfish fisheries incidentally take oceanic whitetip sharks. To monitor the amount of take NMFS established an Incidental Take Statement (ITS) for each fishery as one shark over any five consecutive years for Guam and four sharks over any five consecutive years for CNMI. If the ITS is exceeded, NMFS will reinitiate formal consultation.

2.5.2.1.2 Crustacean Fishery

In an informal consultation completed on September 28, 2007, NMFS concluded that Mariana Archipelago crustacean fisheries are not likely to adversely affect five sea turtle species (loggerhead, leatherback, olive ridley, green, and hawksbill turtles) and five marine mammal species (humpback, blue, fin, sei, and sperm whales).

On September 18, 2018, NMFS concluded that Mariana Archipelago crustacean fisheries will have no effect on the oceanic whitetip shark and giant manta ray.

2.5.2.1.3 Coral Reef Fishery

In an informal consultation completed by NMFS on March 7, 2002, NMFS concluded that fishing activities conducted under the Coral Reef Ecosystems FMP are not likely to adversely affect endangered or threatened species or critical habitat under NMFS's jurisdiction. On May 22, 2002, the USFWS concurred with the determination of NMFS that the activities conducted under the Coral Reef Ecosystems FMP are not likely to adversely affect listed species under USFWS's exclusive jurisdiction (i.e., seabirds) and listed species shared with NMFS (i.e., sea turtles).

^b LAA = likely to adversely affect; NLAA = not likely to adversely affect.

In an informal consultation completed on June 3, 2008, NMFS concluded that the Mariana Archipelago coral reef fisheries are not likely to adversely affect four sea turtle species (leatherback, olive ridley, green, and hawksbill turtles) and five marine mammal species (humpback, blue, fin, sei, and sperm whales).

On September 18, 2018, NMFS concluded that Mariana Archipelago coral reef fisheries will have no effect on the oceanic whitetip shark and giant manta ray.

2.5.2.1.4 Precious Coral Fishery

In a Biological Opinion issued on October 4, 1978, NMFS concluded that the ongoing operation of the Western Pacific Region's precious coral fisheries was not likely to jeopardize the continued existence of any threatened or endangered species under NMFS's jurisdiction or destroy or adversely modify critical habitat. In an informal consultation completed on December 20, 2000, NMFS concluded that Mariana Archipelago precious coral fisheries are not likely to adversely affect humpback whales, green turtles, or hawksbill turtles.

On September 18, 2018, NMFS concluded that Mariana Archipelago precious coral fisheries will have no effect on the oceanic whitetip shark and giant manta ray.

2.5.2.2 Non-ESA Marine Mammals

The MMPA requires NMFS to annually publish a List of Fisheries (LOF) that classifies commercial fisheries in one of three categories based on the level of mortality and serious injury of marine mammals associated with that fishery. According to the 2023 LOF (88 FR 16899, March 21, 2023) the Guam and CNMI bottomfish fisheries operating under the Marianas FEP are classified as Category III fisheries (i.e., a remote likelihood of or no known incidental mortality and serious injury of marine mammals).

2.5.3 Status of Protected Species Interactions in the Marianas FEP Fisheries

2.5.3.1 Bottomfish Fisheries

2.5.3.1.1 Sea Turtle, Marine Mammal, and Seabird Interactions

There are no observer data available for the Guam and CNMI bottomfish fisheries. However, based on current ESA consultations, these fisheries are not expected to interact with any ESA-listed sea turtle, marine mammal, or seabird species in federal waters around Guam or CNMI. NMFS has also concluded that the Mariana Archipelago commercial bottomfish fisheries will not affect marine mammals in any manner not considered or authorized under the MMPA.

Based on fishing effort and other characteristics described in Chapter 1 of this report, no notable changes have been observed in the fishery. There is no other information to indicate that impacts to sea turtle, marine mammal, or seabird species from this fishery have changed in recent years.

2.5.3.1.2 Elasmobranch Interactions

As indicated in Section 2.5.2.1, ESA consultation for newly listed elasmobranch species was completed in 2022. To meet the requirements of the new BiOp for the Marianas bottomfish fisheries, ITS for oceanic whitetip sharks will be monitored on an annual basis to serve as a check for the reinitiation trigger. Available information on elasmobranch interactions in the Guam and CNMI bottomfish fishery are included here.

There is limited data on fishery interactions with oceanic white tip sharks in Pacific Island bottomfish fisheries. Where data exists, some datasets identified oceanic whitetip shark captures to the species level, while others categorized oceanic whitetip sharks and whitetip reef sharks as "whitetip shark." Guam and CNMI bottomfish boat-based creel surveys indicate that fishermen catch whitetip reef sharks more frequently than oceanic whitetip sharks.

From 1982 to 2017, Guam DAWR recorded 39 whitetip reef sharks and 3 oceanic whitetip sharks in the Guam boat-based creel survey (NMFS 2019). No additional interactions with oceanic whitetip sharks have been reported since 2013 for the Guam bottomfish fishery.

There have been no records of oceanic whitetip sharks in the CNMI boat-based creel surveys administered by CNMI DFW since the start of the dataset in 2000. The federal commercial bottomfish logbook form in the CNMI has a write-in space for recording catch by species under the shark category. Between 2009, when logbooks were implemented, and 2017, fishermen recorded 33 sharks as "whitetip shark", which may be whitetip reef sharks or oceanic whitetip sharks. Based on catch composition associated with the whitetip shark captures, most records were associated with shallow-water fish species captures, which are more likely to be whitetip reef sharks. Twelve of the 33 whitetip shark captures were associated with deep-water bottomfish species, which could potentially be oceanic whitetip sharks (NMFS 2019). No additional interactions with oceanic whitetip sharks or unidentified whitetip sharks have been reported in the last five years in the CNMI bottomfish fishery.

Table 62. The number of oceanic whitetip shark interactions expected, including unidentified sharks, as calculated by the 2022 BiOp, representing the ITS, with the reported number of interactions based on the best scientific data as described above.

| Territory | ITS | Reported number in the last five consecutive calendar years |
|-----------|-----|---|
| Guam | 1 | 0 |
| CNMI | 4 | 0 |

While bottomfish fishing surveys in the main Hawaiian Islands (PIFSC unpublished survey) and Guam (Kendall Enterprise Inc. 2014) show records of whitetip reef shark captures, there have not been any oceanic whitetip sharks recorded in bottomfish surveys or other PIFSC research activities. In addition to the bottomfish surveys, PIFSC researchers have conducted limited bottomfish fishing in the Pacific Islands region for life history research purposes since 2007. They typically fish once to twice a year and land a maximum of 1,200 kg of bottomfish each time they fish. In the last five years (2013-2018), there was one trip each to Johnston Atoll, the CNMI, Guam, and American Samoa, and Samoa. There are no records of researchers catching oceanic whitetip sharks while conducting these activities. There was one record in Guam of an oceanic whitetip shark depredating hooked fish but did not become hooked or entangled on the line (NMFS 2019).

2.5.3.2 Coral Reef Fisheries

There are no observer data available for the Guam and CNMI coral reef fisheries. However, based on current ESA consultations, these fisheries are not expected to interact with any ESA-listed species in federal waters around Guam or CNMI. NMFS has also concluded that the

Mariana Archipelago commercial coral reef fisheries will not affect marine mammals in any manner not considered or authorized under the MMPA.

Based on fishing effort and other characteristics described in Chapter 1 of this report, no notable changes have been observed in the fishery. There is no other information to indicate that impacts to protected species from this fishery have changed in recent years.

2.5.3.3 Crustacean and Precious Coral Fisheries

There are currently no crustacean or precious coral fisheries operating in federal waters around Guam or CNMI. However, based on current ESA consultations, crustacean fisheries are not expected to interact with any ESA-listed species in federal waters around Guam or CNMI. NMFS has also concluded that the Mariana Archipelago crustacean and precious coral commercial fisheries will not affect marine mammals in any manner not considered or authorized under the MMPA.

2.5.4 Identification of Emerging Issues

Table 63 summarizes current candidate ESA species, recent listing status, and post-listing activity (critical habitat designation and recovery plan development). Impacts from FEP-managed fisheries on any new listings and critical habitat designations will be considered in future versions of this report.

Table 63. Status of candidate ESA species, recent ESA listing processes, and post-listing activities

| Species | | | Listing Process | , | Post-Listi | ng Activity |
|------------------------------|--|---|--|---|---|---|
| Common Name | Scientific Name | 90-Day Finding | 12-Month Finding / Proposed Rule | Final Rule | Critical Habitat | Recovery Plan |
| Oceanic whitetip shark | Carcharhinus longimanus | Positive (81 FR 1376, 1/12/2016) | Positive, threatened (81 FR 96304, 12/29/2016) | Listed as threatened (83 FR 4153, 1/30/18) | Designation not prudent; no areas within U.S. jurisdiction that meet definition of critical habitat (85 FR 12898, 3/5/2020) | Draft Recovery Plan published January 25, 2023 (88 FR 4817) |
| Giant manta ray | Manta birostris | Positive (81 FR 8874, 2/23/2016) | Positive, threatened (82 FRN 3694, 1/12/2017) | Listed as threatened (83 FR 2916, 1/22/18) | Designation not prudent; no areas within U.S. jurisdiction that meet definition of critical habitat (84 FR 66652, 12/5/2019) | Recovery outline published 12/4/19 to serve as interim guidance until full recovery plan is developed; recovery planning workshop planned for 2021. |
| Corals | N/A | Positive for 82 species (75 FR 6616, 2/10/2010) | Positive for 66 species (77 FR 73219, 12/7/2012) | 20 species listed as threatened (79 FR 53851, 9/10/2014) | Critical habitat proposed (85 FR 76262, 11/27/2021), comment period extended through 5/26/2021 (86 FR 16325) | In development, interim recovery outline in place; recovery workshops convened in May 2021. |
| Giant Clams | Hippopus, H. porcellanus, Tridacna costata, T. derasa, T. gigas, T. Squamosa, and T. tevoroa | Positive (82 FR 28946, 06/26/2017) | TBA (status review ongoing) | TBA | N/A | N/A |

| Sp | ecies | | Listing Process | | Post-Listi | ng Activity |
|------------------------|---------------------------|---|---|--|--|--|
| Common Name | Scientific Name | 90-Day Finding | 12-Month Finding / Proposed Rule | Final Rule | Critical Habitat | Recovery Plan |
| Green sea turtle | Chelonia mydas | Positive (77 FR 45571, 8/1/2012) | Identification of 11 DPSs, endangered and threatened (80 FR 15271, 3/23/2015) | 11 DPSs listed as endangered and threatened (81 FR 20057, 4/6/2016) | In development, proposal expected summer 2023 | TBA |
| Humpback whale | Megaptera novaeangliae | Positive 90-day finding on petition to classify the North Pacific population as DPS and delist the DPS (78 FR 53391, 8/29/2013) | Revision of species-wide listing and listing of four DPSs as threatened or endangered (80 FR 22304) | Revision of species wide listing; Western North Pacific DPS listed as endangered (81 FR 62259, 9/8/2016) | No critical habitat designated for waters around the Mariana Archipelago (86 FR 21082, 4/21/21) | In development for Western North Pacific DPS; anticipated publication of draft documents & public comment period in 2023 |
| Shortfin Mako Shark | Isurus oxyrinchus | Positive (86 FR 19863, 04/15/2021 | Not warranted (87 FR 68236, 11/14/2022) | N/A | N/A | N/A |

2.5.5 Identification of Research, Data, and Assessment Needs

The following research, data, and assessment needs for insular fisheries were identified by the Council's Plan Team:

- Improve species identification of commercial and non-commercial fisheries data (e.g., outreach, use FAO species codes) to improve understanding of potential protected species impacts.
- Define and evaluate innovative approaches to derive robust estimates of protected species interactions in insular fisheries.
- Conduct genetic and telemetry research to improve understanding of population structure and movement patterns for listed elasmobranchs.
- Estimates of post release survival for incidental protected species.

2.6 CLIMATE AND OCEANIC INDICATORS

2.6.1 Introduction

Over the past several years, the Council has incorporated climate change into the overall management of the fisheries over which it has jurisdiction. This 2022 annual SAFE report includes a now standard chapter on indicators of climate and oceanic conditions in the Western Pacific region. These indicators reflect global climate variability and change as well as trends in local oceanographic conditions.

The reasons for the Council's decision to provide and maintain an evolving discussion of climate conditions as an integral and continuous consideration in their deliberations, decisions, and reports are numerous:

- Emerging scientific and community understanding of the impacts of changing climate conditions on fishery resources, the ecosystems that sustain those resources, and the communities that depend upon them;
- Recent Federal Directives including the 2010 implementation of a National Ocean
 Policy that identified Resiliency and Adaptation to Climate Change and Ocean
 Acidification as one of nine National priorities as well as the development of a Climate
 Science Strategy by NMFS in 2015 and the subsequent development of the Pacific
 Islands Regional Action Plan for climate science; and
- The Council's own engagement with NOAA as well as jurisdictional fishery management agencies in American Samoa, CNMI, Guam, and Hawaii as well as fishing industry representatives and local communities in those jurisdictions.

In 2013, the Council began restructuring its Marine Protected Area/Coastal and Marine Spatial Planning Committee to include a focus on climate change, and the committee was renamed as the Marine Planning and Climate Change Committee (MPCCC). In 2015, based on recommendations from the committee, the Council adopted its Marine Planning and Climate Change Policy and Action Plan, which provided guidance to the Council on implementing climate change measures, including climate change research and data needs. The revised Pelagic FEP (February 2016) included a discussion on climate change data and research as well as a new objective (Objective 9) that states the Council should consider the implications of climate change in decision-making, with the following sub-objectives:

- To identify and prioritize research that examines the effects of climate change on Council-managed fisheries and fishing communities.
- To ensure climate change considerations are incorporated into the analysis of management alternatives.
- To monitor climate change related variables via the Council's Annual Reports.
- To engage in climate change outreach with U.S. Pacific Islands communities.

Beginning with the 2015 report, the Council and its partners began providing continuing descriptions of changes in a series of climate and oceanic indicators. However, the MPCCC was disbanded in early 2019, re-allocating its responsibilities among its members already on other committees or teams, such as the Fishery Ecosystem Plan Teams.

This annual report focuses previous years' efforts by refining existing indicators and improving communication of their relevance and status. Future reports will include additional indicators as the information becomes available and their relevance to the development, evaluation, and

revision of the FEPs becomes clearer. Working with national and jurisdictional partners, the Council will make all datasets used in the preparation of this and future reports available and easily accessible.

2.6.2 Response to Previous Plan Team and Council Recommendations

There were no Council recommendations relevant to the climate and oceanic indicators section of the annual SAFE report for the Mariana Archipelago in 2022.

2.6.3 Conceptual Model

In developing this chapter, the Council relied on a number of recent reports conducted in the context of the U.S. National Climate Assessment including, most notably, the 2012 Pacific Islands Regional Climate Assessment (PIRCA) and the Ocean and Coasts chapter of the 2014 report on a Pilot Indicator System prepared by the National Climate Assessment and Development Advisory Committee (NCADAC).

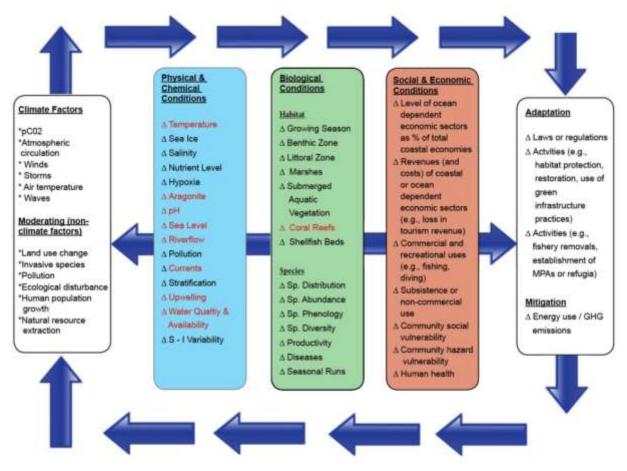
The Advisory Committee Report presented a possible conceptual framework designed to illustrate how climate factors can connect to and interact with other ecosystem components to impact ocean and coastal ecosystems and human communities. The Council adapted this model with considerations relevant to the fishery resources of the Western Pacific region (Figure 14).

As described in the 2014 NCADAC report, the Conceptual Model presents a "simplified representation of climate and non-climate stressors in coastal and marine ecosystems." For the purposes of this Annual Report, the modified Conceptual Model allows the Council and its partners to identify indicators of interest to be monitored on a continuing basis in coming years. The indicators shown in red were considered for inclusion in the annual SAFE reports, though the final list of indicators varied somewhat. Other indicators will be added over time as data become available and an understanding of the causal chain from stressors to impacts emerges.

The Council also hopes that this Conceptual Model can provide a guide for future monitoring and research. This guide will ideally enable the Council and its partners to move forward from observations and correlations to understanding the specific nature of interactions, and to develop capabilities to predict future changes of importance in the developing, evaluating, and adapting of FEPs in the Western Pacific region.

Indicators of Change to Archipelagic Coastal and Marine Systems*

(Items in red to be monitored for 2015 Annual Reports of the Archipelagic Fishery Ecosystem Plans for the Western Pacific Region)



*Adapted from National Climate Assessment and Development Advisory Committee. February 2014. National Climate Indicators System Report. B-59.

Figure 14. Indicators of change of archipelagic coastal and marine systems; conceptual model

2.6.4 Selected Indicators

The primary goal for selecting the indicators used in this (and future reports) is to provide fisheries-related communities, resource managers, and businesses with climate-related situational awareness. In this context, Indicators were selected to:

- Be fisheries relevant and informative;
- Build intuition about current conditions in light of changing climate;
- Provide historical context; and
- Recognize patterns and trends.

In this context, this section includes the following climate and oceanic indicators:

- Atmospheric concentration of carbon dioxide (CO₂)
- Oceanic pH at Station ALOHA;
- Oceanic Niño Index (ONI);

- Pacific Decadal Oscillation (PDO);
- Tropical cyclones;
- Sea surface temperature (SST);
- Coral Thermal Stress Exposure
- Chlorophyll-A (Chl-A)
- Rainfall
- Sea Level (Sea Surface Height)

Figure 15 and Figure 16 provide a description of these indicators and illustrate how they are connected to each other in terms of natural climate variability and anthropogenic climate change.

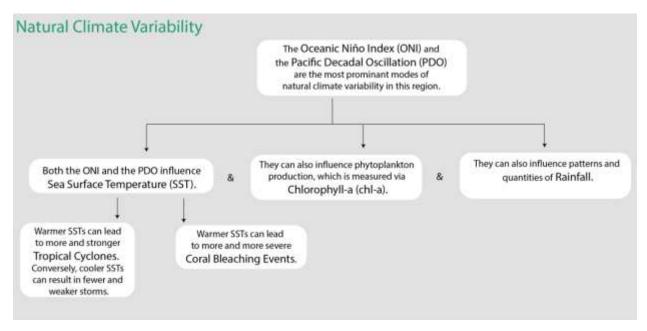


Figure 15. Schematic diagram illustrating how indicators are connected to one another and how they vary as a result of natural climate variability

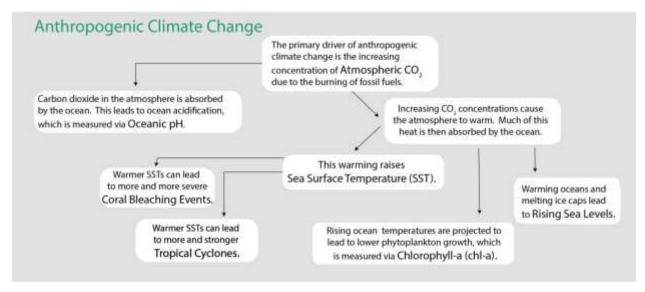


Figure 16. Schematic diagram illustrating how indicators are connected to one another and how they vary as a result of anthropogenic climate change

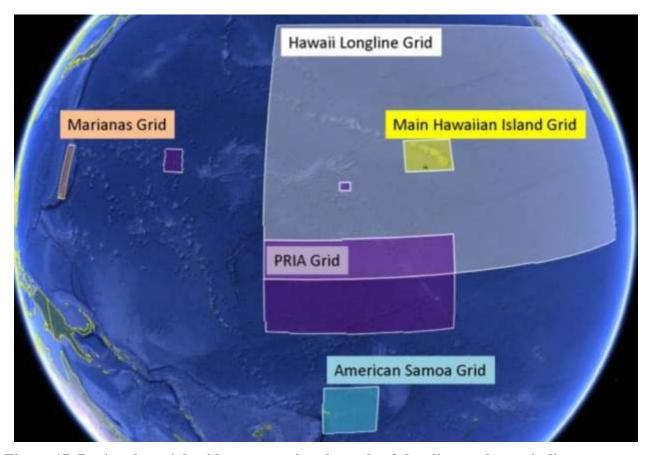


Figure 17. Regional spatial grids representing the scale of the climate change indicators being monitored

2.6.4.1 Atmospheric Concentration of Carbon Dioxide at Mauna Loa

Rationale: Atmospheric carbon dioxide is a measure of what human activity has already done to affect the climate system through greenhouse gas emissions. It provides quantitative information in a simplified, standardized format that decision makers can easily understand. This indicator demonstrates that the concentration (and, in turn, warming influence) of greenhouse gases in the atmosphere has increased substantially over the last several decades.

Status: Atmospheric CO_2 is increasing exponentially. This means that atmospheric CO_2 is increasing more quickly over time. In 2022, the annual mean concentration of CO_2 was 418.56 ppm. This is the highest annual value recorded. This year also saw the highest monthly value, which was 420.99 ppm. In 1959, the first year full of the time series, the atmospheric concentration of CO_2 was 316 ppm. The annual mean passed 350 ppm in 1988 and 400 ppm in 2015.

Description: Monthly mean atmospheric carbon dioxide (CO₂) at Mauna Loa Observatory, Hawaii in parts per million (ppm) from March 1958 to present. The observed increase in monthly average carbon dioxide concentration is primarily due to CO₂ emissions from fossil fuel burning. Carbon dioxide remains in the atmosphere for a very long time, and emissions from any location mix throughout the atmosphere in approximately one year. The annual variations at Mauna Loa, Hawai'i are due to the seasonal imbalance between the photosynthesis and respiration of terrestrial plants. During the summer growing season, photosynthesis exceeds respiration, and CO₂ is removed from the atmosphere. In the winter (outside the growing season), respiration exceeds photosynthesis, and CO₂ is returned to the atmosphere. The seasonal cycle is strongest in the northern hemisphere because of its larger land mass.

Timeframe: Annual, monthly.

Region/Location: Mauna Loa, Hawaii, but representative of global atmospheric carbon dioxide concentration.

Measurement Platform: In-situ station.

Data available at: https://gml.noaa.gov/ccgg/trends/data.html.

Sourced from: Keeling et al. (1976), Thoning et al. (1989), and NOAA (2023a). Graphics produced in part using Stawitz (2022).

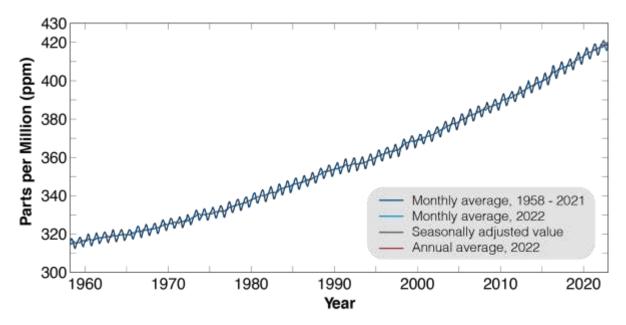


Figure 18. Monthly mean (black) and seasonally corrected (blue) atmospheric carbon dioxide at Mauna Loa Observatory, Hawaii

2.6.4.2 Oceanic pH

Rationale: Oceanic pH is a measure of how greenhouse gas emissions have already impacted the ocean. This indicator demonstrates that oceanic pH has decreased significantly over the past several decades (i.e., the ocean has become more acidic). Increasing ocean acidification limits the ability of marine organisms to build shells and other calcareous structures. Recent research has shown that pelagic organisms such as pteropods and other prey for commercially valuable fish species are already being negatively impacted by increasing acidification (Feely et al. 2016). The full impact of ocean acidification on the pelagic food web is an area of active research (Fabry et al. 2008).

Status: The ocean is roughly 10.9% more acidic than it was 30 years ago at the start of this time series. Over this time, pH has declined by 0.045 at a constant rate. In 2021, the most recent year for which data are available, the average pH was 8.05. Additionally, for the 6th year, small variations seen over the course of the year are outside the range seen in the first year of the time series. The highest pH value reported for the most recent year (8.069) is lower than the lowest pH value reported in the first year of the time series (8.083).

Description: Trends in surface (5 m) pH at Station ALOHA, north of Oahu (22.75°N, 158°W), collected by the Hawai'i Ocean Time Series (HOT) from October 1988 to 2021 (2022 data are not yet available). Oceanic pH is a measure of ocean acidity, which increases as the ocean absorbs carbon dioxide from the atmosphere. Lower pH values represent greater acidity. Oceanic pH is calculated from total alkalinity (TA) and dissolved inorganic carbon (DIC). Total alkalinity represents the ocean's capacity to resist acidification as it absorbs CO₂ and the amount of CO₂ absorbed is captured through measurements of DIC. The multi-decadal time series at Station ALOHA represents the best available documentation of the significant downward trend in oceanic pH since the time series began in 1988. Oceanic pH varies over both time and space, though the conditions at Station ALOHA are considered broadly representative of those across the Western and Central Pacific's pelagic fishing grounds.

Timeframe: Monthly.

Region/Location: Station ALOHA: 22.75°N, 158°W.

Measurement Platform: *In-situ* station.

Data available at: https://hahana.soest.hawaii.edu/hot/hot-dogs/bseries.html.

Sourced from: Fabry et al. (2008), Feely et al. (2016), and the Hawai'i Ocean Time Series as described in Karl and Lukas (1996) and on its website (HOT 2023) using the methodology provided by Zeebe and Wolf-Gladrow (2001). Graphics produced in part using Stawitz (2022).

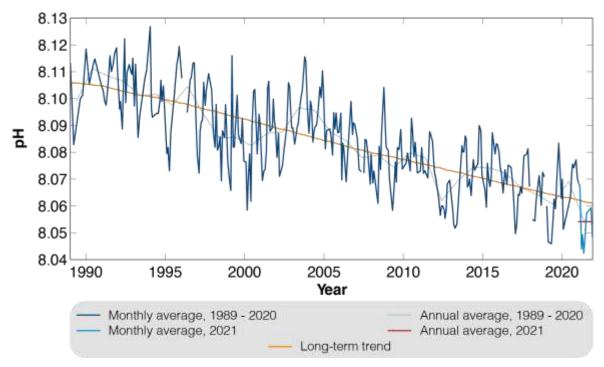


Figure 19. Time series and long-term trend of oceanic pH measured at Station ALOHA from 1989-2021

2.6.4.3 Oceanic Niño Index

Rationale: The El Niño – Southern Oscillation (ENSO) cycle is known to have impacts on Pacific fisheries including tuna fisheries. The ONI focuses on ocean temperature, which has the most direct effect on these fisheries.

Status: The ONI indicated La Niña conditions throughout 2022. In 2022, the ONI ranged from -1.06 to -0.81. This is within the range of values observed previously in the time series.

Description: The three-month running mean (referred to as a season) of satellite remotely-sensed sea surface temperature (SST) anomalies in the Niño 3.4 region ($5^{\circ}S - 5^{\circ}N$, $120^{\circ} - 170^{\circ}W$). The Oceanic Niño Index (ONI) is a measure of the El Niño – Southern Oscillation (ENSO) phase. Warm and cool phases, termed El Niño and La Niña respectively, are based in part on an ONI threshold of \pm 0.5 °C being met for a minimum of five consecutive overlapping seasons. Additional atmospheric indices are needed to confirm an El Niño or La Niña event, as the ENSO is a coupled ocean-atmosphere phenomenon. The atmospheric half of ENSO is measured using the Southern Oscillation Index.

Timeframe: Every three months.

Region/Location: Niño 3.4 region, 5°S – 5°N, 120° – 170°W.

Measurement Platform: In-situ station, satellite, model.

Data available at: https://www.cpc.ncep.noaa.gov/data/indices/oni.ascii.txt.

Sourced from NOAA CPC (2023). Graphics produced in part using Stawitz (2022).

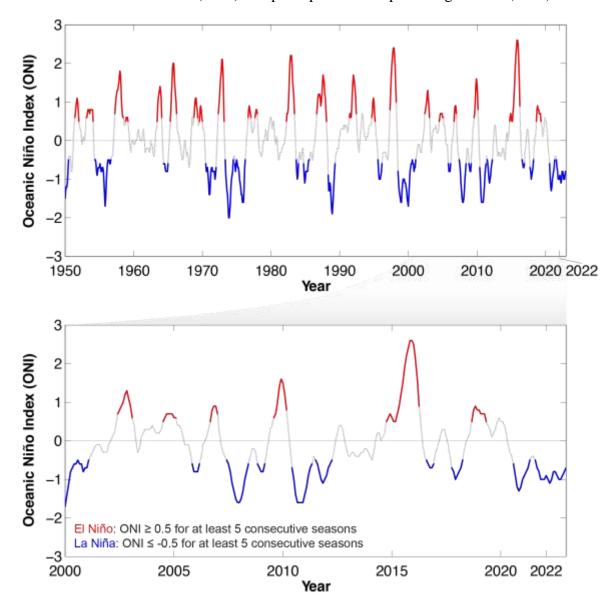


Figure 20. Oceanic Niño Index from 1950-2022 (top) and 2000-2022 (bottom) with El Niño periods in red and La Niña periods in blue

2.6.4.4 Pacific Decadal Oscillation

Rationale: The Pacific Decadal Oscillation (PDO) was initially named by fisheries scientist Steven Hare in 1996 while researching connections between Alaska salmon production cycles and Pacific climate. Like ENSO, the PDO reflects changes between periods of persistently warm or persistently cool ocean temperatures, but over a period of 20 to 30 years (versus six to 18 months for ENSO events). The climatic fingerprints of the PDO are most visible in the Northeastern Pacific, but secondary signatures exist in the tropics.

Status: The PDO was negative in 2022. The index ranged from -2.22 to -1.35 over the course of the year. This is within the range of values observed previously in the time series.

Description: The PDO is often described as a long-lived El Niño-like pattern of Pacific climate variability. As seen with the better-known ENSO, extremes in the PDO pattern are marked by widespread variations in the Pacific Basin and the North American climate. In parallel with the ENSO phenomenon, the extreme cases of the PDO have been classified as either warm or cool, as defined by ocean temperature anomalies in the northeast and tropical Pacific Ocean. When SST is below average in the [central] North Pacific and warm along the North American coast, and when sea level pressures are below average in the North Pacific, the PDO has a positive value. When the climate patterns are reversed, with warm SST anomalies in the interior and cool SST anomalies along the North American coast, or above average sea level pressures over the North Pacific, the PDO has a negative value. Description inserted from NOAA (2023b).

Timeframe: Annual, monthly.

Region/Location: Pacific Basin north of 20°N.

Measurement Platform: *In-situ* station, satellite, model.

Data available at: https://psl.noaa.gov/pdo/.

Sourced from: NOAA (2023b), Mantua (1997), and Newman (2016). Graphics produced in part

using Stawitz (2022).

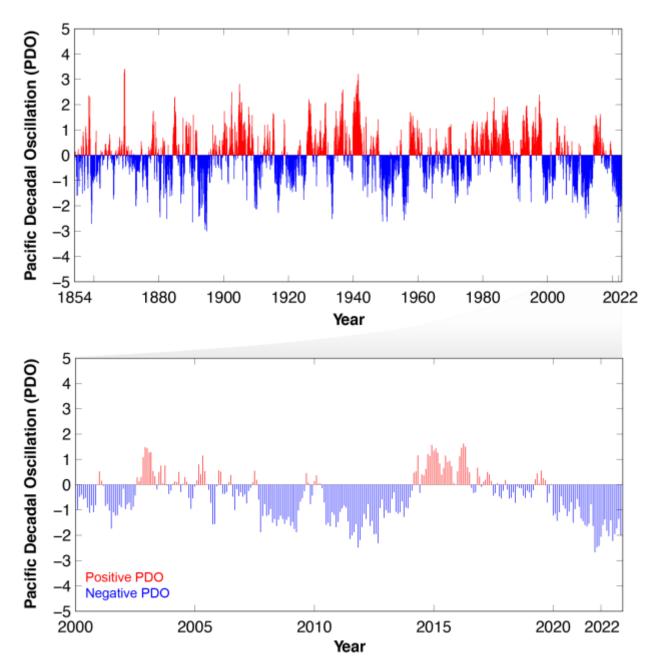


Figure 21. Pacific Decadal Oscillation from 1950-2022 (top) and 2000-2022 (bottom) with positive warm periods in red and negative cool periods in blue

2.6.4.5 Tropical Cyclones

Rationale: The effects of tropical cyclones are numerous and well known. At sea, storms disrupt and endanger shipping traffic as well as fishing effort and safety. The Hawai'i longline fishery, for example, has had serious problems with vessels dodging storms at sea, delayed departures, and inability to make it safely back to Honolulu because of bad weather. When cyclones encounter land, their intense rains and high winds can cause severe property damage, loss of life, soil erosion, and flooding. Associated storm surge, the large volume of ocean water pushed toward shore by cyclones' strong winds, can cause severe flooding and destruction.

Status:

Eastern North Pacific. Tropical cyclone activity was near normal in the Eastern Pacific in 2022. There were 19 named storms, 10 of which were hurricanes. There were 4 major hurricanes (category 3 or higher), which is also near normal. The Accumulated Cyclone Energy (ACE) was near the 1991–2020 average. After four straight years of named storms forming in the Eastern Pacific in November (which is unusually high), conditions returned to normal this November with no storms, named or otherwise. Portions of this summary inserted from https://www.ncei.noaa.gov/access/monitoring/monthly-report/tropical-cyclones/202211.

Central North Pacific. Central Pacific tropical cyclone activity was below the 1991–2020 average in 2022. There was 1 named storm, which reached hurricane status, and no major hurricanes. A weakened Hurricane Darby entered the Central Pacific in July, passing south of the Island of Hawai'i as a tropical depression. On average (1991–2020), the central Pacific sees four named storms, two hurricanes, and one major hurricane each year. The 2022 ACE index was about ten percent of the 1991–2020 average. Portions of this summary inserted from https://www.ncei.noaa.gov/access/monitoring/monthly-report/tropical-cyclones/202207.

Western North Pacific. Tropical cyclone activity got off to a slow but strong start in the Western Pacific, with no storms occurring until Super Typhoon Malakas formed in April. The season overall saw below normal activity for the third year in a row. Tropical cyclone activity was below the 1991–2020 average in 2022. The 22 named storms, 12 typhoons, and 5 major typhoons were all below average (1991–2020), as was the ACE. Portions of the summary inserted from https://www.ncei.noaa.gov/access/monitoring/monthly-report/tropical-cyclones/202204, and https://www.ncei.noaa.gov/access/monitoring/monthly-report/tropical-cyclones/202213.

South Pacific. South Pacific tropical cyclone activity was below average in 2022. There were 4 named storms, none of which became cyclones or major cyclones. The 2022 ACE was also below the 1991–2020 average. Portions of the summary inserted from https://www.ncei.noaa.gov/access/monitoring/monthly-report/tropical-cyclones/202213.

Description: This indicator uses historical data from the NOAA National Climate Data Center (NCDC) International Best Track Archive for Climate Stewardship to track the number of tropical cyclones in the western, central, eastern, and southern Pacific basins. This indicator also monitors the Accumulated Cyclone Energy (ACE) Index and the Power Dissipation Index which are two ways of monitoring the frequency, strength, and duration of tropical cyclones based on wind speed measurements.

The annual frequency of storms passing through each basin is tracked and Figure 22 shows the representative breakdown of Saffir-Simpson hurricane categories.

Every cyclone has an ACE Index value, which is a number based on the maximum wind speed measured at six-hourly intervals over the entire time that the cyclone is classified as at least a tropical storm (wind speed of at least 34 knots; 39 mph). Therefore, a storm's ACE Index value accounts for both strength and duration. Figure 23 shows the ACE values for each hurricane/typhoon season and has a horizontal line representing the average annual ACE value.

Timeframe: Annual.

Region/Location:

Eastern North Pacific: east of 140° W, north of the equator. Central North Pacific: 180° - 140° W, north of the equator. Western North Pacific: west of 180°, north of the equator.

South Pacific: south of the equator.

Measurement Platform: Satellite.

Data available at: https://www.ncei.noaa.gov/data/international-best-track-archive-for-climate-stewardship-ibtracs/v04r00/access/csv.

Sourced from: Knapp et al. (2010), Knapp et al. (2018), and NOAA (2023c).

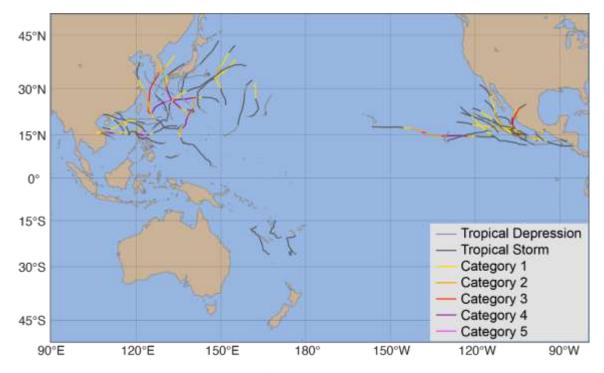


Figure 22. 2022 Pacific basin tropical cyclone tracks

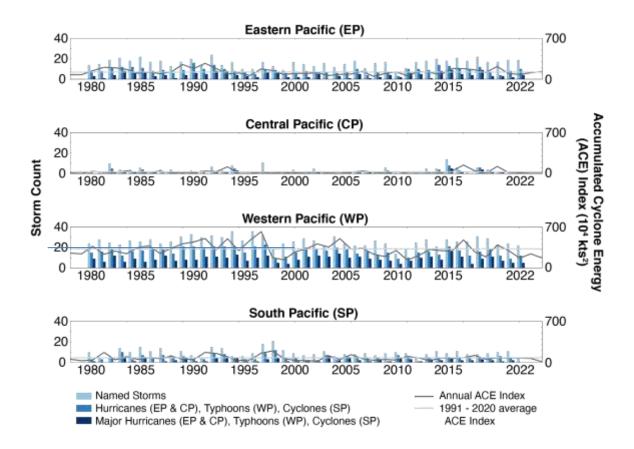


Figure 23. Storm counts (bars) and Accumulated Cyclone Energy (ACE) index values (lines) in each region of the Pacific. Both annual ACE index (black lines) and 1991–2020 average ACE index (grey lines) are shown

2.6.4.6 Sea Surface Temperature and Anomaly

Rationale: Sea surface temperature (SST) is one of the most directly observable existing measures for tracking increasing ocean temperatures. SST varies in response to natural climate cycles such as ENSO and is projected to rise as a result of anthropogenic climate change. Both short-term variability and long-term trends in SST impact the marine ecosystem. Understanding the mechanisms through which organisms are impacted and the time scales of these impacts is an area of active research.

Status: Annual mean SST was 28.8° C in 2022. Over the period of record, annual SST has increased at a rate of 0.0247° C/year. Monthly SST values in 2022 ranged from $27.12 - 30.03^{\circ}$ C, within the climatological range of $25.60 - 30.60^{\circ}$ C. The annual anomaly was 0.43° C hotter than average, with intensification in the northern islands.

Note that from the top to bottom in Figure 24, panels show climatological SST (1985–2021), 2022 SST anomaly, time series of monthly mean SST, and time series of monthly SST anomaly.

Description: Satellite remotely-sensed monthly sea surface temperature (SST) is averaged across the Marianas Grid ($13^{\circ} - 21^{\circ}N$, $144^{\circ} - 146^{\circ}E$). A time series of monthly mean SST averaged over the Marianas Grid Region is presented. Additionally, spatial climatology and anomalies are shown.

Timeframe: Monthly.

Region/Location: Marianas Grid (13° – 21°N, 144° – 146°E).

Measurement Platform: Satellite.

Source: NOAA OceanWatch (2023a).

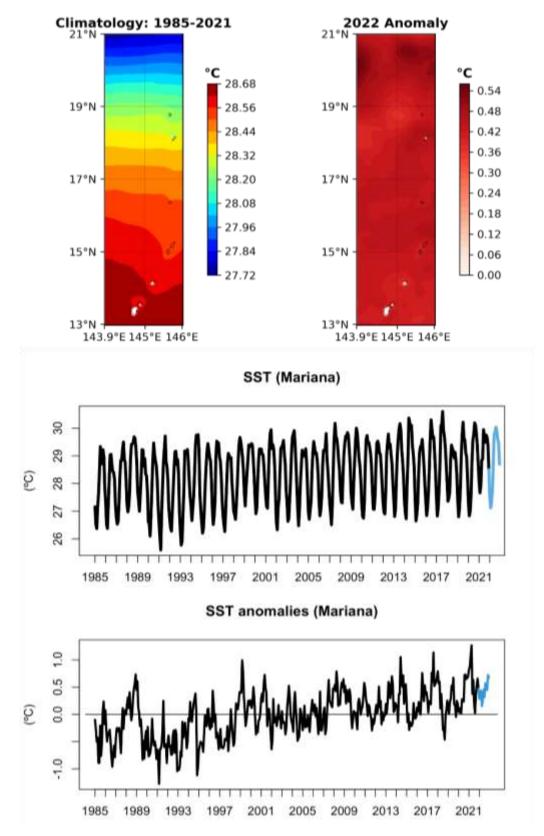


Figure 24. Sea surface temperature climatology and anomalies from 1985–2022

2.6.4.7 Coral Thermal Stress Exposure: Degree Heating Weeks

Rationale: Degree heating weeks are one of the most widely used metrics for assessing exposure to coral bleaching-relevant thermal stress.

Status: After a series of stress events in 2013, 2014, 2016, 2017, 2019, and 2020, the Marianas experienced another coral heat stress event in the second half of 2022 with mass bleaching expected.

Description: Here we present a metric of exposure to thermal stress that is relevant to coral bleaching. Degree Heating Weeks (DHW) measure time and temperature above a reference 'summer maximum', presented as rolling sum weekly thermal anomalies over a 12-week period. Higher DHW measures imply a greater likelihood of mass coral bleaching or mortality from thermal stress.

The NOAA Coral Reef Watch program uses satellite data to provide current reef environmental conditions to quickly identify areas at risk for <u>coral bleaching</u>. Bleaching is the process by which corals lose the symbiotic algae that give them their distinctive colors. If a coral is severely bleached, disease and death become likely.

The NOAA Coral Reef Watch (CRW) daily 5-km satellite coral bleaching Degree Heating Week (DHW) product presented here shows accumulated heat stress, which can lead to coral bleaching and death. The scale goes from 0 to 20 °C-weeks. The DHW product accumulates the instantaneous bleaching heat stress (measured by Coral Bleaching HotSpots) during the most-recent 12-week period. It is directly related to the timing and intensity of coral bleaching. Significant coral bleaching usually occurs when DHW values reach 4 °C-weeks. By the time DHW values reach 8 °C-weeks, widespread bleaching is likely and significant mortality can be expected (NOAA Coral Reef Watch 2023).

Timeframe: 2014–2022, daily data.

Region/Location: Global.

Sourced from: NOAA Coral Reef Watch (2023).

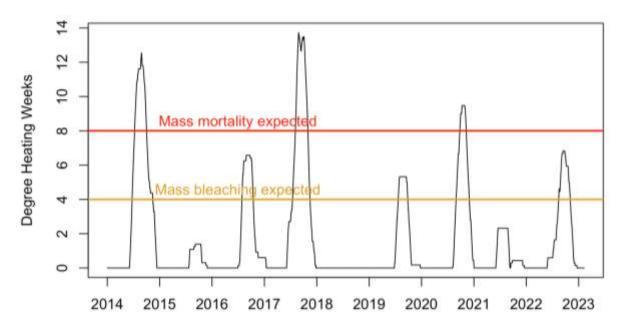


Figure 25. Coral Thermal Stress Exposure measured at CNMI Virtual Station 2014–2022 (Coral Reef Watch Degree Heating Weeks)

2.6.4.8 Chlorophyll-a and Anomaly

Rationale: Chlorophyll-a (Chl-A) is one of the most directly observable measures we have for tracking increasing ocean productivity.

Status: Annual mean Chl-A was $0.054~\text{mg/m}^3$ in 2022. Over the period of record, annual Chl-A has shown weak but significant linear decrease at a rate of $0.00039~\text{mg/m}^3$ /year. Monthly Chl-A values in 2022 ranged from 0.044- $0.079~\text{mg/m}^3$, within the climatological range of $0.043-0.095~\text{mg/m}^3$. The annual anomaly was $0.0015~\text{mg/m}^3$ lower than average.

Description: Chlorophyll-a concentration from 1998–2022, derived from the ESA Ocean Color Climate Change Initiative dataset, v6.0. A monthly climatology was generated across the entire period (1998–2021) to provide both a 2022 spatial anomaly, and an anomaly time series.

ESA Ocean Color Climate Change Initiative dataset is a merged dataset, combining data from SeaWIFS, MODIS-Aqua, MERIS, and VIIRS to provide a homogeneous time-series of ocean color. Data was accessed from the OceanWatch Central Pacific portal.

Timeframe: 1998–2022, daily data available, monthly means shown.

Region/Location: Global.

Measurement Platform: SeaWIFS, MODIS-Aqua, MERIS, and VIIRS.

Sourced from: NOAA OceanWatch (2023b).

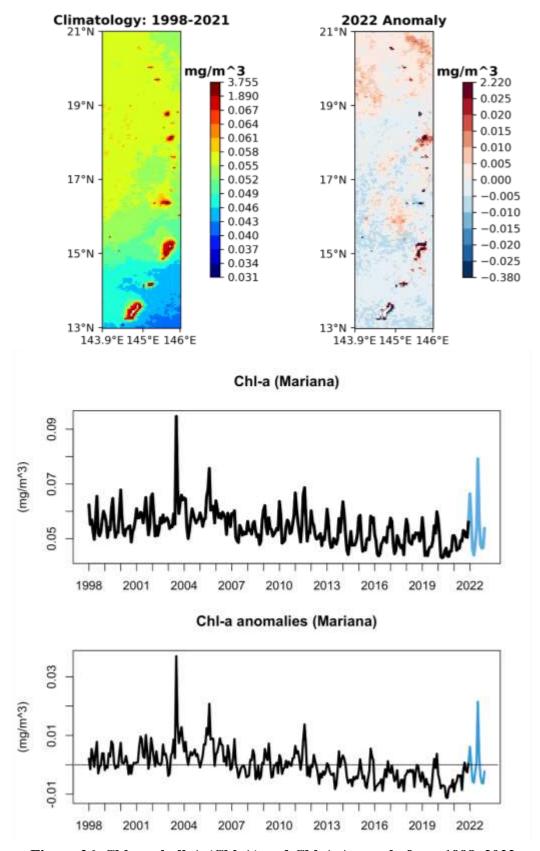


Figure 26. Chlorophyll-A (Chl-A) and Chl-A Anomaly from 1998–2022

2.6.4.9 Rainfall

Rationale: Rainfall may have substantive effects on the nearshore environment and is a potentially important co-variate with the landings of particular stocks.

Description: The CPC Merged Analysis of Precipitation (CMAP) is a technique which produces pentad and monthly analyses of global precipitation in which observations from rain gauges are merged with precipitation estimates from several satellite-based algorithms, such as infrared and microwave (NOAA, 2002). The analyses are on a 2.5 x 2.5 degree latitude/longitude grid and extend back to 1979. CMAP Precipitation data are provided by the NOAA/OAR/ESRL PSD, Boulder, Colorado, USA, from their website at https://www.esrl.noaa.gov/psd/. The data are comparable (but should not be confused with) similarly combined analyses by the Global Precipitation Climatology Project described in Huffman et al. (1997).

It is important to note that the input data sources to make these analyses are not constant throughout the period of record. For example, SSM/I (passive microwave - scattering and emission) data became available in July 1987; prior to that the only microwave-derived estimates available are from the MSU algorithm (Spencer 1993), which is emission-based, thus, precipitation estimates are available only over oceanic areas. Furthermore, high temporal resolution IR data from geostationary satellites (every 3-hr) became available during 1986; prior to that, estimates from the OPI technique (Xie and Arkin 1997) are used based on OLR from orbiting satellites.

The merging technique is thoroughly described in Xie and Arkin (1997). Briefly, the methodology is a two-step process. First, the random error is reduced by linearly combining the satellite estimates using the maximum likelihood method, in which case the linear combination coefficients are inversely proportional to the square of the local random error of the individual data sources. Over global land areas the random error is defined for each time period and grid location by comparing the data source with the rain gauge analysis over the surrounding area. Over oceans, the random error is defined by comparing the data sources with the rain gauge observations over the Pacific atolls. Bias is reduced when the data sources are blended in the second step using the blending technique of Reynolds (1988).

Text inserted from

https://www.cpc.ncep.noaa.gov/products/global_precip/html/wpage.cmap.shtml.

Timeframe: Monthly.

Region/Location: Global.

Measurement Platform: *In-situ* station gauges and satellite data.

Sourced from: NOAA ESRL (2023).

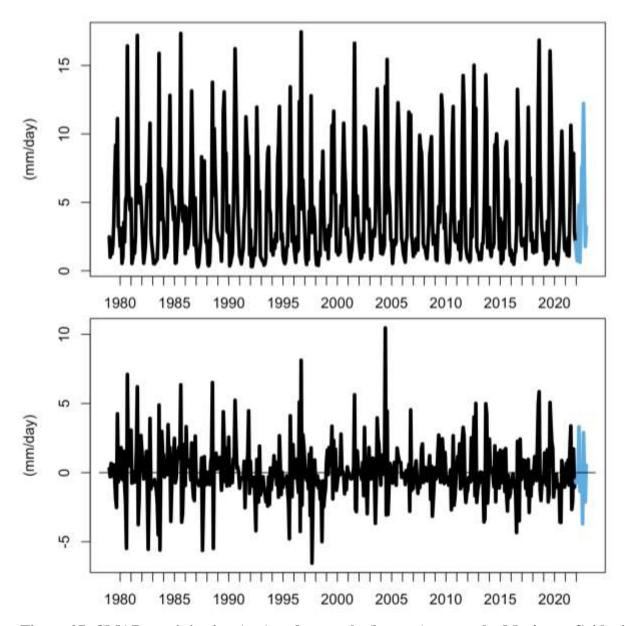


Figure 27. CMAP precipitation (top) and anomaly (bottom) across the Marianas Grid with 2022 values in blue

2.5.3.9 Sea Level (Sea Surface Height and Anomaly)

Rationale: Coastal: Rising sea levels can result in a number of coastal impacts, including inundation of infrastructure, increased damage resulting from storm-driven waves and flooding, and saltwater intrusion into freshwater supplies.

Description: Monthly mean sea level time series of local and basin-wide sea surface height and sea surface height anomalies, including extremes.

Timeframe: Monthly

Region/Location: Observations from selected sites within the Samoan Archipelago

Measurement Platform: Satellite and in situ tide gauges

Sourced from: Aviso (2023), NOAA (2023e), and NOAA CoastWatch (2023).

2.5.3.9.1 Basin-Wide Perspective

This image of the mean sea level anomaly for March 2022 compared to 1993-2016 climatology from satellite altimetry provides a glimpse into the 2022 continued La Niña conditions across the Pacific Basin. The image captures the fact that sea level is higher in the Western Pacific and lower in the Central and Eastern Pacific (this basin-wide perspective provides a context for the location-specific sea level/sea surface height images that follow).

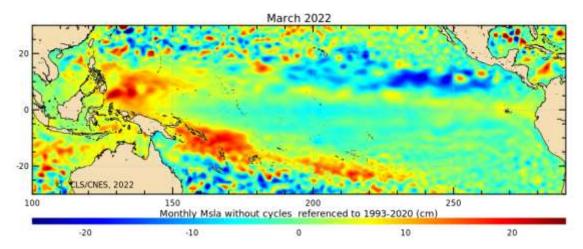


Figure 28a. Sea surface height and anomaly

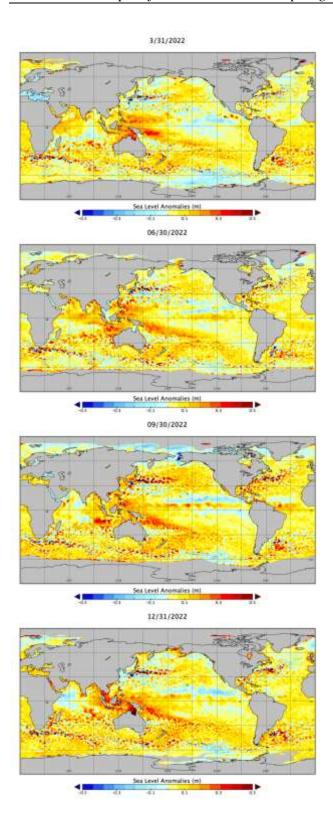


Figure 28b. Quarterly time series of mean sea level anomalies during 2022

Altimetry data are provided by the NOAA Laboratory for Satellite Altimetry, accessed from NOAA CoastWatch (2022).

2.5.3.9.2 Local Sea Level

These time-series from *in situ* tide gauges provide a perspective on sea level trends within each Archipelago (Tide Station Time Series from NOAA Center for Operational Oceanographic Products and Services, or CO-OPS).

The following figures and descriptive paragraphs were inserted from the NOAA Tides and Currents website. Figure 29 shows the monthly mean sea level without the regular seasonal fluctuations due to coastal ocean temperatures, salinities, winds, atmospheric pressures, and ocean currents. The long-term linear trend is also shown, including its 95% confidence interval. The plotted values are relative to the most recent Mean Sea Level datum established by CO-OPS. The calculated trends for all stations are available as a table in millimeters/year and in feet/century. If present, solid vertical lines indicate times of any major earthquakes in the vicinity of the station and dashed vertical lines bracket any periods of questionable data or datum shift.

The relative sea level trend is 5.01 millimeters/year with a 95% confidence interval of +/- 3.30 mm/yr based on monthly mean sea level data from 1993 to 2022 which is equivalent to a change of 1.64 feet in 100 years. The trend for 1948–1993 was -0.85 +/- 1.76 mm/yr.

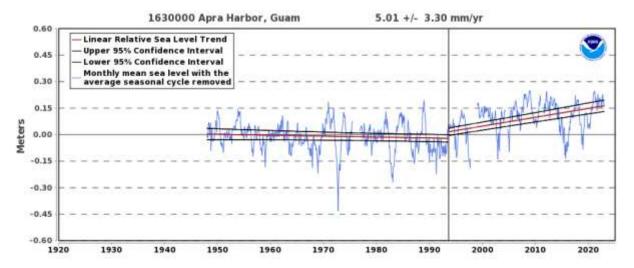


Figure 29. Monthly mean sea level without regular seasonal variability due to coastal ocean temperatures, salinities, winds, atmospheric pressures, and ocean currents

2.7 ESSENTIAL FISH HABITAT

2.7.1 Introduction

The Magnuson-Stevens Fishery Conservation and Management Act (MSA) includes provisions concerning the identification and conservation of essential fish habitat (EFH) and, under the EFH final rule, habitat areas of particular concern (HAPC) (50 Code of Federal Regulations [CFR] 600.815). The MSA defines EFH as "those waters and substrate necessary to fish for spawning, breeding, feeding, or growth to maturity." HAPC are those areas of EFH identified pursuant to 50 CFR 600.815(a)(8), and meeting one or more of the following considerations: (1) ecological function provided by the habitat is important; (2) habitat is sensitive to human-induced environmental degradation; (3) development activities are, or will be, stressing the habitat type; or (4) the habitat type is rare.

NMFS and the regional fishery management councils must describe and identify EFH in fishery management plans (FMPs) or fishery ecosystem plans (FEPs), minimize to the extent practicable the adverse effects of fishing on EFH, and identify other actions to encourage the conservation and enhancement of EFH. Federal agencies that authorize, fund, or undertake actions that may adversely affect EFH must consult with NMFS, and NMFS must provide conservation recommendations to federal and state agencies regarding actions that would adversely affect EFH. Councils also have the authority to comment on federal or state agency actions that would adversely affect the habitat, including EFH, of managed species. Fishery management actions must be evaluated for impacts to all EFH and HAPC in the area of effect and not just the EFH and HAPC for the fishery to which the management action applies.

The EFH Final Rule strongly recommends regional fishery management councils and NMFS to conduct a review and revision of the EFH components of FMPs every five years (600.815(a)(10)). The council's FEPs state that new EFH information should be reviewed, as necessary, during preparation of the annual reports by the Plan Teams. Additionally, the EFH Final Rule states "Councils should report on their review of EFH information as part of the annual Stock Assessment and Fishery Evaluation (SAFE) report prepared pursuant to §600.315(e)." The habitat portion of the annual SAFE report is designed to meet the FEP requirements and EFH Final Rule guidelines regarding EFH reviews.

National Standard 2 guidelines recommend that the SAFE report summarize the best scientific information available (BSIA) concerning the past, present, and possible future condition of EFH described by the FEPs.

2.7.1.1 EFH Information

The EFH components of FMPs include the description and identification of EFH, lists of prey species and locations for each managed species, and optionally, HAPC. Impact-oriented components of FMPs include federal fishing activities that may adversely affect EFH, non-federal fishing activities that may adversely affect EFH; non-fishing activities that may adversely affect EFH, conservation and enhancement recommendations, and a cumulative impacts analysis on EFH. The last two components include the research and information needs section, which feeds into the Council's Five-Year Research Priorities, and the EFH update procedure, which is described in the FEP but implemented in the annual SAFE report.

The Council has described EFH for five management unit species (MUS) under its management authority, most of which are no longer MUS: pelagic (PMUS), bottomfish (BMUS), crustaceans (CMUS), former coral reef ecosystem species (CREMUS), and precious corals (PCMUS).

EFH reviews of the biological components, including the description and identification of EFH, lists of prey species and locations, and HAPC, consist of three to four parts:

- Updated species descriptions, which can be found appended to the SAFE report. These can be used to directly update the FEP;
- Updated EFH levels of information tables, which can be found in Section 2.7.4;
- Updated research and information needs, which can be found in Section 2.7.5. These can be used to directly update the FEP; and
- An analysis that distinguishes EFH from all potential habitats used by the species, which is the basis for an options paper for the Council. This part is developed if enough information exists to refine EFH.

2.7.1.2 Habitat Objectives of FEP

The habitat objective of the FEP is to refine EFH and minimize impacts to EFH, with the following sub-objectives:

- Review EFH and HAPC designations every five years based on the best available scientific information and update such designations based on the best available scientific information, when available; and
- Identify and prioritize research to assess adverse impacts to EFH and HAPC from fishing (including aquaculture) and non-fishing activities, including, but not limited to, activities that introduce land-based pollution into the marine environment.

This annual report reviews the precious coral EFH components and non-fishing impacts components, resetting the five-year timeline for review. The Council's support of non-fishing activities research is monitored through the program plan and five-year research priorities, not the annual report.

2.7.1.3 Response to Previous Council Recommendations

At its 172nd meeting in March 2018, the Council recommended that staff develop an omnibus amendment updating the non-fishing impact to EFH sections of the FEPs, incorporating the non-fishing impacts EFH review report by Minton (2017) by reference. An options paper has been developed. The CNMI Joint Advisory Group provided comments on the non-fishing impacts review at a meeting held November 15, 2017, in Garapan. The Guam Joint Advisory Group also reviewed the report at their meeting held on November 17, 2017, in Tumon.

At its 187th meeting in September 2021, the Council recommended that the Chair recommend at the October 2021 CCC meeting that NMFS work with the Council to review EFH guidance in terms of how that guidance requiring the Council to identify and describe how EFH has been applied in the Western Pacific Region.

2.7.2 Habitat Use by MUS and Trends in Habitat Condition

The Mariana Archipelago is a chain of islands in the western Pacific roughly oriented north-south. It is anchored at the southern end by the relatively large island of Guam at 13.5° north

latitude. The Commonwealth of the Northern Mariana Islands (CNMI) stretches off to the north. The entire chain is approximately 425 miles long. The archipelago was named by Spanish explorers in the 16th Century in honor of Spanish Queen Mariana of Austria.

The total land area of Guam is approximately 212 square miles and its EEZ is just over 84,000 square miles. The CNMI consists of 14 main islands. From north to south these are: Farallon de Pajaros, Maug, Asuncion, Agrihan, Pagan, Alamagan, Guguan, Sarigan, Anatahan, Farallon de Medinilla, Saipan, Tinian, Aguijan, and Rota. Only Saipan, Rota, and Tinian are permanently inhabited, with 90% of the population residing on the island of Saipan. The total land area of the CNMI is 176.5 square miles and its EEZ is almost 300,000 square miles.

Guam and the southern islands of the CNMI are limestone, with level terraces and fringing coral reefs. The CNMI's northern islands are volcanic and sparsely inhabited, with active volcanoes on several islands, including Anatahan, Pagan, and Agrihan (the highest, at 3,166 feet). The archipelago has a tropical maritime climate moderated by seasonal northeast trade winds. While there is little seasonal temperature variation, there is a dry season (December to June) and a rainy season (July to November). The rainy season coincides with hurricane season, and the Mariana Archipelago is periodically impacted by powerful typhoons.

The Mariana Trench is located to the east of the chain and includes the deepest point in the world's oceans. The vertical measurement from the seafloor to Mount Tapotchau is 37,752 ft.

Essential fish habitat in the Marianas for the four MUS comprises all substrate from the shoreline to the 700 m isobath. The entire water column is described as EFH from the shoreline to the 700 m isobath, and the water column to a depth of 400 m is described as EFH from the 700 m isobath to the limit or boundary of the EEZ. The coral reef ecosystems surrounding the islands in the Mariana Archipelago been the subject of a comprehensive monitoring program through the PIFSC Coral Reef Ecosystem Division (CRED) biennially since 2002, surveys are focused on the nearshore environments surrounding the islands, atolls, and reefs. PIFSC CRED was replaced by the Coral Reef Ecosystem Program (CREP) within the PIFSC Ecosystem Sciences Division (ESD) before being shifted to the Archipelagic Research Program (ARP).

2.7.2.1 Habitat Mapping

The NOS Hydrographic Survey (Survey ID: H13572) was conducted in 2022 in the Mariana Archipelago to gather valuable information on the area's oceanic features. The primary data collected during the survey included water depths, as well as details on the presence of features such as rocks, wrecks, navigation aids, shoreline identification, and bottom type information. The survey results offer a fresh and comprehensive understanding of the Mariana Archipelago, adding new bathymetry and backscatter data to previously recorded findings. Additionally, the survey has complemented existing coral reef habitat data, thus contributing to a deeper understanding of the archipelago's ecology.

2.7.2.2 Benthic Habitat

Juvenile and adult bottomfish EFH extends from the shoreline to the 400 m isobath (64 FR 19067, April 19, 1999).

2.7.2.2.1 NCRMP Indicators

Benthic percent cover of coral, macroalgae, and crustose coralline algae are surveyed as a part of the NOAA's National Coral Reef Monitoring Program (NCRMP) led by the PIFSC ESD. In 2022, the PIFSC ESD conducted NCRMP surveys in Mariana Archipelago. These surveys provide EHF-relevant information to manage by assessing indicators like ocean acidification, water quality, temperature, and salinity. Various oceanographic instruments, deployed from the ship and underwater moored instruments, collect data on the physical and biological characteristics of the coral reef environment.

2.7.2.3 Oceanography, Water Quality, and Other Environmental Data

The water column is also designated as EFH for selected MUS life stages at various depths. For larval stages of all species except deepwater shrimp, the water column is EFH from the shoreline to the EEZ. Coral reef species egg and larval EFH is to a depth of 100 m; crustaceans, 150 m; and bottomfish, 400 m. Please see the Climate and Oceanic Indicators section (Section 2.6) for information related to oceanography and water quality.

In addition to the 2022 NOS Hydrographic Survey (Survey ID: H13572), satellite and buoy data are continuously collected and archived. PIFSC staff recently developed an advanced data compilation tool, the Environmental Data Summary (EDS), that gives users a simple, consistent way to enhance existing in situ observations with external gridded environmental data. The EDS is written in R and provides users an interface to NOAA CoastWatch and OceanWatch datasets through the ERDDAP server protocol. The EDS allows users to download, filter, and/or extract large amounts of gridded and tabular data given user-defined time stamps and geographical coordinates. The various external environmental data summarized at individual survey sites can aid scientists in assessing and understanding how environmental variabilities impact living marine resources. The EDS outputs were summarized at the National Coral Reef Monitoring Program (NCRMP) Rapid Ecological Assessment (REA) site level from 2000 to 2022 across 57 islands covered by the survey. PIFSC is planning to expand the utility of EDS with a broader range of gridded NOAA CoastWatch and OceanWatch data products (e.g., wave, wind) at finer spatiotemporal scales (e.g., water columns). Target data content includes spatial data (e.g., remote sensing), modeled data (e.g., Regional Ocean Modeling Systems), and socioeconomic data, including human density.

2.7.3 Report on Review of EFH Information

There were no EFH reviews completed in 2021. A review of the biological components of crustacean EFH in Guam and Hawaii was finalized in 2019 and can be found in Appendix C of the 2019 reports for the Hawaiian and Mariana Archipelagos (WPRFMC 2020a, WPRFMC 2020b). The non-fishing impacts and cumulative impacts components were reviewed in 2016 through 2017, which can be found in Minton (2017).

2.7.4 EFH Levels

NMFS guidelines codified at 50 C.F.R. § 600.815 recommend Councils organize data used to describe and identify EFH into the following four levels:

• Level 1: Distribution data are available for some or all portions of the geographic range of the species.

- Level 2: Habitat-related densities of the species are available.
- Level 3: Growth, reproduction, or survival rates within habitats are available.
- Level 4: Production rates by habitat are available.

The Council adopted a fifth level, denoted Level 0, for situations in which there is no information available about the geographic extent of a managed species' life stage. The existing level of data for individual MUS in each fishery are presented in tables per fishery.

2.7.4.1 Bottomfish and Seamount Groundfish

EFH for bottomfish was originally designated in Amendment 6 to the Bottomfish and Seamount Groundfish FMP (64 FR 19067, April 19, 1999). To analyze the potential effects of a proposed fishery management action on EFH, one must consider all designated EFH, but research examining depth and habitat requirements for most species is generally lacking (PIFSC 2021). The levels of information available for Mariana Archipelago BMUS did not change in 2022.

Table 64. Level of EFH information available for the Mariana Archipelago BMUS complex

| Life History Stage | Eggs | Larvae | Juvenile | Adult |
|---|------|--------|----------|-------|
| Aphareus rutilans (red snapper/silvermouth) | 0 | 0 | 0 | 1 |
| Caranx ignobilis (giant trevally/jack) | 0 | 0 | 1 | 1 |
| C. lugubris (black trevally/jack) | 0 | 0 | 0 | 1 |
| Etelis carbunculus (red snapper) | 0 | 0 | 1 | 1 |
| E. coruscans (red snapper) | 0 | 0 | 1 | 1 |
| L. rubrioperculatus (redgill emperor) | 0 | 0 | 0 | 1 |
| Lutjanus kasmira (blueline snapper) | 0 | 0 | 1 | 1 |
| Pristipomoides auricilla (yellowtail snapper) | 0 | 0 | 0 | 1 |
| P. filamentosus (pink snapper) | 0 | 0 | 1 | 1 |
| P. flavipinnis (yelloweye snapper) | 0 | 0 | 0 | 1 |
| P. sieboldii (pink snapper) | 0 | 0 | 1 | 1 |
| P. zonatus (snapper) | 0 | 0 | 0 | 1 |
| Variola louti (lunartail grouper) | 0 | 0 | 0 | 1 |

Table 65. EFH and HAPC for Mariana Archipelago BMUS

| Mariana Archipelago BMUS | EFH | HAPC |
|--|--|---|
| Lehi (Aphareus rutilans) Giant trevally (Caranx ignobilis) Black trevally (Caranx lugubris) Ehu (Etelis carbunculus) Onaga (E. coruscans) | EFH Eggs and larvae: the water column extending from the shoreline to the outer limit of the EEZ down to a depth of 400 m (200 fathoms, fm). | All slopes and escarpments between 40-280 m (20 and 140 |
| Redgill emperor (<i>Lethrinus rubrioperculatus</i>) Blueline snapper (<i>Lutjanus kasmira</i>) Yellowtail snapper (<i>Pristipomoides auricilla</i>) Opakapaka (<i>P. filamentosus</i>) Yelloweye snapper (<i>P. flavipinnis</i>) | Juvenile/adults: the water column and all bottom habitat extending from the shoreline to a depth of 400 m (200 fm). | fm). |

| Mariana Archipelago BMUS | EFH | НАРС |
|-----------------------------------|-----|------|
| Kalekale (P. sieboldii) | | |
| Gindai (P. zonatus) | | |
| Lunartail grouper (Variola louti) | | |

2.7.5 Project Updates

The PIFSC ESD planned to conduct the NOAA's National Coral Reef Monitoring Program (NCRMP) - Pacific Region surveys aboard the NOAA Ship *Rainier*, which provides scientific information to support ecosystem approaches to management and conservation of coral reefs. Diver-based surveys include fine-scale, rapid ecological assessment (REA) surveys of reef fishes and corals, as well as surveys to monitor nearshore physical and ecological factors associated with ocean acidification and general water quality, including data on water temperature, salinity, and other physical and biological characteristics of the coral reef environment using an assortment of oceanographic sampling and monitoring instruments, including systems deployed from the ship and underwater moored instruments. Survey areas include reef area around Guam and the Northern Mariana Islands of Rota, Aguijan, Tinian, Saipan, Sarigan, Zealandia Bank, Guguan, Alamagan, Pagan, Agrihan, Asuncion, Maug, Supply Reef, and Farallon de Pajaros. Since its inception in 2000, NCRMP-Pacific (formally known as Pacific Reef Assessment and Monitoring Program, or RAMP) established baseline ecosystem assessments and conducted long-term monitoring that integrates biological observations with oceanographic data as part of a long-term NOAA effort to monitor the status and trends of U.S. coral reef ecosystems. This cruise was delayed due to COVID, and the cruise departed in March 2022.

Research is ongoing to analyze a synthesized dataset (i.e., federal and jurisdictional data) to look at trends in benthic communities over space in time across the Mariana Archipelago. In 2022, the response of fish communities will be layered on top of this effort. A group of PIFSC staff is analyzing changes in benthic and fish composition across the PRIA. There's particular interest in both of these projects to determine variance in response to bleaching events.

Life history research is ongoing that attempts to determine spatial variability along latitudinal gradients in the Mariana Archipelago. The research plans to identify spatial variability in life history parameters across the archipelago and provide insights into how fish may respond to climate change as well as specific extreme thermal events. Creation of multiple individual time series (chronologies) for a complex of species (deepwater snappers, coral reef fish, coral)

2.7.6 Research and Information Needs

Based, in part, on the information provided in the tables above the Council identified the following scientific data which are needed to more effectively address the EFH provisions:

2.7.6.1 All FMP Fisheries

- Distribution of early life history stages (eggs and larvae) of management unit species by habitat.
- Juvenile habitat (including physical, chemical, and biological features that determine suitable juvenile habitat).
- Food habits (feeding depth, major prey species etc.).
- Habitat-related densities for all MUS life history stages.
- Growth, reproduction, and survival rates for MUS within habitats.

2.7.6.2 Bottomfish Fishery

- Inventory of marine habitats in the EEZ of the Western Pacific region.
- Data to obtain a better SPR estimate for American Samoa's bottomfish complex.
- Baseline (virgin stock) parameters (CPUE, percent immature) for the Guam/CNMI deep-water and shallow water bottomfish complexes.
- High resolution maps of bottom topography/currents/water masses/primary productivity.
- Habitat utilization patterns for different life history stages and species.

2.8 MARINE PLANNING

2.8.1 Introduction

Marine planning is a science-based management tool being utilized regionally, nationally, and globally to identify and address issues of multiple human uses, ecosystem health and cumulative impacts in the coastal and ocean environment. The Council's efforts to formalize incorporation of marine planning in its actions began in response to Executive Order (EO) 13547, *Stewardship of the Ocean, Our Coasts, and the Great Lakes*. EO 13158, *Marine Protected Areas*, proposes that agencies strengthen the management, protection, and conservation of existing MPAs, develop a national system of MPAs representing diverse ecosystems, and avoid causing harm to MPAs through federal activities. MPAs, or marine managed areas (MMAs) are one tool used in fisheries management and marine planning.

At its 165th meeting in March 2016, in Honolulu, Hawai`i, the Council approved the following objective for the FEPs: Consider the Implications of Spatial Management Arrangements in Council Decision-making. The following sub-objectives apply:

- Identify and prioritize research that examines the positive and negative consequences of areas that restrict or prohibit fishing to fisheries, fishery ecosystems, and fishermen, such as the Bottomfish Fishing Restricted Areas (BRFAs), military installations, NWHI restrictions, and Marine Life Conservation Districts (MLCDs).
- Establish effective spatially-based fishing zones.
- Consider modifying or removing spatial-based fishing restrictions that are no longer necessary or effective in meeting their management objectives.
- As needed, periodically evaluate the management effectiveness of existing spatialbased fishing zones in federal waters.

In order to monitor implementation of this objective, this annual report includes the Council's spatially-based fishing restrictions, or MMAs, the goals associated with those, and the most recent evaluation. Council research needs are identified and prioritized through the Five Year Research Priorities and other processes and are not tracked in this report.

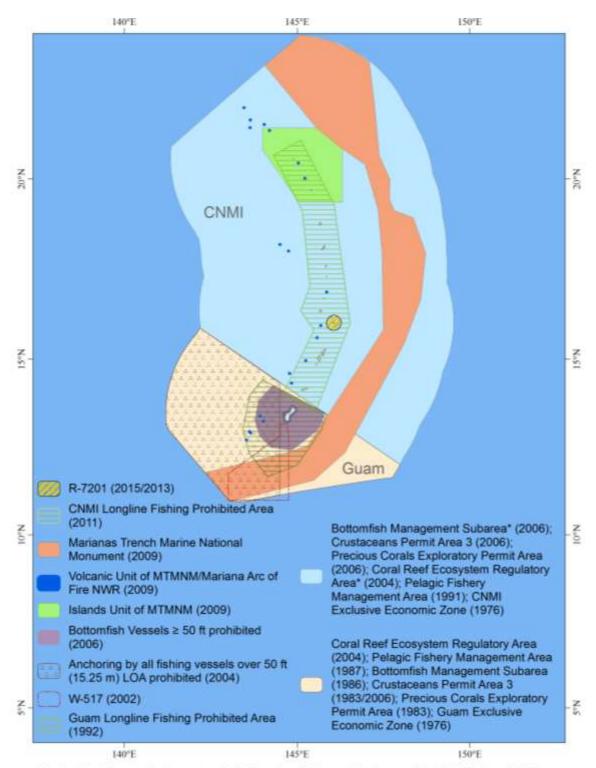
In order to meet the EFH and National Environmental Policy Act (NEPA) mandates, this annual report tracks activities that occur in the ocean that are of interest to the Council, and incidents or facilities that may contribute to cumulative impact. NMFS is responsible for NEPA compliance, and the Council must assess the environmental effects of ocean activities for the FEP's EFH cumulative impacts section.

2.8.2 Response to Previous Council Recommendations

There are no Council recommendations indicating review deadlines for Marianas MMAs.

2.8.3 Marine Managed Areas Established under FEPs

Council-established MMAs were compiled in Table 66 from 50 CFR § 665, Western Pacific Fisheries, the *Federal Register*, and Council amendment documents. All regulated fishing areas and large scale access restrictions, including the Mariana Trench Marine National Monument (MTMNM), are shown in Figure 30.



* The Coral Reef Ecosystem Regulatory Area excluded the portion of EEZ waters 0-3 miles around the CNMI. The Bottomfish Management Subarea was divided in the CNMI Inshore Area, which was that portion of the EEZ shoreward of 3 nautical miles of the shoreline of CNMI, and the CNMI Offshore Area, which was that portion of the EEZ seaward of 3 nautical miles from the CNMI shoreline.

Figure 30. Regulated fishing areas of the Mariana Archipelago

Table 66. MMAs established under FEPs from <u>50 CFR § 665</u>

| Name | FEP | Island | 50 CFR /FR /Amendment Reference | Marine Area (km²) | Fishing Restriction | Goals | Most Recent Evaluation | Review Deadline | | |
|--|------------------------|------------------------|--|----------------------|--|--|---------------------------|--------------------|--|--|
| Pelagic Restrictions | | | | | | | | | | |
| Guam Longline Prohibited Area | Pelagic | Guam | 665.806(a)(3) 57 FR 7661 Pelagic FMP Am. 5 | 50,192.88 | Longline fishing prohibited | Prevent gear conflicts between longline vessels and troll/handline vessels. | 1992 | - | | |
| CNMI Longline Prohibited Area | Pelagic | Mariana Archipelago | 665.806(a)(4) 76 FR 37287 | 88,112.68 | Longline fishing prohibited | Reduce potential for nearshore localized fish depletion from longline fishing, and to limit catch competition and gear conflicts between the CNMI-based longline and trolling fleets. | 2011 | - | | |
| Bottomfish Rest | trictions | | | | | | | | | |
| Guam Large Vessel Prohibited Area | Mariana Archipelago | Guam | 665.403(a) 71 FR 64474 Bottomfish FMP Am. 9 | 29,384.06 | Vessels ≥ 50 feet prohibited | To maintain viable participation and bottomfish catch rates by small vessels in the fishery. | 2006 | - | | |
| Other Restrictio | ns | | | | | | | | | |
| Guam No Anchor Zone | Mariana Archipelago | Guam | 665.399 69 FR 8336 Coral Reef Ecosystem FMP | 138,992.51 | Anchoring by all fishing vessels ≥ 50 ft. prohibited on the offshore southern banks located in the U.S. EEZ off Guam | Minimize adverse human impacts on coral reef resources. | 2004 | - | | |
| Marianas Trench Marine National Monument | Mariana Archipelago | Mariana Archipelago | 665.901(a) 78 FR 33003 Mariana Archipelago FEP Am. 3 | - | Commercial fishing prohibited; non- commercial fishing authorized under permit | Minimize adverse human impacts on marine resources within the marine national monument. | 2013 | - | | |

2.8.3.1 Mariana Trench National Marine Sanctuary Nomination – Five-Year Review

On January 21, 2022, the NOAA Office of National Marine Sanctuaries (ONMS) began facilitation of a review of the nomination for the Mariana Trench National Marine Sanctuary (NMS) at the five-year interval by requesting written and oral comments (87 FR 3284). On March 10, 2022, the NOAA OMNS extended the public comment period by an additional 45 days through April 25, 2022 (87 FR13709). ONMS will review information to its 11 evaluation criteria for inclusion in the inventory of nominations, emphasizing any new information about the significance of the area's natural or cultural resources, changes to any threats to these resources, and any updates to the management framework of the area. The original nominating parties for the NMS were Pew Charitable Trusts and Friends of the Marianas Trench, which will also have an opportunity to provide input on relevant information. Following information gathering and internal analysis, NOAA will make a final determination on whether or not the Mariana Trench NMS nomination will remain in the inventory for another five-year period.

The potential development of an NMS for the Mariana Trench is an issue of debate for residents of the Mariana Archipelago. The Marianas Trench Marine National Monument (MTMNM) already exists in the area, and the creation of an NMS would have the potential to further restrict fishing access or limit the potential for fisheries development within the EEZ around the CNMI. Regarding the nomination of the Mariana Trench NMS, concerns have been raised about the potential to expand the NMS beyond the boundaries of the MTMNM, the expansion of fishing restrictions to the water column within the Trench Unit of the MTMNM, and the current lack of community support and public confusion surrounding the nomination letter submitted by Friends of the Marianas Trench (Tenorio, pers. comm., April 4, 2022).

2.8.4 Fishing Activities and Facilities

There are no proposed or existing offshore aquaculture projects in federal waters of neither Guam nor CNMI.

2.8.5 Non-Fishing Activities and Facilities

The following section includes activities or facilities associated with known uses and predicted future uses. The Plan Team will add to this section as new facilities are proposed and/or built. Due to the sheer volume of ocean activities and the annual frequency of this report, only major activities on multi-year planning cycles are tracked in this report. Activities which are no longer reasonably foreseeable or have been replaced with another planning activity are removed from the report, though may occur in previous reports.

2.8.5.1 Alternative Energy Facilities

There are no proposed or existing alternative energy facilities in federal waters of neither Guam nor CNMI.

2.8.5.2 Military Training and Testing Activities and Impacts

The Department of Defense major planning activities in the region are summarized in Table 67. Activities that are no longer reasonably foreseeable or have been replaced with another planning activity were removed from the report, though may occur in previous reports. When a particular offshore area is in use for training or testing exercises by the U.S. military, a notice to mariners (NTM) is issued, and vessels attempting to use the area are advised to be cautious of objects in the water and other small vessels. This discourages access to many popular fishing areas. NTMs

from the military and the number of days affected for Guam and the CNMI are included in Table 68. The areas for which NTMs are issued are presented in Figure 31.

In addition to the Department of Defense activities detailed in Table 67, the U.S. military proposed the development of a small arms firing range near the shoreline in northwest Tinian. Several concerns were brought to the military regarding the proposed firing range, including issues with fishing access in the associated spatial closure area, issues with access to dive sites in the area, issues with the increased distance that boaters would have to travel to transit between Tinian and Saipan, and issues regarding boater safety due to having to travel further west from Tinian and losing access to the calmer waters nearshore (Tenorio, pers. comm., April 4, 2022).

Table 67. Department of Defense major planning activities

| Action | Description | Phase | Impacts |
|---|--|--|---|
| Guam and CNMI Military Relocation SEIS | Relocate Marines to Guam and build a cantonment/family housing unit on Finegayan/Andersen Air Force Base, a live-fire individual training range complex at the Ritidian Unit of the Guam National Wildlife Refuge. | Record of Decision (ROD) published August 29, 2015 after release of Final SEIS on July 18, 2015 (80 FR 55838). Lawsuit filed for segmentation and range of reasonable alternatives under NEPA. The case was lost in 2018 when a judge from the District Court of CNMI stated that the Guam buildup and proposed training in the CNMI are not connected actions. The case was appealed, and the US Court of Appeals for the Ninth Circuit affirmed the District Court's dismissal in 2020. Marine Corps Base Camp Blaz was activated on October 1, 2020. The US Army Corps of Engineers published a final rule on Oct. 8, 2021, amending regulations to establish a danger zone in the Pacific Ocean adjacent to the Mason Live-Fire Training Range Complex at Camp Blaz. | Danger zone established in the waters adjacent to Ritidian – access restricted during training. Northern District Wastewater Treatment Plant will significantly impact nearshore water quality until it is upgraded. |
| Mariana Islands Training and Testing — Supplemental | The supplement to the 2015 Final EIS/OEIS was prepared to support ongoing and future activities conducted at sea and on Farallon de Medinilla (FDM) beyond 2020. New information, including an updated acoustic effects model, updated marine mammal density data, and evolving and emergent BSIA, were used to update the MITT. | The MITT Final Supplemental EIS/OEIS was released in June 2020. ROD published on August 7, 2020, to continue training and testing activities in the study area (85 FR 47952). Meetings occurred to discuss FDM research activities and exercises. Meetings were previously held to discuss the Integrated Natural Resources Management Plan and plans for future surveys around FDM. In July 2020, NMFS implemented regulations regarding to the incidental take of marine mammals in the MITT area (85 FR 46302). | Access and habitat impact similar to previously analyzed activities in the 2015 EIS/OEIS (80 FR 29701). |

| Action | Description | Phase | Impacts |
|---|--|--|---|
| CNMI Joint Military Training | Establish unit and combined level training ranges on Tinian and Pagan. | The revised Draft EIS was expected in late 2018 or early 2019, but there is no new information on the status of the EIS. Lawsuit filed for segmentation and range of reasonable alternatives under NEPA. DOJ asked U.S. District Court for the NMI to dismiss the plaintiff's complaint with prejudice to prevent refiling. The case was lost in 2018 after a judge from the district court of CNMI agreed with the military that the Guam buildup and proposed training in the CNMI are not connected actions. The case was appealed, and the U.S. Court of Appeals for the Ninth Circuit affirmed the District Court's dismissal in 2020. Several meetings have been held with DFW and military officials to discuss relevant natural resource, land use, and social concerns regarding the proposed activities and prompted the reconsideration of proposed alternatives. Community engagement is also ongoing. | Significant access and habitat impacts. |
| Tinian Divert Infrastructure Improvements, Marianas | Improvements to airport and seaport (improving roads, installing fuel line) in CNMI for expanding mission requirements in the Western Pacific. | ROD for Tinian Divert Infrastructure Improvements published in 2016 (81 FR 92791). The USAF has published a NOI to prepare a SEIS for the proposed Tinian Divert Infrastructure Improvements. The NOI began the public scoping process for the SEIS, which ended on May 31, 2018. Substantive comments received during the public scoping period were taken into consideration during preparation of the Draft SEIS. The USAF published a Notice of Availability (NOA) for the Draft SEIS on May 17, 2019. The NOA began the public review period for the Draft SEIS, which ended on July 1, 2019. Substantive comments received during the public review period were taken into consideration during preparation of the Final SEIS, which had an NOA published in July 2020 (85 FR 43580). Ground was broken on the Tinian Divert airfield on Feb. 22, 2022, which is expected to be completed by Oct. 9, 2025. | Adverse impacts to EFH minimal; access near Port of Tinian fuel transfer facility affected. Access and transit to fishing grounds. |
| Garapan Anchorage | Military pre-positioned ships anchor and transit. | Expired Memorandum of Understanding with the CNMI Government. As of 2022, a new MOU had not been signed. | Access, invasive species, unmitigated damage to reefs. |

| Action | Description | Phase | Impacts |
|--------------------------|--|---|---|
| Farallon de Medinilla | Restricted airspace covering the island to 12 nmi radius to conduct military training scenarios using air-to-ground ordnance delivery, naval gunfire, lasers, and special operations training. | Final rule published March 13, 2017, effective June 22, 2017, designating a new area, R-2701A, that surrounds existing R-2701, encompassing airspace between a 3 nmi radius and 12 nmi radius of FDM (82 FR 13389). Proposed surface danger zone to 12 nmi. Meetings with military officials established that the 12 nmi radius is closed when exercises are being conducted, but a 3 nmi closure would instead be in effect year-round when exercises are not being conducted. Damage to submerged lands and fisheries to be included within consultation establishing continued U.S. interest in the island and compensation to the CNMI (Report to the President on 902 Consultations 2017). | Access – to fishing grounds and transit to fishing grounds and damage to submerged lands. |

Table 68. NTMs for Military Exercises in the Mariana Archipelago

| Year | Location | Number of Notices to Mariners Issued | Number of Days Affected |
|------|----------|---|----------------------------|
| 2013 | FDM | 45 | 159 |
| 2015 | W-517 | 24 | 54 |
| 2014 | FDM | 38 | 145 |
| 2014 | W-517 | 24 | 49 |
| 2015 | FDM | 37 | 164 |
| 2015 | W-517 | 33 | 87 |
| | FDM | 35 | 142 |
| 2016 | W-517 | 50 | 139 |
| 2016 | W-11 | NA | NA |
| | W-12 | NA | NA |
| | FDM | 56 | 191 |
| 2017 | W-517 | 46 | 119 |
| 2017 | W-12 | 2 | 5 |
| | W-11 | NA | NA |
| | FDM | 38 | 150 |
| 2019 | W-517 | 49 | 107 |
| 2018 | W-12 | 6 | 13 |
| | W-11 | 1 | 1 |
| 2019 | FDM | 39 | 165 |

| Year | Location | Number of Notices to Mariners Issued | Number of Days Affected |
|-------|----------|---|----------------------------|
| | W-517 | 27 | 65 |
| | W-12 | 3 | 22 |
| | W-11 | 6 | 27 |
| | W-13 | 15 | 37 |
| | FDM | 17 | 62 |
| | W-517 | 12 | 26 |
| 2020 | W-12 | 5 | 10 |
| | W-11 | 3 | 8 |
| | W-13 | 15 | 62 |
| | FDM | N/A | 49 |
| | W-517 | N/A | 80 |
| 2021* | W-12 | N/A | 32 |
| | W-11 | N/A | 41 |
| | W-13 | N/A | 63 |

^{*}Data for 2021 are incomplete. The number of notices to mariners is not able to be reported for 2021 due to changes in how the Department of Defense presents aggregate NTM data. Additionally, military departments did not issue NTMs from August to December of 2021, so the presented data are from January to July 2021.

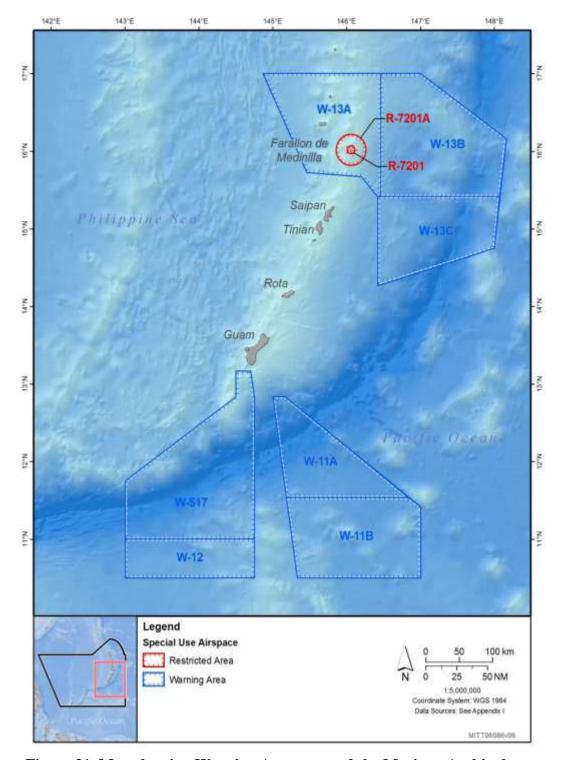


Figure 31. Map showing Warning Areas around the Mariana Archipelago

2.8.6 Mariana Archipelago Spatial Planning Initiatives

Spatial planning has occurred in CNMI in Saipan Lagoon. CNMI Division of Coastal Resources Management developed the <u>Saipan Lagoon Use Management Plan</u>, which was updated in 2017 and has an associated mapping tool.

3 DATA INTEGRATION

3.1 INTRODUCTION

3.1.1 Potential Indicators for Nearshore Fisheries

The purpose of this section ("Chapter 3") of the Stock Assessment and Fishery Evaluation (SAFE) annual report is to identify and evaluate potential fishery ecosystem relationships between fishery parameters and ecosystem variables to assess how changes in the ecosystem affect fisheries in the Mariana Archipelago and across the Western Pacific region. "Fishery ecosystem relationships" are those associations between various fishery-dependent data measures (e.g., catch, effort, or catch-per-unit-effort), and other environmental attributes (e.g., precipitation, sea surface temperature, primary productivity) that may contribute to observed trends or act as potential indicators of the status of prominent stocks in the fishery. These analyses represent a first step in a sequence of exploratory analyses that will be utilized to inform new assessments of what factors may be useful going forward.

To support the development of Chapter 3 of the annual SAFE report, staff from the Council, NMFS PIFSC and PIRO, and Triton Aquatics (consultants), held a SAFE Report Data Integration Workshop (hereafter, "the Workshop") convened on November 30, 2016 to identify potential fishery ecosystem relationships relevant to local policy in the WPR and determine appropriate methods to analyze them. The archipelagic fisheries group developed nearly 30 potential fishery ecosystem relationships to examine across bottomfish, coral reef, and crustacean fisheries based on data reliability, suitability of methodology, repeatability on an annual basis, and how well analyses could potentially inform management decisions.

Brief introductory analyses, presented in this section and initially introduced in the 2017 report, were intended to be "proof of concept" such that similar evaluations could be carried out on remaining fishery data for the Mariana Archipelago in the future. However, the Archipelagic Fishery Ecosystem Plan Team determined that the quantitative analyses presented here were not sufficient to act as a model for future evaluations. Using the direction from the Plan Team, the data integration module was updated for the Hawaii Archipelagic annual SAFE report in 2018, but each of the remaining archipelagic reports still contains data integration assessments from 2017. The annual SAFE report for the Mariana Archipelago will be updated in the coming years similar to the annual SAFE report for the Hawaii Archipelago pending Plan Team support.

Going forward, relationships deemed potentially relevant will be emphasized and recommended for further analysis. In subsequent years, this chapter will be updated with these analyses through the SAFE report process as the strength of certain fishery ecosystem relationships relevant to advancing ecosystem-based fishery management are determined.

To begin, this chapter described feedback from the Plan Team, SSC, and Council members on the initial drafts of the data integration module. Next, the chapter includes brief descriptions of past work on fishery ecosystem relationship assessment in coral reefs of the U.S. Western Pacific, followed by initial evaluations of relationships previously recommended for evaluation by participants of the Workshop using current data streams from the Mariana Archipelago. The evaluations completed were exploratory in nature, being the first step of analyses to know which comparisons may be more useful to focus on going forward.

Going forward with the analyses and presentation of results for the data integration chapter of the Marianas Archipelago Annual SAFE Report, the Plan Team suggested several improvements to implement in the coming year: standardizing and correcting values in CPUE time series, incorporating longer stretches of phase lag, completing comparisons on the species-level and by dominant gear types, incorporating local knowledge on shifts in fishing dynamics over the course of the time series, and utilizing the exact environmental data sets presented in the ecosystem consideration chapter of the annual report. Many of these recommendations were applied to datasets from Hawaii in 2018 and will similarly be done for Mariana Archipelago data integration analyses in the upcoming report cycles. Implementation of these suggestions will allow for the preparation of a more finalized version of the data integration chapter in future report cycles.

3.1.2 2018 Recommendations for Chapter Development

At the FEP Plan Team Meeting held on April 30th – May 1st, 2018, participants were presented preliminary data integration results shown here, and provided detailed recommendations to support the ongoing development of the data integration section of the Archipelagic Annual SAFE Report. These suggestions, both general and specific, will be implemented in the coming year to ensure that more refined analyses comprise the data integration section. FEP Plan Team participants recommended that:

- CPUE data should be standardized and calculated in a more robust fashion, measuring the average catch per unit effort rate over the course of a year to analyze variance.
- Analyses of fishery performance data against environmental variables should focus on dominant gear types rather than the entirety of the fishery or other gear aggregates (e.g., purse seine harvest of *Selar crumenophthalmus* in the MHI).
- There should be additional phase lag implemented in the analyses.
- Local knowledge of fishery dynamics, especially pertaining to shifting gear preferences, should be utilized. Changes in dynamics that may have impacted observed fishery trends over the course of available time series, both discreetly and long-term for taxa-specific and general changes should be emphasized.
- Spatial specificity and precision should be increased for analyses of environmental variables in relation to areas commonly fished.

The analyses presented in the data integration chapter of the 2018 Hawaii annual SAFE report are a reflection of a thoughtful re-approaching to these data integration evaluations based on this feedback. Additional data can be added to either time series as they are made available. Incorporating such recommendations into the 2018 version of the Mariana Archipelago Annual SAFE Report will mark the beginning of a standardized process to implement current data integration analyses on an annual basis. Doing so will promote more proactive management action with respect to ecosystem-based fishery management objectives.

3.1.3 Past Work

Richards et al. (2012) performed a study on a range of environmental factors that could potentially affect the distribution of large-bodied coral reef fish in Mariana Archipelago. Large-bodied reef fish were determined to typically be at the greatest risk of overfishing, and their distribution in the region was shown to be negatively associated with human population density. Additionally, depth, sea surface temperature (SST), and distance to deep water were identified as important environmental factors to large-bodied coral reef fish, whereas topographic complexity,

benthic habitat structure, and benthic cover had little association with reef fish distribution in the Mariana Archipelago.

Kitiona et al. (2016) completed a study of the impacts on climate and/or ecosystem change on coral reefs fish stocks of American Samoa using climate and oceanic indicators (see Section 2.5.4). The evaluation of environmental variables showed that certain climate parameters (e.g., SST anomaly, sea level height, precipitation, and tropical storm days) are likely linked to fishery performance. It was also noted that larger natural disturbances in recent decades, such as cyclones and tsunamis, negatively impacted reef fish assemblages and lowed reef fishery CPUE in American Samoa (Ochavillo et al. 2012).

On a larger spatial scale, an analysis of various drivers on coral reef fish populations across 37 U.S.-affiliated islands in the Central and Western Pacific was performed by Williams et al. (2015) and evaluated relationships between fish biomass in these reefs with human and environmental factors. Again, reef fish assemblages were negatively associated with increasing human population density (even at relatively low levels) across the WRP but were positively associated with elevated levels of ocean productivity across islands. The authors warned, however, that the ability of reefs surrounding uninhabited islands to maintain fish populations varies, and that high biomass observed in remote areas (e.g., the NWHI) may not necessarily be reflective of baselines or recovery response levels for all reef systems.

A common method of EBFM used in coral reef ecosystems is the implementation of biological reference points, statistical indicators of potential overfishing used to help determine how a fishery is performing relative to these points at a given time (McClanahan et al. 2011). Hawhee (2007) adapted this idea, generating biological reference points in the form of CPUE-based proxies to be used as indicators for reef fish stocks in the WPR. However, the devised method was determined to be inappropriate for application in management of reef stocks in the U.S. Western Pacific due to the lack of a historical CPUE to use as a baseline for the reference points and their limit thresholds (Remington and Field 2016).

3.2 PRECIPITATION

3.2.1 **Guam**

Participants of the Workshop determined that the potential fishery ecosystem relationships between precipitation levels and atulai and opelu (bigeye scad and mackerel scad, *Selar crumenophthalmus* and *Decapterus macarellus*, respectively) were among the highest priority of those involving coral reef fisheries in the Mariana Archipelago. It has been suggested that the recruitment of small tropical pelagic fish is related to annual rainfall and subsequent runoff enrichment (Longhurst and Pauly 1987; Weng and Sibert 2000). The direct freshwater and nutrient input to reefs associated with increased precipitation can alter the physiochemical composition of the water, and it has been shown that reef assemblages are positively associated with this sort of increased ocean productivity (Williams et al. 2015). Data for precipitation in the Mariana Archipelago was gathered from local databases maintained by the National Weather Service (NWS-G). The time series of total annual precipitation from showed a non-significant, slightly variable trend over the last 30 years (R² = 0.05, CV = 19.5; Figure 32).

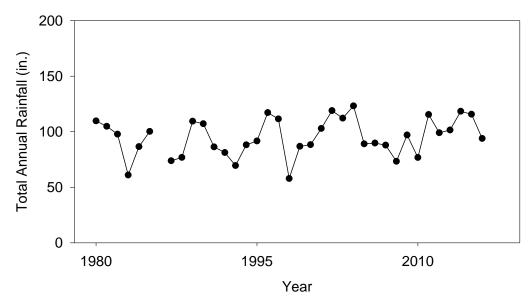


Figure 32. Total annual precipitation (in.) in Guam from 1980-2016

3.2.1.1 Evaluating relationship with atulai

Total annual estimated atulai catch in the Guam recreational coral reef fishery according to shore- and boat-based creel surveys showed no general trend over the last thirty years, with relatively large variability likely due to several years of catch orders of magnitude greater than previous or subsequent years (e.g., 2009; $R^2 = 0.01$; CV = 119.5; Figure 33). Combined effort statistics between shore- and boat-based creel survey statistics could not be generated because the proxies used to measure effort in each survey are different (i.e., number of gear hours versus number of boat trips). Similarly, because effort could not be standardized across the data sets, CPUE could not be generated on the individual family level at which these evaluations are taking place.

Examining effort, Guam shore-based creel survey data show that there are considerable differences in the number of samples recorded across gear types. The most frequently sampled gear in the shore-based survey was hook and line by an order of magnitude and had catch estimated to be several times greater than that in the expanded dataset (Figure 34a-b). Effort data also revealed that, despite catch statistics, the gill net had been sampled the least frequently among the top gears (Figure 34a-b). Boat-based effort data show that bottom fishing was sampled approximately twice as much than the other three top gears, but the difference in the expanded estimates between were at least an order of magnitude greater (Figure 34c-d). Generally, each of the time series for prominent gear types in Guam showed a slight shift but seemingly no net change over the course of available data despite interannual variability.

Total estimated atulai catch and rainfall in Guam showed no statistical association with one another such that would allow for assessment of the fishery ecosystem relationship between the two (R^2 =0.02; Figure 35). However, there seemed to be a slight observable negative relationship between the two (r = -0.15), indicating that catch may have experienced a minor decrease in years with more rainfall. Additionally, there was no association between annual rainfall amounts and total estimated atulai catch in Guam when only considering shore-based data, boat-based data, or prominent gear types.

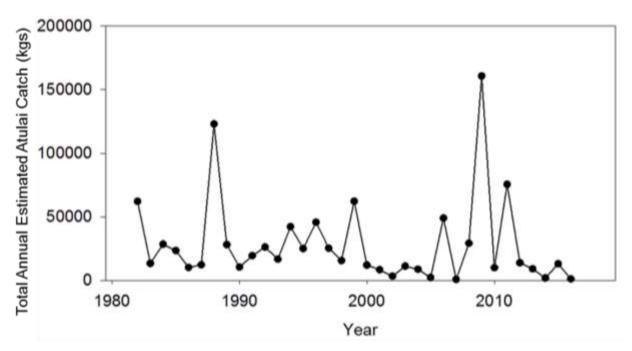


Figure 33. Time series of total annual estimated (i.e., expanded) landings of atulai in kilograms from Guam shore-and boat-based creel survey records from 1982-2016

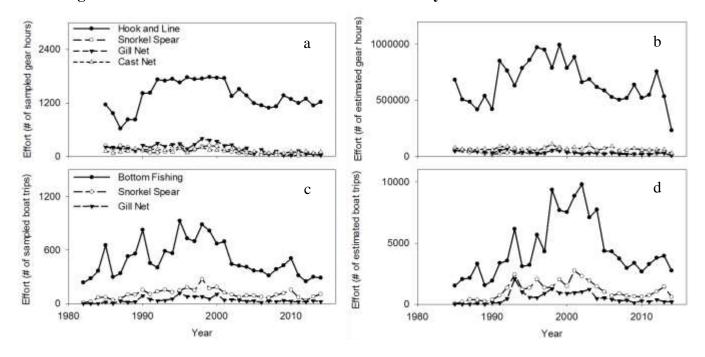


Figure 34. Time series of total sampled (left) and expanded (right) effort for top gear types in shore-based (top) and boat-based (bottom) creel surveys in Guam from 1982-2016

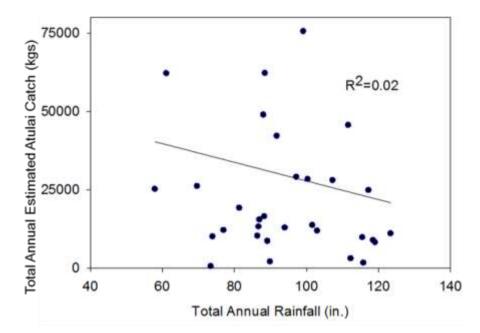


Figure 35. Linear regression between total atulai catch (kg) in the Guam shore-based and boat-based creel survey records and total annual rainfall (in.) from 1982-2016

3.2.1.2 Evaluating relationship with *D. macarellus*

Decapterus macarellus (i.e., mackerel scad) records from creel surveys in Guam were scant and had high variability, with estimated catch for many years being close to zero while others had close to 8,000 kg ($R^2 = 0.01$; CV = 278.4; Figure 36). Several years where mackerel scad catch data were available, they indicated a total amount landed of just a few kilograms (e.g., 1999, 2001, 2013, etc.; Figure 36). Because there were 17 of 35 total years with available mackerel scad catch data across gear types for the entire territory since 1982, many with extremely low catch estimates, the time series were not able to be used for comparison to rainfall records in the same region over the last thirty years.

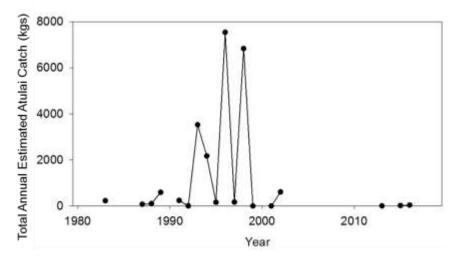


Figure 36. Time series of total annual expanded landings of *Decapterus macarellus* (kg) in Guam shore-and boat-based creel survey records from 1982-2016

In summary, no fishery ecosystem relationship could be established between atulai or mackerel scad catch with precipitation in Guam from 1982 till present without the incorporation of phase lag, and no standardized index/threshold characteristic of the association between the parameters could be identified representative of an immediate population response. The general lack of recreational harvest data for mackerel scad in Guam hindered the ability to determine whether a relationship exists with rainfall in that portion of the fishery. Analyses including atulai data had similar comparisons with rainfall data completed in the MHI as well, though no notable relationship between atulai catch and annual precipitation was identified there.

3.3 SEA SURFACE TEMPERATURE

Sea surface temperature (SST) is a commonly used diagnostic tool in monitoring climate change and its affects both regionally and globally, as it is representative of changes in ocean temperatures over time that can affect coastal fisheries (see Section 2.5.4). The potential influence of temperature-derived variables in fishery ecosystem relationships for U.S. Western Pacific coral reef stocks was deemed to be among the highest priority by the participants of the Workshop. Data for SST was gathered from the NOAA's AVHRR Pathfinder v5.0 through the OceanWatch program in the Central Pacific (NOAA/NESDIS/OceanWatch).

A time series of SST for the CNMI from 1985-2016 is shown in Figure 37. SST here had slightly less variability over time than Guam (CV = 0.55), again indicating relative stability. Unlike Guam, the CNMI did not seem to be observably increasing or decreasing over the time series of available data. The hottest temperature in the last three decades was approximately 29°C, where preceding SST had largely been stable over time. The average SST over the course of evaluated data was 28.8°C, slightly warmer than observed in Guam. The lowest recorded SST over the course of the time series was just about 27.5°C in the year 1996 (Figure 37).

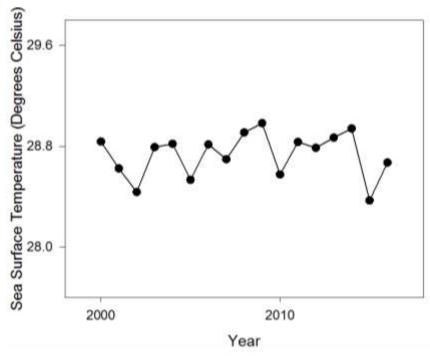


Figure 37. Time series of SST ($^{\circ}$ C) in the CNMI from 1985-2016 (CV = 0.55)

A time series of SST for Guam from 1985-2016 is shown in Figure 38. Temperature had low variability over time (CV = 1.38), suggesting relative stability. There was also a seeming increase in temperature over the last three decades, with some of the hottest temperatures recorded observed in the last five years. The average SST over the course of evaluated data was 28.6°C. The highest recorded SST over the course of the time series was just over 29°C in the year 1999, whereas the lowest was earlier in the 1990s (27.7°C; Figure 38).

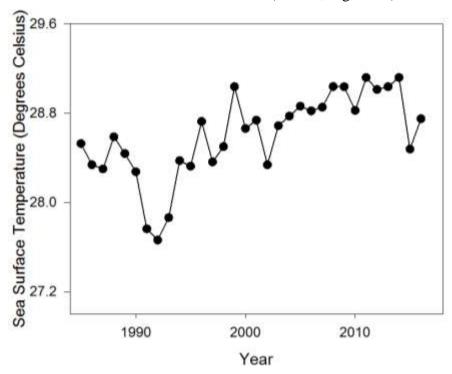


Figure 38. Time series of SST ($^{\circ}$ C) from 1985-2016 in Guam (CV = 1.38)

3.3.1 CNMI

3.3.1.1 Evaluating relationship for entire reef fishery

A plot showing the relationship between SST and catch time series from the recreational coral reef fishery in the CNMI from 2000-2016 is depicted in Figure 39. Landings were variable over the course of the time series (CV = 19.4), but less so than observed in catch time series in Guam. Total annual catch in the fishery has been observably decreasing over the last decade and a half despite an abrupt increase in 2013 resulting in the recorded maximum catch over this period (\sim 338,000 kg). Recent recorded catch levels (i.e., for 2016) were the lowest for the fishery through the available time series of data (\sim 165,000 kg; Figure 39).

In performing comparisons between fishery parameters and environmental variables such as SST, data were grouped in taxa categories based on family due to scarcity of data on the species level in many cases. Table 69 displays the different dominant family groups considered as well as their common names.

Linear regressions and correlation analyses performed on the time series of recreational coral reef fishery catch (kg) and annual mean SST from the CNMI are reported in Table 70. The comparisons between the two parameters showed a negatively significant relationship between 2000 and 2016 ($R^2 = 0.30$, p = 0.02; Table 70; Figure 40). The relationship between the total

annual catch and average annual SST for the whole fishery were associated such that for every degree Celsius of temperature increase, catch would decrease by approximately 105,000 kg (Figure 40).

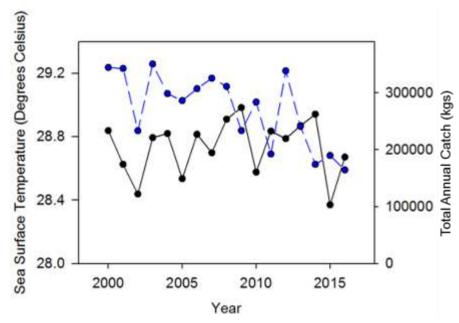


Figure 39. Time series of total annual catch (kg; blue) for the CNMI recreational coral reef fishery plotted alongside average annual SST ($^{\circ}$ C; black) from 2000-2016

Table 69. Families in creel surveys from the U.S. Western Pacific analyzed in this report

| Four-letter code | Family | Common Name |
|------------------|---------------|--------------------------|
| LUTJ | Lutjanidae | snappers |
| LETH | Lethrinidae | emperors |
| CARA | Carangidae | jacks/mackerel/trevally |
| ACAN | Acanthuridae | unicornfish/tang |
| SERR | Serranidae | Sea bass/grouper |
| SIGA | Siganidae | rabbitfish |
| SCAR | Scaridae | parrotfish |
| MULL | Mullidae | goatfish |
| MUGI | Mugilidae | mullet |
| LABR | Labridae | wrasse |
| HOLO | Holocentridae | squirrelfish/soldierfish |
| BALI | Balistidae | triggerfish |

Table 70. Correlation coefficients (r) between recreational coral reef fishery catch (kg) and SST (°C) in the CNMI for 12 top taxa harvested from 2000-2016

| Taxa Code | Total Catch | LUTJ | LETH | CARA | ACAN | SERR | SIGA | SCAR | MULL | MUGI | LABR | HOLO | BALI |
|-----------|--------------------|------|-------|-------|-------|-------|------|-------|-------|------|------|------|-------|
| n = 17 | | | | | | | | | | | | | |
| p | 0.02 | 0.49 | 0.54 | 0.26 | 0.70 | 0.91 | 0.99 | 0.88 | 0.06 | - | 0.59 | 0.91 | 0.82 |
| r | -0.55 | 0.18 | -0.16 | -0.29 | -0.10 | -0.03 | 0.00 | -0.04 | -0.47 | - | 0.14 | 0.03 | -0.06 |
| R^2 | 0.30 | 0.03 | 0.02 | 0.09 | 0.01 | 0.00 | 0.00 | 0.00 | 0.22 | - | 0.02 | 0.00 | 0.00 |

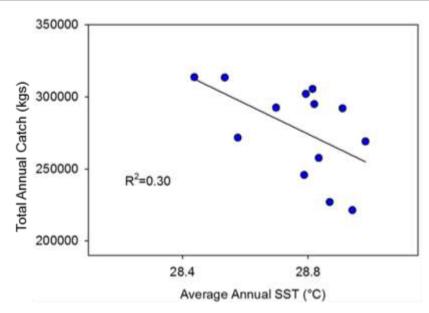


Figure 40. Linear regression showing the correlation between total annual catch (kg) in creel survey records and average annual SST (°C) in the CNMI from 2000-2016

3.3.1.2 Evaluating relationship for dominant taxa

Correlation and regression analyses were performed on prominent taxa in the CNMI recreational coral reef fishery, and it was found that no individual taxa had significant relationships with SST data (Table 70). The strongest associations between fishery catch and SST were observed from the Mullids ($R^2 = 0.22$, p = 0.06; Figure 41a), Carangids ($R^2 = 0.09$, p = 0.26; Figure 41b), and Lutjanids ($R^2 = 0.03$, p = 0.49; Figure 41c). While the relationship between catch and temperature for families Mullidae and Carangidae were negative, the Lutjanidae family had a positive relationship (Table 70).

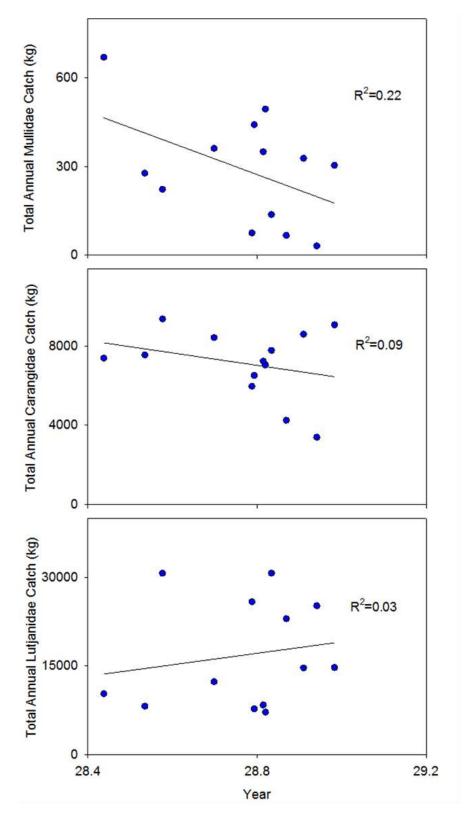


Figure 41. Linear regressions showing the three top correlations between total annual catch (kg) from creel survey records and average annual SST (°C) in the CNMI from for (a) Mullids, (b) Carangids, and (c) Lutjanids from 2000–2016

3.3.2 **Guam**

3.3.2.1 Evaluating relationship for entire reef fishery

An individual plot depicting the comparisons of time series of SST and catch from the recreational coral reef fishery in Guam from 1985-2016 is shown in Figure 42. Landings were variable over the course of the time series (CV = 28.1) though relatively stable, especially before the year 2000. There was a relatively abrupt observed decrease in total annual catch from 1998 to 2005, where recorded landings went from over half a million kg to approximately 180,000 kg in less than a decade. Catch has slightly rebounded since that minimum, with landings reaching over 400,000 kg in six of the last seven years (Figure 42).

Multiple linear regressions and correlation analyses were performed on time series of recreational coral reef fishery catch and annual mean SST from Guam (Table 71). Evaluations measuring the association between SST and total catch for the entirety of the recreational coral reef fishery in Guam showed a negatively significant relationship between 1985 and 2016 ($R^2 = 0.20$, p = 0.02; Table 71; Figure 43). The relationship between the total annual catch and average annual SST were associated such that for every degree Celsius of temperature increase, catch would decrease by approximately 120,000 kg (Figure 43).

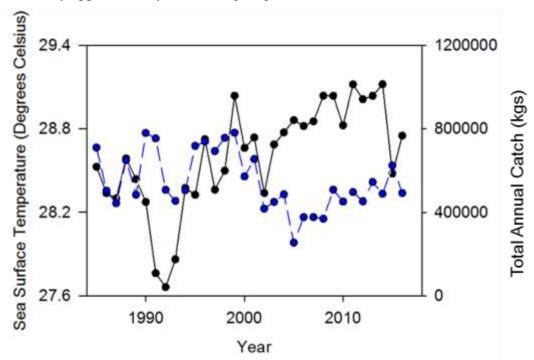


Figure 42. Time series of total annual catch (kg; blue) in the Guam shore-and boat-based creel survey records plotted with average annual SST (°C; black) from 1985-2016

Table 71. Correlation coefficients (r) between recreational coral reef fishery catch (in kg) and SST (°C) in Guam for 12 top taxa harvested from 1985-2016

| Taxa Code | Total Catch | LUTJ | LETH | CARA | ACAN | SERR | SIGA | SCAR | MULL | MUGI | LABR | HOLO | BALI |
|-----------|--------------------|-------|-------|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| n = 28 | | | | | | | | | | | | | |
| p | 0.02 | 0.01 | 0.00 | 0.01 | 0.39 | 0.01 | 0.00 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 |
| r | -0.45 | -0.80 | -0.48 | 0.17 | -0.50 | -0.54 | -0.71 | -0.51 | -0.56 | -0.66 | -0.60 | -0.63 | -0.43 |
| R^2 | 0.20 | 0.64 | 0.23 | 0.03 | 0.25 | 0.30 | 0.50 | 0.26 | 0.31 | 0.43 | 0.35 | 0.39 | 0.18 |

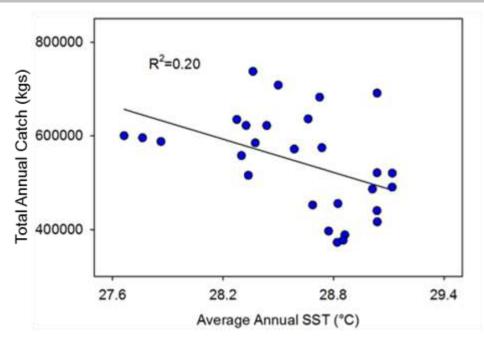


Figure 43. Linear regression between total annual catch (kg) for shore- and boat-based creel survey records and average annual SST (°C) in Guam from 1985-2016

3.3.2.2 Evaluating relationship for dominant taxa

Comparisons were made for the time series of catch for prevalent taxa in Guam's recreational reef fishery as well, and it was found that all except for the Acanthuridae family showed negative statistically significant correlations with SST (Table 71). The strongest relationship observed was of that between SST and annual Lutjanidae catch, where the regression suggested that for every degree Celsius of temperature increase, catch would decrease by approximately 7,500 kg ($R^2 = 0.64$, p = 0.00; Table 71; Figure 44a). The next two strongest associations observed were for families Siganidae ($R^2 = 0.50$, p = 0.00; Figure 44b) and Mugilidae ($R^2 = 0.43$, p = 0.01; Figure 44c). The regressions performed with temperature for taxa, suggesting negative relationships with temperature, also showed that for every degree of temperature increase in degrees Celsius, Siganidae and Mugilidae recreational catch in Guam would decrease by approximately 10,000 kg and 7,500 kg, respectively.

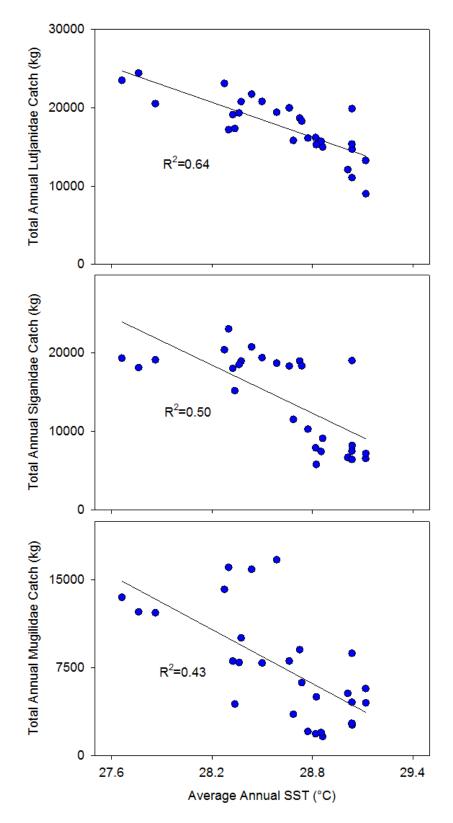


Figure 44. Linear regressions showing three top correlations between total annual catch (kg) for shore-and boat-based creel survey records and average annual SST (°C) in Guam for (a) Lutjanids, (b) Siganids, and (c) Mugilids from 1985–2016

In summary, Guam and the CNMI had fishery ecosystem relationships that could be identified for the entirety of the recreational coral reef fishery. The relationship between the total annual catch and average annual SST in Guam were associated such that for every degree Celsius of temperature increase, catch would decrease by approximately 120,000 kg. The relationship between the total annual catch and average annual SST in the CNMI were associated such that for every degree Celsius of temperature increase, catch would decrease by approximately 105,000 kg.

In Guam, the linear regressions performed showed that all evaluated taxa except for the Acanthurids had a statistically significant negative relationship with average annual temperature. The three strongest associations with SST were with the Lutjanids, Siganids, and Mugilids, such that the total annual catch for each would decrease by approximately 7,500-10,000 kg for every increase in SST by one degree Celsius. In the CNMI, conversely, there were no individual family groups whose catch data had statistically significant associations with temperature, though the strongest associations observed were the Mullids (relatively close to the threshold of significance, p = 0.06), Carangids, and Lutjanids. The relationships for families Mullidae and Carangidae were negative, though the Lutjanidae family displayed a positive relationship with SST.

3.4 PRIMARY PRODUCTIVITY

3.4.1 CNMI

Concentrations of the pigment chlorophyll-a are commonly used as an index of phytoplankton biomass that represents primary production, a commonly utilized tool in identifying eutrophication also noted to be among the highest priority fishery ecosystem relationships in the WPR by participants of the Workshop (Islam and Tanaka 2004). In Pacific regions where interannual precipitation and associated coastal runoff are relatively high, the physiochemistry of nearshore reefs is especially impacted from accompanying nutrient input resulting in increased primary production (Ansell et al. 1996).

Long-term changes in regional primary productivity have the potential to change reef fish population abundance due to the susceptibility of these assemblages in shallow areas of coastal reefs to variations in water chemistry, especially when combined with the variability of other environmental parameters like sea surface temperature (Kitiona et al. 2016). For example, it has been suggested that warming ocean temperatures coupled with decreasing environmental productivity led to waning reef fish assemblages in the Southern California Bight, likely due to a reduction in upwelling that isolated nutrients at depth (Roemmich and McGowan 1995). With recent progress in satellite and fluorometric measurements of oceanic surface waters, time series of global and regional primary production estimated using concentrations of chlorophyll-*a* have become increasingly available and can be used for evaluating the impact of environmental productivity on reef fish population abundance and the marine food web in general (Behrenfeld et al. 2006; Messié and Radenac 2006). Data for the study at hand were gathered from the ESA Ocean Colour Climate Change Initiative dataset version 3.1.

Considering the Ocean Colour Climate Change Initiative dataset (v3.1) for CNMI, the time series of fluorometric chlorophyll-*a* concentrations (mg/m³) for the years 1998-2016 in the region is shown in Figure 45. The chlorophyll concentrations had less variability than Guam (CV = 6.28) but was relatively higher in overall average concentration. Unlike Guam, however, pigment levels appeared to have been decreasing over the course of the time series despite the

non-significant nature of the associated regression. Over the 15 years of evaluated data, the average chlorophyll-*a* concentration was 0.049 mg/m³, though the lowest recorded level was seen in 2014 at 0.042 mg/m³ Figure 45.

A time series of fluorometric chlorophyll-*a* concentrations (mg/m³) for the years 1998-2016 in Guam is shown in Figure 46. Pigment concentration in the upper 200 meters had moderate variability over the course of the time series (CV=7.03). Also, there seemed to be a slight increase in pigment concentrations over the course of collected data despite the lack of a significant trend over the same time. The average chlorophyll-*a* concentration over this time was 0.048 mg/m³, with the highest recorded levels being observed in 2005 at 0.055 mg/m³ and the lowest occurring earlier in 2002 (0.042 mg/m³; Figure 46).

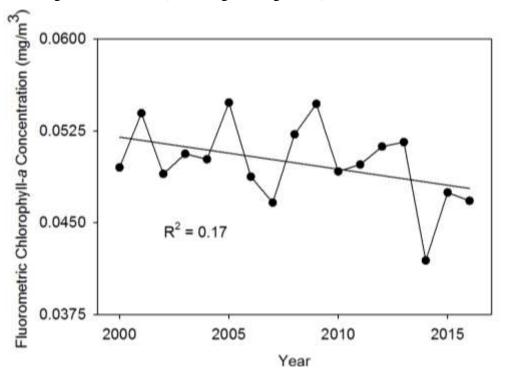


Figure 45. Time series of fluorometric chlorophyll-a concentrations (mg/m^3) around the CNMI from 1998-2016 (CV=6.28)

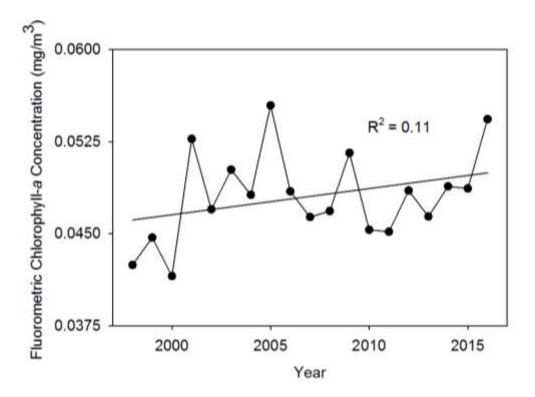


Figure 46. Time series of fluorometric chlorophyll-a concentrations (mg/m³) around Guam from 1998-2016 (CV=7.03)

3.4.1.1 Evaluating relationship for entire reef fishery

A plot showing the relationship between these same chlorophyll levels and catch time series from the recreational coral reef fishery in the CNMI from 2000-2016 is depicted in Figure 47. Catch, again, was even more variable than the environmental data evaluated (CV=19.4) and was at about the same levels as Guam. Total annual catch in the fishery has been decreasing over the last decade and a half despite a spike in catch during 2013 that gave the maximum observed annual catch over this time series (~338,000 kg). The levels of current catch (i.e., for 2014-2016) are the lowest for the entirety of the recreational fishery over the past decade and a half (~165,000 kg; Figure 47).

In pattern with the analyses completed for Guam, linear regressions and correlation analyses were conducted for the time series of the CNMI recreational coral reef fishery catch (with phase lag) with fluorometric chlorophyll-a concentrations (mg/m³) gathered for the 15 years between 2000-2014. The chlorophyll-a concentrations and total annual catch for the all harvested taxa had a positive relationship between 2000 and 2014, though the relationship was far from being considered statistically significant (r = 0.32, p = 0.25; Table 72; Figure 48). Though not significant, the regression was extrapolated to determine that, following this pattern, every increase of 0.01 mg/m³ in chlorophyll-a concentration would cause increase by nearly 62,000 kg two years later for all the CNMI recreational reef fishery (R^2 =0.11, P = 0.25; Figure 48).

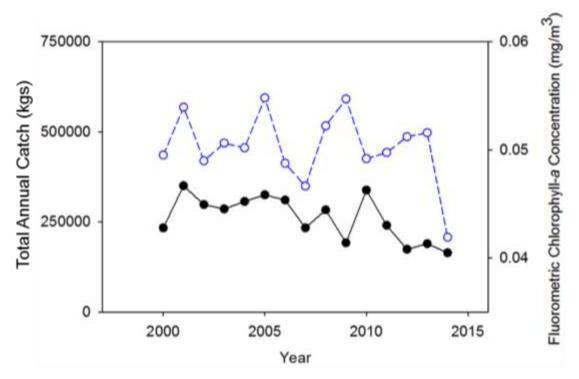


Figure 47. Comparison of the CNMI recreational reef fish catch (kg; black) from creel survey records with two years of time lag (t+2 years) and fluorometric chlorophyll-a concentrations (mg/m³; blue) from 2000-2014 (r = 0.32)

Table 72. Correlation coefficients (r) from comparisons of time series of the CNMI recreational coral reef fishery annual catch (kg) and fluorometric chlorophyll-a concentrations (mg/m³) from 2000-2014

| Taxa Code | Total Catch | LUTJ | LETH | CARA | ACAN | SERR | SIGA | SCAR | MULL | MUGI | LABR | HOLO | BALI |
|-----------|-------------|-------|-------|------|------|------|-------|-------|------|-------|-------|------|------|
| n = 15 | | | | | | | | | | | | | |
| p | 0.25 | 0.47 | 0.14 | 0.67 | 0.37 | 0.09 | 0.72 | 0.80 | 0.99 | 0.83 | 0.83 | 0.10 | 0.72 |
| r | 0.32 | -0.20 | -0.04 | 0.12 | 0.25 | 0.45 | -0.10 | -0.07 | 0.00 | -0.06 | -0.06 | 0.44 | 0.10 |
| R^2 | 0.11 | 0.04 | 0.00 | 0.02 | 0.06 | 0.20 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.20 | 0.01 |

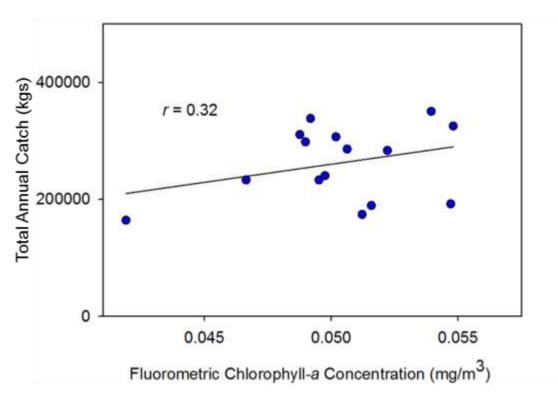


Figure 48. Linear regression between total annual catch (kg) phase lag (t+2 years) and fluorometric chlorophyll-a concentrations (mg/m³) in CNMI from 2000-2014

3.4.1.2 Evaluating relationship for dominant taxa

Out of the many linear regressions completed for catch time series of dominant taxa in the CNMI's recreational coral reef fishery, none of them were determined to be significantly related to the recorded chlorophyll-a concentrations from the same area (Table 72). Of the 12 analyzed groups, the three with the strongest (non-significant) relationship with local chlorophyll concentrations were the Serranids, the Acanthurids, and the Holocentrids ($R^2 = 0.20, 0.20, 0.06$, respectively; Figure 49a-c). It is interesting to note that, unlike Guam, the overall relationship between pigment concentration and catch for the entirety of the reef fishery in the region was positive, though non-significant (r = 0.32, p = 0.25), and the strongest determined associations among the analyzed taxa were all positive as well (Table 72).

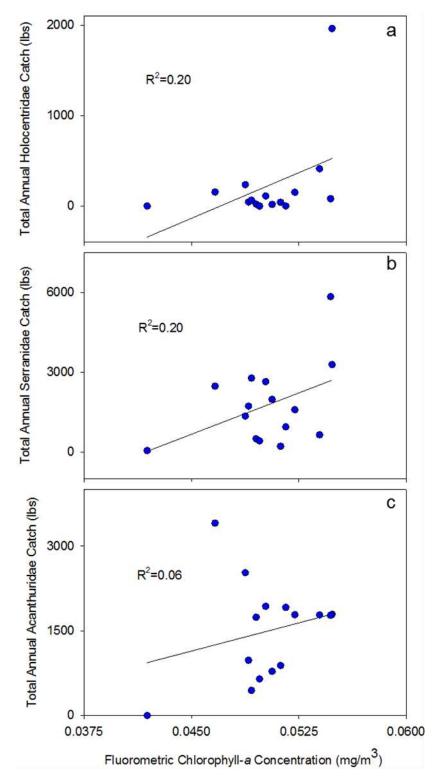


Figure 49. Linear regressions showing the three top correlations between total annual catch (kg) for the CNMI from creel survey records with phase lag (t+2 years) and fluorometric chlorophyll-a concentrations (mg/m³) for (a) Holocentrids, (b) Serranids, and (c)

Acanthurids from 2000–2014

3.4.2 **Guam**

3.4.2.1 Evaluating relationship for entire reef fishery

A plot depicting the comparison of the fluorometric chlorophyll-*a* concentrations and recreational coral reef fishery catch time series from 1998 - 2014 in Guam is shown in Figure 50. Catch levels were relatively variable over the course of the time series when considering the variation in pigment levels (CV=26.2; Figure 50). A gradual drop in total annual catch was observed starting from 1998 before stabilizing in the late 2000s, where recorded catch decreased to approximately a quarter million. and rose back up to over half a million kilograms in more recent years; it is of note that the minimum catch and maximum chlorophyll concentration depicted in this plot both occurred in the year 2005 (Figure 50).

Linear regressions and correlation analyses were conducted for the time series of the Guam recreational coral reef fishery catch (with phase lag) with fluorometric chlorophyll-a concentrations (mg/m³) gathered from the Ocean Colour Climate Change Initiative dataset (v3.1) for the 17 years between 1998 and 2014. It was found that the chlorophyll concentrations and total annual catch for all harvested taxa had a negative relationship between 1989 and 2015, though it was slightly over the threshold of significance (r = -0.45, p = 0.02; Table 73; Figure 51). The association was statistically significant, and it was determined that for every increase of 0.01 mg/m³ in chlorophyll-a concentration, catch would approximately decrease by 180,000 kg after two years all of the Guam recreational fishery ($R^2 = 0.20$, p = 0.02; Table 73; Figure 51).

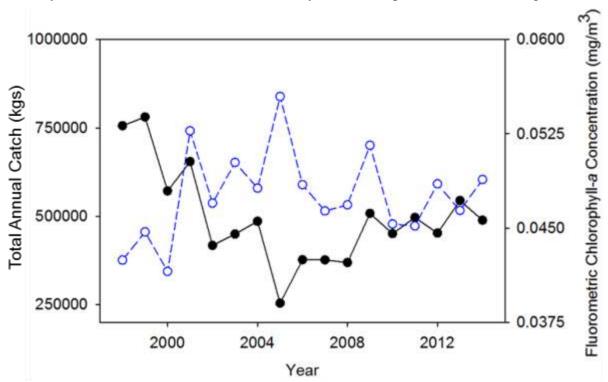


Figure 50. Comparison of Guam recreational reef fish catch for shore-and boat-based creel survey records (kg; black) with two years of time lag (t+2 years) and fluorometric chlorophyll-*a* concentrations (mg/m³; blue) from 1998-2014

Table 73. Correlation coefficients (r) from comparisons of time series of for shore-and boat-based creel survey records in Guam (kg) and fluorometric chlorophyll-a concentrations

(mg/m 3) for 12 top taxa harvested from 1998 - 2014. Significant correlations are indicated in bold (α =0.05)

| Taxa Code | Total Catch | LUTJ | LETH | CARA | ACAN | SERR | SIGA | SCAR | MULL | MUGI | LABR | HOLO | BALI |
|-----------|-------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| n = 17 | | | | | | | | | | | | | |
| P | 0.07 | 0.62 | 0.16 | 0.73 | 0.44 | 0.51 | 0.17 | 0.42 | 0.08 | 0.04 | 0.47 | 0.21 | 0.03 |
| r | -0.45 | -0.13 | -0.36 | -0.09 | -0.20 | -0.17 | -0.35 | -0.21 | -0.43 | -0.50 | -0.19 | -0.32 | -0.53 |
| R^2 | 0.20 | 0.02 | 0.13 | 0.01 | 0.04 | 0.03 | 0.12 | 0.04 | 0.19 | 0.25 | 0.03 | 0.11 | 0.28 |

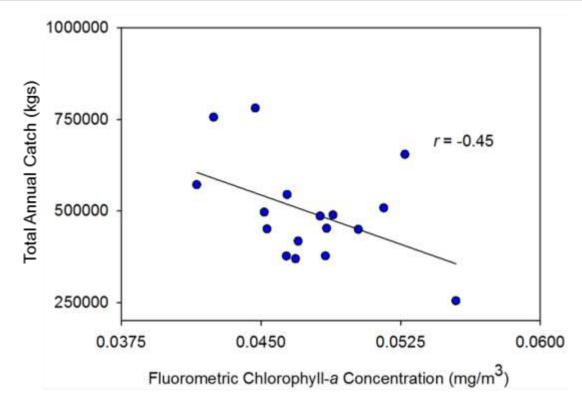


Figure 51. Linear regression between total annual catch (kg) for Guam shore-and boat-based creel survey records with phase lag (t+2 years) and fluorometric chlorophyll-*a* concentrations (mg/m³) from 1998-2014

3.4.2.2 Evaluating relationship for dominant taxa

The several linear regression and correlation analyses performed for time series of catch on the taxa level of Guam's recreational reef fishery showed that for dominant taxa in the fishery, and only two of the 12 analyzed groups had statistically significant relationships with local chlorophyll concentrations: the Balistids and the Mugilids (Table 73). The relationship between catch of species in the Balistidae group and chlorophyll concentration was shown to have negatively significant relationship such that for every increase of 0.01 mg/m³ in chlorophyll-a concentration, catch would drop by more than 1,700 kg two years later when harvesting members of the Balistidae family (R^2 =0.28, p = 0.03; Table 73; Figure 52a). The relationship between catch of members of the Mugilidae group and chlorophyll concentration was also shown to be negatively significant, but to a lesser degree. With a rise of 0.01 mg/m³ in chlorophyll-a levels, recreational catch of the Mugilids would decrease by approximately over 4,600 kg after two years for the group (R^2 =0.25, p = 0.04; Table 73; Figure 52b;). The next strongest relationship as determined by the regressions was not significant but was similarly negative

(Mullidae; R²=0.19, p=0.08; Table 73; Figure 52c); all four of these potential fishery ecosystem relationships, however, were positive.

In the CNMI, there were no statistically significant relationships discovered between chlorophyll concentrations and any of the 12 prevalent taxa evaluated in this study, nor to the total fishery annual catch in its entirety. The lack of identifiable associations could have been attributed to the relatively short time series of data available for comparison at 15 years. While there were several families observed that had relationships on the cusp of being deemed significant according to resulting coefficients of determination, such as Serranidae and Holocentridae, they were positively associated.

In summary for Guam, it was determined that there existed a negatively significant relationship between reef recreational catch and fluorometric chlorophyll-*a* concentrations (mg/m³) from the Ocean Colour Climate Change Initiative dataset (v3.1) for the entirety of the fishery. For every increase of 0.01 mg/m³ in chlorophyll-*a* concentration, catch would approximately decrease by 180,000 kg across all harvested taxa two years later. Potential statistically significant fishery ecosystem relationships were also observed for the Balistidae and Mugilidae groups, where the catch of each group would decrease by approximately 1,700 and 4,600 kg, respectively, given two years of phase lag with a similar increase in fluorometric chlorophyll.

Uncertainty levels were relatively high in evaluations including chlorophyll-a concentrations due to the nature of incorporating phase lag and not smoothing the catch data. The largest issue in performing comparison analyses between catch from reef fisheries in the Mariana Archipelago and fluorometric chlorophyll-a concentrations was the relatively short time series (i.e., small sample size). Robust, homogenous time series highlighting interdecadal patterns in these regions were difficult to obtain due to time series merging several sources of chlorophyll concentration to elongate the range of continuous data. For example, the ESA's OCC CCI dataset only permitted the use of less than two decades of data when evaluating the territories with the incorporation of phase lag. The length of the applied lag has a large impact in the patterns observed, so the relatively short extent of the available time series may obfuscate some of the identified relationships.

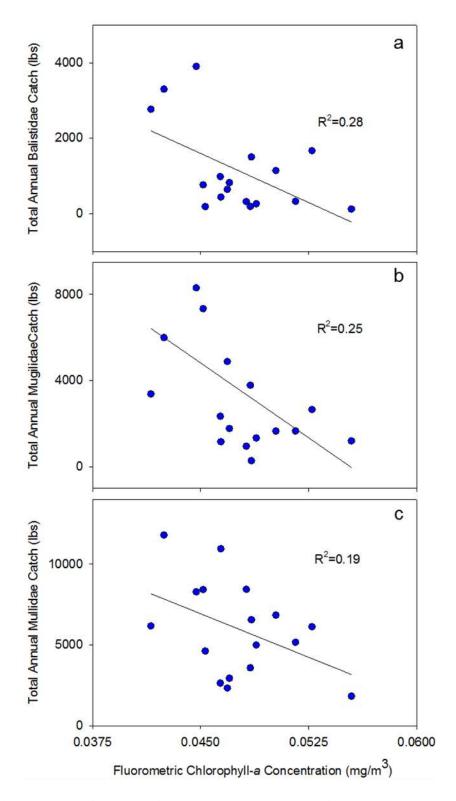


Figure 52. Linear regressions showing the three top correlations between total annual catch (kg) for Guam for shore-and boat-based creel survey records with phase lag (t+2 years) and fluorometric chlorophyll-a concentrations (mg/m³) for (a) Balistidae, (b) Mugilidae, and (c) Mullidae from 1998–2014.

3.5 MULTIVARIATE ASSESSMENTS OF OTHER ECOSYSTEM VARIABLES

3.5.1 Non-metric Multidimensional Scaling

There were several other prioritized fishery ecosystem relationships for coral reefs in the Mariana Archipelago involving environmental parameters that were not to be addressed in this initial evaluation including: the Oceanic Niño Index (ONI), the Pacific Decadal Oscillation (PDO), sea level height, pH, dissolved oxygen, and salinity. Further descriptions of these climate and oceanic indicators are available in Section 2.5. Sea surface height data were aggregated from the Ocean Service, Tides, and Currents, and Sea Level database operated (NOAA/NOS/CO-OPS). Basin-wide data ONI were taken from NOAA's Nation Centers for Environmental Information- Equatorial Pacific Sea Surface Temperature Database (CPC 2015). Similarly, PDO data were obtained from NOAA's Earth System Research Laboratory Physical Sciences Division originally derived from OI.v1 and OI.v2 SST parameters (NOAA PDO). Salinity data for the Marianas were gathered from Simple Ocean Data Assimilation (SODA) version 3.3.1 (Carton and Giese 2008). Rainfall estimates were obtained through the National Weather Service in the Mariana Archipelago (NWS-G).

Non-metric multidimensional scaling (NMS), a form of multivariate analysis that orders sample units along synthetic axes to reveal patterns of composition and relative abundance (Peck 2016), is most commonly utilized when looking to identify patterns in heterogeneous species response data (Peck 2016). For this study, NMS was used to help identify associations between coral reef fishery parameters and environmental factors using the program PCORD 7. To ensure the same length of time series for all catch and environmental variables considered, data was analyzed from 1989-2015 to allow for the inclusion of more parameters (e.g., pH) for which longer-term time series were unavailable. The generated axes represent the best fit of patterns of redundancy in the catch data used as input, and the resulting ordination scores are a rank-order depiction of associations in the original dataset.

NMS produces robust results even in the presence of outliers by avoiding parametric and distributional assumptions (Peck 2016). The only assumption to be met in NMS is that the relationship between the original rank ordered distances between sample units and the reduced distances in the final solution should be monotonic; that is, the slope of the association between the two is flat or positive, as determined by the stress statistic. In the most general terms, interpretable and reliable ordination axes have stress less than 10 up to 25 for datasets with large sample size, but large stress scores (i.e., greater than 30) may suggest that the final ordination results have little association with the original data matrix. Additionally, NMS ordination scores vary depending on the number of dimensions/axes designated to be solved (Peck 2016). Dimensionality (i.e., number of axes for the final solution) for each test was identified though PCORD result recommendations based on final stress being lower than that for 95% of randomized runs (i.e., p < 0.05). Tau is a statistic that represents the rank correlations of the ordination scores to the original data matrices and was used to identify explanatory variables with associations to the ordination axes. For the test, data from 13 species/taxa groups from 1989 - 2015 (27 years) were included along with 10 variables of environmental data collected during the same time period.

3.5.1.1 CNMI

The resulting ordination scores from the NMS analysis performed on boat-based expanded creel survey catch records and the previously mentioned environmental parameters recommended a

one dimensional solution, which accounts for 87.2% of the cumulated variance observed in the CNMI boat-based creel survey data. The NMS final stress was morderate for the real runs (13.9), but low relative to stress from the randomization runs (31.0; Figure 53. NMS scree plot showing the stress test to determine dimensionality for the final solution for the CNMI multivariate analysis). The final ordination scores for the families considered were scaled on a gradient relative to the individual ordination axis, the overlying environmental joint biplot is situated to the left of the final ordination points (Figure 53).

The only environmental parameter included in this analysis that displayed a significant relationship with the lone axis was PDO, though that assoication was negative. (tau = -0.47), Although this NMS run was not able to identify any other environmental parameters significantly correlated to the ordination axis, additionally relatively strong associations exist between sea level height (tau = 0.33) and pH (-0.31; Figure 54). Replicate NMS runs had similar stress levels for the final generated result.

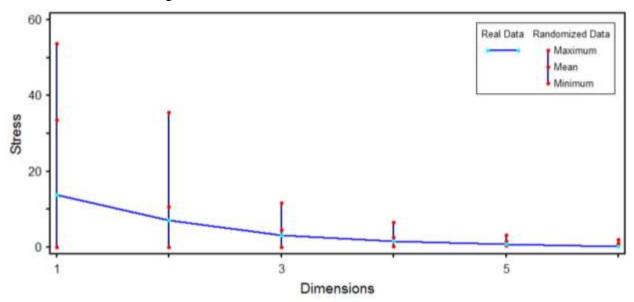


Figure 53. NMS scree plot showing the stress test to determine dimensionality for the final solution for the CNMI multivariate analysis; a one-axis solution was recommended



Figure 54. One-dimensional scatterplot overlaid with a joint biplot depicting ordination scores resulting from an NMS analysis on creel survey expanded catch data and prominent environmental parameters in the CNMI from 2000-2014

3.5.1.2 Guam

The Guam NMS identified two orthogonal axes for the final solution that accounted for 93.6% of the cumulative observed variance in shore- and boat-based creel survey data from Guam. The

final stress for the Guam NMS barely less than 10, though it was notable lower than the average final stress from randomizations (14.2; Figure 55). A majority of the families were clustered in ordination space, with the notable exception of Carangidae (Figure 56).

The final ordination scores for the Guam NMS did not show any environmental parameters with a statistically significant correlation to the first axis ($r^2 = 0.62$; Figure 56). SST (tau = -0.50) and SSTA (tau = -0.50) were both negatively associated with the Axis 2 ($r^2 = 0.32$), and pH had a significantly positive relationship with the axis (tau = 0.56). Additionally, Axis 2 was shown to also be negatively associated with pH (tau = -0.37; Figure 56). Replicate NMS runs had similar stress levels for the final generated result.

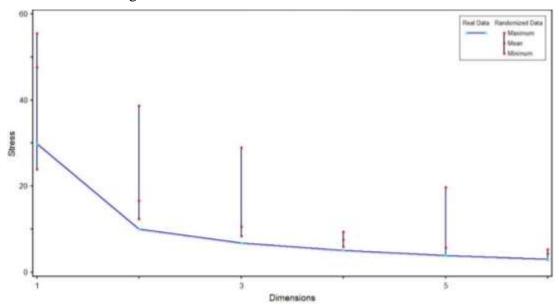


Figure 55. NMS scree plot showing the stress test to determine dimensionality for the final solution for the Guam multivariate analysis; two-axis solution was recommended

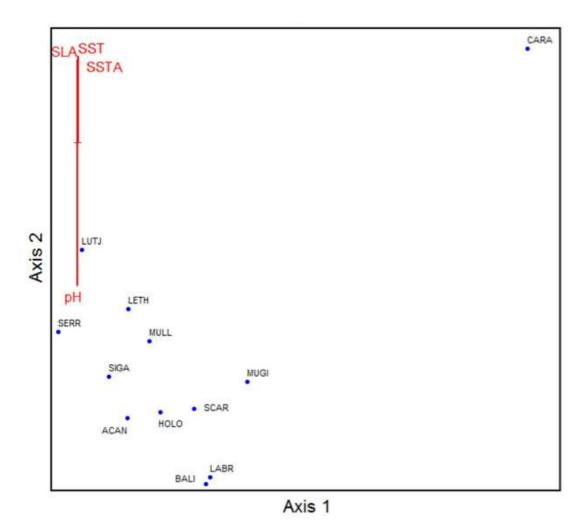


Figure 56. Two-dimensional scatterplot overlaid with a joint biplot depicting ordination scores resulting from an NMS analysis on creel survey expanded catch data and prominent environmental parameters in Guam from 1989-2014

Ultimately, stress values for all analyses were relatively low, suggesting that the generated ordination scores were robust and useful for interpretation relative to the ordination axes. Nearly all included environmental parameters had a statistically significant relationship with at least one ordination axis in at least one of the final solutions, suggesting that these parameters likely intertwine in complicated processes to produce observed impacts on coral reef fisheries in the U.S. Western Pacific. Though a fishery ecosystem relationship may have not been explicitly identified in NMS runs of this preliminary evaluation, it does not preclude the possibility that an association may still exist.

3.6 RECENT RELEVANT ABSTRACTS

In this section, abstracts from primary journal articles published in 2022 and relevant to data integration are compiled. Collecting the abstracts of these articles is intended to further the goal of this section being used to guide adaptive management.

Arostegui MC, Gaube, P, Woodworth-Jefcoats PA, et al. 2022. Anticyclonic eddies aggregate pelagic predators in a subtropical gyre. Nature (2022) https://doi.org/10.1038/s41586-022-05162-6.

Ocean eddies are coherent, rotating features that can modulate pelagic ecosystems across many trophic levels. These mesoscale features, which are ubiquitous at mid-latitudes 1, may increase productivity of nutrient-poor regions, accumulate prey and modulate habitat conditions in the water column. However, in nutrient-poor subtropical gyres—the largest marine biome—the role of eddies in modulating behaviour throughout the pelagic predator community remains unknown despite predictions for these gyres to expand and pelagic predators to become increasingly important for food security. Using a large-scale fishery dataset in the North Pacific Subtropical Gyre, we show a pervasive pattern of increased pelagic predator catch inside anticyclonic eddies relative to cyclones and non-eddy areas. Our results indicate that increased mesopelagic prey abundance in anticyclone cores may be attracting diverse predators, forming ecological hotspots where these predators aggregate and exhibit increased abundance. In this energetically quiescent gyre, we expect that isolated mesoscale features (and the habitat conditions in them) exhibit primacy over peripheral submesoscale dynamics in structuring the foraging opportunities of pelagic predators. Our finding that eddies influence coupling of epi- to mesopelagic communities corroborates the growing evidence that deep scattering layer organisms are vital prey for a suite of commercially important predator species and, thus, provide valuable ecosystem services.

Asner GP, Vaughn NR, Martin RE, Foo SA, Heckler J, Neilson BJ, Gove JM. 2022. Mapped coral mortality and refugia in an archipelago-scale marine heat wave. Proceedings of the National Academy of Sciences. 119(19) https://doi.org/10.1073/pnas.2123331119.

Corals are a major habitat-building life-form on tropical reefs that support a quarter of all species in the ocean and provide ecosystem services to millions of people. Marine heat waves continue to threaten and shape reef ecosystems by killing individual coral colonies and reducing their diversity. However, marine heat waves are spatially and temporally heterogeneous, and so too are the environmental and biological factors mediating coral resilience during and following thermal events. This combination results in highly variable outcomes at both the coral bleaching and mortality stages of every event. This, in turn, impedes the assessment of changing reef-scale patterns of thermal tolerance or places of resistance known as reef refugia. We developed a large-scale, high-resolution coral mortality monitoring capability based on airborne imaging spectroscopy and applied it to a major marine heat wave in the Hawaiian Islands. While water depth and thermal stress strongly mediated coral mortality, relative coral loss was also inversely correlated with preheat-wave coral cover, suggesting the existence of coral refugia. Subsequent mapping analyses indicated that potential reef refugia underwent up to 40% lower coral mortality compared with neighboring reefs, despite similar thermal stress. A combination of human and environmental factors, particularly coastal development and sedimentation levels, differentiated resilient reefs from other more vulnerable reefs. Our findings highlight the role that coral

mortality mapping, rather than bleaching monitoring, can play for targeted conservation that protects more surviving corals in our changing climate.

Boland RC, Hyrenbach KD, DeMartini EE, Parrish FA, Rooney JJ. 2022. Quantifying mesophotic fish assemblages of Hawai`i`s Au`au channel: associations with benthic habitats and depth. Frontiers in Marine Science. Volume 8:1990. https://doi.org/10.3389/fmars.2021.785308.

Mesophotic reefs (30–150 m) occur in the tropics and subtropics at depths beyond most scientific diving, thereby making conventional surveys challenging. Towed cameras, submersibles, and mixed-gas divers were used to survey the mesophotic reef fish assemblages and benthic substrates of the Au'au Channel, between the Hawaiian Islands of Maui and Lana'i. Nonparametric multivariate analysis: Non-metric Multidimensional Scaling (NMDS), Hierarchical Cluster Analysis (HCA), Multi-Response Permutation Procedure (MRPP), and Indicator Species Analysis (ISA) were used to determine the association of mesophotic reef fish species with benthic substrates and depth. Between 53 and 115-m depths, 82 species and 10 genera of fish were observed together with 10 types of benthic substrate. Eight species of fish (Apolemichthys arcuatus, Centropyge potteri, Chaetodon kleinii, Chromis leucura, Chromis verater, Forcipiger sp., Naso hexacanthus, and Parupeneus multifasciatus) were positively associated with increasing depth, Leptoseris sp. coral cover, and hard-bottom cover, and one species (Oxycheilinus bimaculatus) of fish was positively associated with increasing Halimeda sp. algae cover. Fish assemblages associated with rubble were not significantly different from those associated with sand, Montipora coral beds and Leptoseris coral beds, but were distinct from fish assemblages associated with hard bottom. The patterns in the data suggested two depth assemblages, one "upper mesophotic" between 53 and 95 m and the other deeper, possibly part of a "lower mesophotic" assemblage between 96 and 115 m at the edge of the rariphotic and bottomfish complex.

Domokos R. 2022. Seamount effects on micronekton at a subtropical central Pacific seamount. Deep Sea Research Part I: Oceanographic Research Papers, Volume 186: 103829. https://doi.org/10.1016/j.dsr.2022.103829.

Seamounts are globally ubiquitous features with potential for increased biodiversity and biomass, including those of economically important fish. Although their ecological and economical importance is well known, the mechanisms for supporting seamount-associated communities are not well understood. In this study, the effects of an intermediate depth seamount (Cross Seamount) on the micronekton communities, forage for economically important bigeye tuna, are investigated. Relative biomass and composition estimates were calculated from multi-frequency active acoustic data from surveys over 3 years. Mean micronekton biomass was significantly higher than in the ambient environment and its composition differed over the flanks and plateau of Cross Seamount. The effects of the seamount extended ~3.5 km away from the plateau's edge, possibly further below 400 m depth at the flanks. Micronekton occupied the water column from the surface to the 400 m deep plateau with dense aggregations immediately over the bottom at night. During the day, these micronekton migrated both horizontally and downward, occupying depths of 500–700 m, preferably along the upstream flank of the seamount. Descending micronekton from near-surface waters appeared to be temporarily blocked by the topography before swimming below the plateau at the flanks. Mechanisms supporting the increase in

micronekton biomass are uncertain, although hydrographic data support topographic trapping of zooplankton and the existence of transient or semi-permanent Taylor caps.

Giddens J, Kobayashi DR, Mukai GNM, Asher J, Birkeland C, Fitchett M, Hixon MA, Hutchinson M, Mundy BC, O'Malley JM, Sabater M Scott M, Stahl J, Toonen R, Trianni M, Woodworth-Jefcoats PA, Wren JLK, Nelson M. 2022. Assessing the vulnerability of marine life to climate change in the Pacific Islands region. PLoS One,17(7):e0270930. https://doi.org/10.1371/journal.pone.0270930.

Our changing climate poses growing challenges for effective management of marine life, ocean ecosystems, and human communities. Which species are most vulnerable to climate change, and where should management focus efforts to reduce these risks? To address these questions, the National Oceanic and Atmospheric Administration (NOAA) Fisheries Climate Science Strategy called for vulnerability assessments in each of NOAA's ocean regions. The Pacific Islands Vulnerability Assessment (PIVA) project assessed the susceptibility of 83 marine species to the impacts of climate change projected to 2055. In a standard Rapid Vulnerability Assessment framework, this project applied expert knowledge, literature review, and climate projection models to synthesize the best available science towards answering these questions. Here we: (1) provide a relative climate vulnerability ranking across species; (2) identify key attributes and factors that drive vulnerability; and (3) identify critical data gaps in understanding climate change impacts to marine life. The invertebrate group was ranked most vulnerable and pelagic and coastal groups not associated with coral reefs were ranked least vulnerable. Sea surface temperature, ocean acidification, and oxygen concentration were the main exposure drivers of vulnerability. Early Life History Survival and Settlement Requirements was the most data deficient of the sensitivity attributes considered in the assessment. The sensitivity of many coral reef fishes ranged between Low and Moderate, which is likely underestimated given that reef species depend on a biogenic habitat that is extremely threatened by climate change. The standard assessment methodology originally developed in the Northeast US, did not capture the additional complexity of the Pacific region, such as the diversity, varied horizontal and vertical distributions, extent of coral reef habitats, the degree of dependence on vulnerable habitat, and wide range of taxa, including data-poor species. Within these limitations, this project identified research needs to sustain marine life in a changing climate.

Gulland FMD, Baker JD, Howe M, LaBrecque E, Leach L, Moore SE, Reeves RR, Thomas PO. 2022. A review of climate change effects on marine mammals in United States waters: Past predictions, observed impacts, current research and conservation imperatives. Climate Change Ecology. Volume 3: 100054. https://doi.org/10.1016/j.ecochg.2022.100054.

We consider the current evidence of climate change effects on marine mammals that occur in U.S. waters relative to past predictions. Compelling cases of such effects have been documented, though few studies have confirmed population-level impacts on abundance or vital rates. While many of the observed effects had been predicted, some unforeseen and relatively acute consequences have also been documented. Effects often occur when climate-induced alterations are superimposed upon marine mammals' ecological (e.g., predator-prey) relationships or coincident human activities. As they were unanticipated, some of the unpredicted effects of climate change have strained the ability of existing conservation and management systems to respond effectively. The literature is replete with cases suggestive of climate change impacts on

marine mammals, but which remain unconfirmed. This uncertainty is partially explained by insufficient research and monitoring designed to reveal the connections. Detecting and mitigating the impacts of climate change will require some realignment of research and monitoring priorities, coupled with rapid and flexible management that includes both conventional and novel conservation interventions.

Hall R, Parke M. 2022. PIFSC-PIRO ecosystem-based fisheries management workshop April 6-7, 2021 final report. Pacific Islands Fisheries Science Center, PIFSC Administrative Report, H-22-02, 42 p. https://doi.org/10.25923/5f6x-sk11.

NOAA Fisheries strives to maintain and build productive and sustainable fisheries and healthy marine and aquatic ecosystems, as well as to protect threatened and endangered species, through use of an ecosystem-based approach to science and management. To further our goal of implementing ecosystem-based fisheries management (EBFM) in the Pacific Islands region, NOAA Fisheries Pacific Islands Fisheries Science Center (PIFSC) and Pacific Islands Regional Office (PIRO) held an EBFM Workshop on April 6 & 7, 2021.

Huntington B, Vargas-Angel B, Couch CS, Barkley HC, Abecassis M. 2022. Oceanic productivity and high-frequency temperature variability -not human habitation- supports calcifier abundance on central Pacific coral reefs. Frontiers in Marine Science. 9:1075972. https://doi.org/10.3389/fmars.2022.1075972.

Past research has demonstrated how local-scale human impacts—including reduced water quality, overfishing, and eutrophication—adversely affect coral reefs. More recently, globalscale shifts in ocean conditions arising from climate change have been shown to impact coral reefs. Here, we surveyed benthic reef communities at 34 U.S.-affiliated Pacific islands spanning a gradient of oceanic productivity, temperature, and human habitation. We re-evaluated patterns reported for these islands from the early 2000s in which uninhabited reefs were dominated by calcifiers (coral and crustose coralline algae) and thought to be more resilient to global change. Using contemporary data collected nearly two decades later, our analyses indicate this projection was not realized. Calcifiers are no longer the dominant benthic group at uninhabited islands. Calcifier coverage now averages 26.9% ± 3.9 SE on uninhabited islands (compared to 45.18% in the early 2000s). We then asked whether oceanic productivity, past sea surface temperatures (SST), or acute heat stress supersede the impacts of human habitation on benthic cover. Indeed, we found variation in benthic cover was best explained not by human population densities, but by remotely sensed metrics of chlorophyll-a, SST, and island-scale estimates of herbivorous fish biomass. Specifically, higher coral and CCA cover was observed in more productive waters with greater biomass of herbivores, while turf cover increased with daily SST variability and reduced herbivore biomass. Interestingly, coral cover was positively correlated with daily variation in SST but negatively correlated with monthly variation. Surprisingly, metrics of acute heat stress were not correlated with benthic cover. Our results reveal that human habitation is no longer a primary correlate of calcifier cover on central Pacific island reefs, and highlight the addition of oceanic productivity and high-frequency SST variability to the list of factors supporting reef builder abundance.

Huntington B, Weible R, Halperin A, et al. 2022. Early successional trajectory of benthic community in an uninhabited reef system three years after mass coral bleaching. Coral Reefs (2022) https://doi.org/10.1007/s00338-022-02246-7.

Severe thermal stress events occurring on the backdrop of globally warming oceans can result in mass coral mortality. Tracking the ability of a reef community to return to pre-disturbance composition is important to inform the likelihood of recovery or the need for active management to conserve these ecosystems. Here, we quantified annual, temporal changes in the benthic communities for the three years following mass coral mortality at Jarvis Island—an uninhabited island in the Pacific Remote Islands Marine National Monument. While Jarvis experienced catastrophic coral mortality in 2015 due to heat stress resulting from the 2015/16 El Niño. significant annual shifts were documented in the benthic community in the three years postdisturbance. Macroalgal and turf dominance of the benthos was temporary—likely reflecting the high biomass of herbivorous reef fishes post-bleaching—giving way to calcifiers such as crustose coralline algae and *Halimeda*, which may facilitate rather than impede coral recovery. By 2018, indications of recovery were detectable in the coral community itself as juvenile densities increased and stress-tolerant genera, such as *Pavona*, exceeded their pre-disturbance densities. However, densities of *Montipora* and *Pocillopora* remain low, suggesting recovery will be slow for these formerly dominant taxa. Collectively, the assemblage and taxon-specific shifts observed in the benthic and coral community support cautious optimism for the potential recovery of Jarvis Island's coral reefs to their pre-disturbance state. Continued monitoring will be essential to assess whether reassembly is achieved before further climate-related disturbance events affect this reef system.

Iwane M, Hospital J. 2022. Hawai'i fishing communities' vulnerability to climate change: Climate vulnerable species and adaptive capacity. U.S. Dept. of Commerce, NOAA Technical Memorandum NOAA-TM-NMFS-PIFSC-136, 34 p. https://doi.org/10.25923/4vvb-pv29.

In this report, we propose a framework that could be useful to select candidate communities from the main Hawaiian Islands for future qualitative research on the vulnerability of fishing communities to climate change. We adopted the IPCC framework (2001) that defines climate change vulnerability as a function of sensitivity (S), exposure (E), and adaptive capacity (AC). We tested and finalized community selection criteria based on available quantitative data and CSVIs relevant to MHI communities' social and climate change vulnerability.

Kinney MJ, Carvalho F, Kai M, Semba Y, Liu KM, Tsai WP, Leonardo CGJ, Horacio HA, Daniel CCL, Teo SLH. 2022. Cluster analysis used to re-examine fleet definitions of North Pacific fisheries with spatiotemporal consideration of blue shark size and sex data. Pacific Islands Fisheries Science Center, PIFSC Working Paper, WP-22-001, 18 p. https://doi.org/10.25923/zet2-sk13.

This study looked at re-examining the North Pacific fleets that have been used for previous assessments of blue shark by investigating the size and sex composition data from observer records, port and scientific samples in greater detail. Our goal is to provide information that can be used by the ISC shark working group to more appropriately define fleet structure for the assessment based on size and sexual composition of the catch. Ultimately, refining fleet structure

within the model with greater consideration for the spatiotemporal characteristics of blue shark catch may help reduce model misspecification in future assessments. We analyzed nearly 600,000 individual records of blue shark size and sex information divided across 240 5 x 5° grid cells covering the North Pacific. A clustering approach was taken to discern areas with related size and sex compositions. Results suggested four distinct clusters, where Clusters 1 and 4 (made up primarily of smaller immature animals) predominate in the catch at higher latitudes (north of ~25°N), especially in the eastern and western edges of the North Pacific (waters nearer the coasts). While Cluster 2 (mature males and females) and Cluster 3 (mostly males, both mature and immature) predominate in a band from ~ 20°N to near the equator. During fall and winter (seasons 1 and 4) this band of mature animals expands north in central Pacific waters, loosely around Hawaii, as high up as ~40°N. We suggest that this work, along with several other studies carried out by various members of the ISC shark working group over the years, be used to better define the fleets used in future assessments of blue sharks in the North Pacific

Kleiber D, Iwane M, Kamikawa K, Leong K, Hospital J. 2022. Pacific Islands Region Fisheries and COVID-19: Impacts and adaptations. U.S. Dept. of Commerce, NOAA Technical Memorandum NOAA-TM-NMFS-PIFSC-130, 36 p. https://doi.org/10.25923/2fpm-c128.

The Pacific Islands Region has experienced a number of unique risks from COVID-19, and the measures put in place to stop its spread. In this report, we detail the impacts of COVID-19 on the Pacific Islands Region fisheries from March 2020 to February 2021, and highlight the adaptations made by the diverse fishers, marketers, and fishing communities of this region. We gathered information from different sources, including publicly available statistics, news reports, government rules, as well as short open-ended phone interviews.

Lisi P, Hogan J.D, Holt G, Moody K, Wren J, Kobayashi D, Blum M, McIntyre P. 2022. Stream and ocean hydrodynamics mediate partial migration strategies in an amphidromous Hawaiian goby. Ecology, e3800. https://doi.org/10.1002/ecy.3800.

Partial migration strategies, in which some individuals migrate but others do not, are widely observed in populations of migratory animals. Such patterns could arise via variation in migratory behaviors made by individual animals, via genetic variation in migratory predisposition, or simply by variation in migration opportunities mediated by environmental conditions. Here we use spatiotemporal variation in partial migration across populations of an amphidromous Hawaiian goby to test whether stream or ocean conditions favor completing its life cycle entirely within freshwater streams rather than undergoing an oceanic larval migration. Across 35 watersheds, microchemical analysis of otoliths revealed that most adult Awaous stamineus were freshwater residents (62% of n = 316 in 2009, 83% of n = 274 in 2011), but we found considerable variation among watersheds. We then tested the hypothesis that the prevalence of freshwater residency increases with the stability of stream flows and decreases with the availability of dispersal pathways arising from ocean hydrodynamics. We found that streams with low variation of daily discharge were home to a higher incidence of freshwater residents in each survey year. The magnitude of the shift in freshwater residency between survey years was positively associated with predicted interannual variability in the success of larval settlement in streams on each island based on passive drift in ocean currents. We built on these findings by developing a theoretical model of goby life history to further evaluate whether

mediation of migration outcomes by stream and ocean hydrodynamics could be sufficient to explain the range of partial migration frequency observed across populations. The model illustrates that the proportion of larvae entering the ocean and differential survival of freshwater-resident versus ocean-going larvae are plausible mechanisms for range-wide shifts in migration strategies. Thus, we propose that hydrologic variation in both ocean and stream environments contributes to spatiotemporal variation in the prevalence of migration phenotypes in *A. stamineus*. Our empirical and theoretical results suggest that the capacity for partial migration could enhance the persistence of metapopulations of diadromous fish when confronted with variable ocean and stream conditions.

Mazur MD, Tanaka KR, Shank B, Chang J, Hodgson CT, Reardon KM, Friedland KD, Chen Y. 2022. Incorporating spatial heterogeneity and environmental impacts into stock-recruitment relationships for Gulf of Maine lobster. ICES Journal of Marine Science.0:1-11. https://doi.org/10.1093/icesjms/fsab266.

Functional stock-recruitment relationships (SRRs) are often difficult to quantify and can differ over space. Additionally, climate change adds to the complexity of recruitment dynamics. This paper's aim was to incorporate spatial heterogeneity and environmental effects on productivity in SRRs with American lobster in the Gulf of Maine (GOM) as a case study. GOM lobster recruitment has substantially increased since the mid-2000s, due to improved survival rates of pre-recruits and increased spawning stock biomass (SSB). GOM bottom water temperatures have increased at a rate of 0.2°C per decade, which caused lobster settlement area to expand and improved survival rates. We first estimated local SSB using bottom trawl survey data and a geostatistical model. Using estimated SSB, recruitment data from a ventless trap survey, and an interpolated bottom water temperature field, we developed modified Ricker stock-recruitment models accounting for spatial heterogeneity and temperature impacts with varying coefficient generalized additive models. Results showed that temperature significantly impacted recruitment. Changes in temperature mediated productivity differed between the eastern and western GOM. Our study demonstrated that the incorporation of spatial heterogeneity and environmental effects impacts our understanding of SRRs. These methods can be applied to other species to understand recruitment dynamics influenced by climate change.

Panelo J, Wiegner TN, Colbert SL, Goldberg S, Abaya LM, Conklin E, Couch C, Falinski K, Gove J, Watson L, Wiggins C. 2022. Spatial distribution and sources of nutrients at two coastal developments in South Kohala, Hawai'i. Marine Pollution Bulletin. Volume 174:113143. https://doi.org/10.1016/j.marpolbul.2021.113143.

Nutrient sources to coastal waters with coral reefs are not well-characterized. This study documented spatial distributions of nutrients within coastal waters along two developments with coral reefs, and identified nutrient sources through nutrient mixing plots, $\delta^{15}N$ measurements in macroalgal tissue, and NO_3^- stable isotope mixing models. Nutrients decreased from fresh groundwaters to offshore waters, with some surface waters higher in concentrations than benthic ones. Conservative and non-conservative mixing between fresh and ocean waters occurred, the latter suggestive of local nutrient sources and biological removal. $\delta^{15}N$ in macroalgal tissue and NO_3^- concurred that fresh groundwater, ocean water, and fertilizers were dominant nutrient sources. Benthic salinity and $NO_3^- + NO_2^-$ concentrations illustrated that submarine groundwater discharge delivered nutrients to reefs in pulses ranging from minutes to days. Information

generated from this study is imperative for developing management actions to improve water quality and make coral reefs more resilient to stressors.

Smith J, Halperin A, Barkley H. 2022. A 'perfect storm' of cumulative and acute heat stress, and a warming trend, lead to bleaching events in Tutuila, American Samoa. U.S. Dept. of Commerce, NOAA Technical Memorandum NOAA-TM-NMFS-PIFSC-129, 52 p. https://doi.org/10.25923/yphg-pq04.

To better understand vertical thermal structure of reefs at depth and identify predictors of mass bleaching events using high frequency time series data, we used long-term (2012–2018) in situ temperature data collected at multiple reefs and depths around the island of Tutuila in American Samoa. Located in the central South Pacific, Tutuila is 1 of 5 volcanic islands and 2 atolls that comprise American Samoa. Lying just a few kilometers from shore, Tutuila contains shallow fringing reefs and a deep offshore bank (Birkeland et al. 2008). American Samoa experienced severe bleaching in 1994, 2003, 2015 and 2017 (Coward et al. 2020). The objectives of our study are to (1) conduct a time series analysis on in situ temperature data (2012–2018) and calculate heating metrics and (2) determine whether heating metrics predicted coral bleaching prevalence during the 2015 bleaching event.

Tanaka KR, Schmidt AL, Kindinger TL, Whitney JL, Samson JC. 2022. Spatiotemporal assessment of *Aprion virescens* density in shallow main Hawaiian Islands waters, 2010-2019. U.S. Dept. of Commerce, NOAA Technical Memorandum NOAA-TM-NMFS-PIFSC-132, 33 p. https://doi.org/10.25923/f24q-k056.

The Magnuson-Stevens Fishery Conservation and Management Act of 1996 directs regional fishery management councils and the National Marine Fisheries Service (NMFS) to identify and describe "essential fish habitat (EFH)" for all federally managed species to ensure conservation and sustainable management of living marine resources. This report summarizes the statistically-derived density patterns of *Aprion virescens* in shallow coastal waters of the main Hawaiian Islands (MHIs) from 2010 to 2019.

Tanaka KR, Van Houtan KS. 2022. The recent normalization of historical marine heat extremes. PLOS Climate. 1(2): e0000007. https://doi.org/10.1371/journal.pclm.0000007.

Climate change exposes marine ecosystems to extreme conditions with increasing frequency. Capitalizing on the global reconstruction of sea surface temperature (SST) records from 1870-present, we present a centennial-scale index of extreme marine heat within a coherent and comparable statistical framework. A spatially (1° × 1°) and temporally (monthly) resolved index of the normalized historical extreme marine heat events was expressed as a fraction of a year that exceeds a locally determined, monthly varying 98th percentile of SST gradients derived from the first 50 years of climatological records (1870–1919). For the year 2019, our index reports that 57% of the global ocean surface recorded extreme heat, which was comparatively rare (approximately 2%) during the period of the second industrial revolution. Significant increases in the extent of extreme marine events over the past century resulted in many local climates to have shifted out of their historical SST bounds across many economically and ecologically important marine regions. For the global ocean, 2014 was the first year to exceed the 50% threshold of extreme heat thereby becoming "normal", with the South Atlantic (1998) and Indian (2007)

basins crossing this barrier earlier. By focusing on heat extremes, we provide an alternative framework that may help better contextualize the dramatic changes currently occurring in marine systems.

Winston M, Oliver T, Couch C, Donovan MK, Asner GP, et al. 2022. Coral taxonomy and local stressors drive bleaching prevalence across the Hawaiian Archipelago in 2019. PLOS ONE 17(9): e0269068. https://doi.org/10.1371/journal.pone.0269068.

The Hawaiian Archipelago experienced a moderate bleaching event in 2019—the third major bleaching event over a 6-year period to impact the islands. In response, the Hawai'i Coral Bleaching Collaborative (HCBC) conducted 2,177 coral bleaching surveys across the Hawaiian Archipelago. The HCBC was established to coordinate bleaching monitoring efforts across the state between academic institutions, non-governmental organizations, and governmental agencies to facilitate data sharing and provide management recommendations. In 2019, the goals of this unique partnership were to: 1) assess the spatial and temporal patterns of thermal stress; 2) examine taxa-level patterns in bleaching susceptibility; 3) quantify spatial variation in bleaching extent; 4) compare 2019 patterns to those of prior bleaching events; 5) identify predictors of bleaching in 2019; and 6) explore site-specific management strategies to mitigate future bleaching events. Both acute thermal stress and bleaching in 2019 were less severe overall compared to the last major marine heatwave events in 2014 and 2015. Bleaching observed was highly site- and taxon-specific, driven by the susceptibility of remaining coral assemblages whose structure was likely shaped by previous bleaching and subsequent mortality. A suite of environmental and anthropogenic predictors was significantly correlated with observed bleaching in 2019. Acute environmental stressors, such as temperature and surface light, were equally important as previous conditions (e.g. historical thermal stress and historical bleaching) in accounting for variation in bleaching during the 2019 event. We found little evidence for acclimation by reefs to thermal stress in the main Hawaiian Islands. Moreover, our findings illustrate how detrimental effects of local anthropogenic stressors, such as tourism and urban runoff, may be exacerbated under high thermal stress. In light of the forecasted increase in severity and frequency of bleaching events, future mitigation of both local and global stressors is a high priority for the future of corals in Hawai'i.

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APPENDIX A: LIST OF SPECIES

CNMI AND GUAM MANAGEMENT UNIT SPECIES

1. Bottomfish Multi-species Stock Complex (FSSI)

| DFW Creel Species Code | DAWR Creel Species Code | Species Name | Scientific Name |
|---------------------------|----------------------------------|--------------------------------------|-----------------------------|
| 214 | 32302 | red snapper, silvermouth (lehi) | Aphareus rutilans |
| 112 | 31404 | giant trevally, jack | Caranx ignobilis |
| 111 | 31405 | black trevally, jack | Caranx lugubris |
| 241 | 28941 | lunartail grouper (lyretail grouper) | Variola louti |
| 203 | 32304 | red snapper (ehu) | Etelis carbunculus |
| 210 | 32305 | red snapper (onaga) | Etelis coruscans |
| 350 | 32809 | redgill emperor | Lethrinus rubrioperculatus |
| 253 | 32310 | blueline snapper | Lutjanus kasmira |
| None | 32317 | yellowtail snapper | Pristipomoides auricilla |
| 212 | 32318 | pink snapper (paka) | Pristipomoides filamentosus |
| 209 | 32319 | yelloweye snapper | Pristipomoides flavipinnis |
| 207 | 32320 | pink snapper (kalekale) | Pristipomoides sieboldii |
| 204 | 32321 | flower snapper (gindai) | Pristipomoides zonatus |

CNMI AND GUAM MONITORED ECOSYSTEM COMPONENT SPECIES

1. Species Selected for Monitoring by DFW (CNMI)

| DFW Creel Species Code | Species Name | Scientific Name |
|---------------------------------|------------------------------|------------------------------|
| 380 | lined surgeonfish | Acanthurus lineatus |
| 319 | orangespine unicornfish | Naso lituratus |
| 384 | bluespine unicornfish | Naso unicornis |
| None | redlip parrotfish | Scarus rubroviolaceus |
| 317 | blue-barred parrotfish | Scarus ghobban |
| 353 | thumbprint/blackspot emperor | Lethrinus harak |
| 304 | forktail rabbitfish | Siganus argenteus |
| 370 | yellowstripe goatfish | Mulloidichthys flavolineatus |

2. Species Selected for Monitoring by DAWR (Guam)

| DAWR Creel Species Code | Species Name | Scientific Name |
|----------------------------------|------------------------------|-----------------------|
| 41225 | bluespine unicornfish | Naso unicornis |
| 41305 | scribbled rabbitfish | Siganus spinus |
| 32804 | thumbprint/blackspot emperor | Lethrinus harak |
| 36408 | Pacific slopehead parrotfish | Chlorurus frontalis |
| 28917 | blacktip grouper | Epinephelus fasciatus |
| 31406 | bluefin trevally | Caranx melampygus |
| 32806 | ornate emperor | Lethrinus olivaceus |
| 32308 | flametail snapper | Lutjanus fulvus |
| 36414 | redlip parrotfish | Scarus rubroviolaceus |

3. Species Monitored by Trophic, Taxonomic, and Functional groups

The species presented in Section 2.1 are displayed according to both trophic level and functional group as an effort to foster continued monitoring of ecosystem component species that are no longer categorized as management unit species. These species are monitored according to their ecosystem function as opposed to individually. Monitoring based on these factors allows for a broader outlook on the ecological composition of fish communities in areas of the Western Pacific. For trophic groupings, "H" stands for "Herbivore", "Cor" stands for "Corallivore", "PK" stands for "Planktivore", "MI" stands for "Mobile Invertebrate Feeder", "SI" stands for "Sessile-Invertebrate Feeder, "Om" stands for "Omnivore", and "Pisc" stands for "Piscovore".

| Family | Scientific Name | Trophic Group | Functional Group |
|--------------|---------------------------|------------------|---------------------------|
| Acanthuridae | Naso lituratus | Н | Browsing Surgeons |
| Acanthuridae | Naso tonganus | Н | Browsing Surgeons |
| Acanthuridae | Naso unicornis | Н | Browsing Surgeons |
| Acanthuridae | Naso brachycentron | Н | Browsing Surgeons |
| Acanthuridae | Ctenochaetus cyanocheilus | Н | Mid-Large Target Surgeons |
| Acanthuridae | Ctenochaetus strigosus | Н | Mid-Large Target Surgeons |
| Acanthuridae | Acanthurus nigroris | Н | Mid-Large Target Surgeons |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|-----------------------------|------------------|---------------------------|
| Acanthuridae | Ctenochaetus hawaiiensis | Н | Mid-Large Target Surgeons |
| Acanthuridae | Ctenochaetus striatus | Н | Mid-Large Target Surgeons |
| Acanthuridae | Ctenochaetus marginatus | Н | Mid-Large Target Surgeons |
| Acanthuridae | Acanthurus lineatus | Н | Mid-Large Target Surgeons |
| Acanthuridae | Acanthurus blochii | Н | Mid-Large Target Surgeons |
| Acanthuridae | Acanthurus dussumieri | Н | Mid-Large Target Surgeons |
| Acanthuridae | Acanthurus xanthopterus | Н | Mid-Large Target Surgeons |
| Chaetodontidae | Chaetodon flavocoronatus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon multicinctus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon punctatofasciatus | MI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon mertensii | Н | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon citrinellus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon pelewensis | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon lunulatus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon melannotus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon rafflesii | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon ulietensis | MI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon fremblii | SI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon quadrimaculatus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon meyeri | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon reticulatus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon trifascialis | Cor | Non-PK Butterflyfish |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|-------------------------|------------------|-------------------------|
| Chaetodontidae | Heniochus chrysostomus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon bennetti | MI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon tinkeri | SI | Non-PK Butterflyfish |
| Chaetodontidae | Heniochus varius | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon ornatissimus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon unimaculatus | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon lunula | SI | Non-PK Butterflyfish |
| Chaetodontidae | Forcipiger longirostris | MI | Non-PK Butterflyfish |
| Chaetodontidae | Forcipiger flavissimus | SI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon ephippium | MI | Non-PK Butterflyfish |
| Chaetodontidae | Heniochus monoceros | MI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon auriga | SI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon vagabundus | SI | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon semeion | Н | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodontidae | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Heniochus singularius | Cor | Non-PK Butterflyfish |
| Chaetodontidae | Chaetodon lineolatus | SI | Non-PK Butterflyfish |
| Caracanthidae | Caracanthus typicus | MI | No Group |
| Gobiidae | Eviota sp. | MI | No Group |
| Pomacentridae | Chrysiptera traceyi | Н | No Group |
| Apogonidae | Ostorhinchus luteus | Pk | No Group |
| Caracanthidae | Caracanthus maculatus | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|-----------------|------------------------------------|------------------|------------------|
| Pseudochromidae | Pseudochromis jamesi | MI | No Group |
| Pomacentridae | Chromis acares | Pk | No Group |
| Serranidae | Luzonichthys whitleyi | Pk | No Group |
| Pomacentridae | Pomachromis guamensis | Pk | No Group |
| Pomacentridae | Pomachromis richardsoni | Pk | No Group |
| Gobiidae | Fusigobius duospilus | MI | No Group |
| Pomacentridae | Plectroglyphidodon imparipennis | MI | No Group |
| Microdesmidae | Nemateleotris helfrichi | Pk | No Group |
| Pomacentridae | Chromis leucura | Pk | No Group |
| Syngnathidae | Doryrhamphus excisus | Pk | No Group |
| Pomacentridae | Pomacentrus coelestis | Pk | No Group |
| Clupeidae | Spratelloides delicatulus | Pk | No Group |
| Pomacentridae | Chrysiptera biocellata | Н | No Group |
| Pseudochromidae | Pictichromis porphyreus | MI | No Group |
| Pomacanthidae | Centropyge fisheri | Н | No Group |
| Cirrhitidae | Cirrhitops hubbardi | MI | No Group |
| Gobiidae | Amblyeleotris fasciata | Pk | No Group |
| Pomacentridae | Chromis lepidolepis | Pk | No Group |
| Pomacentridae | Chromis margaritifer | Pk | No Group |
| Pomacentridae | Chromis ternatensis | Pk | No Group |
| Pomacentridae | Chromis viridis | Pk | No Group |
| Pomacentridae | Chrysiptera cyanea | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|-----------------|------------------------------------|------------------|------------------|
| Pomacentridae | Dascyllus aruanus | Pk | No Group |
| Pomacentridae | Dascyllus reticulatus | Pk | No Group |
| Engraulidae | Encrasicholina purpurea | Pk | No Group |
| Pomacentridae | Neopomacentrus metallicus | Pk | No Group |
| Pomacentridae | Chromis amboinensis | Н | No Group |
| Pomacentridae | Chromis iomelas | Н | No Group |
| Pomacentridae | Chrysiptera glauca | Н | No Group |
| Pomacentridae | Chrysiptera taupou | Н | No Group |
| Labridae | Labroides pectoralis | MI | No Group |
| Labridae | Pseudocheilinus hexataenia | MI | No Group |
| Labridae | Pseudocheilinus tetrataenia | MI | No Group |
| Scorpaenidae | Sebastapistes cyanostigma | MI | No Group |
| Labridae | Wetmorella nigropinnata | MI | No Group |
| Pseudochromidae | Pseudochromis sp. | MI | No Group |
| Monacanthidae | Pervagor marginalis | Om | No Group |
| Pomacentridae | Chromis alpha | Pk | No Group |
| Pomacentridae | Plectroglyphidodon phoenixensis | Н | No Group |
| Gobiidae | Amblyeleotris guttata | Pk | No Group |
| Atherinidae | Atherinomorus insularum | Pk | No Group |
| Pomacentridae | Chromis caudalis | Pk | No Group |
| Pomacentridae | Chromis hanui | Pk | No Group |
| Labridae | Cirrhilabrus katherinae | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|--------------------------------|------------------|------------------|
| Microdesmidae | Nemateleotris magnifica | Pk | No Group |
| Apogonidae | Ostorhinchus angustatus | Pk | No Group |
| Serranidae | Pseudanthias bartlettorum | Pk | No Group |
| Tetraodontidae | Canthigaster jactator | Н | No Group |
| Tetraodontidae | Canthigaster janthinoptera | Н | No Group |
| Tetraodontidae | Canthigaster valentini | Н | No Group |
| Pomacanthidae | Centropyge shepardi | Н | No Group |
| Pomacentridae | Chrysiptera brownriggii | Н | No Group |
| Monacanthidae | Oxymonacanthus longirostris | Cor | No Group |
| Cirrhitidae | Amblycirrhitus bimacula | MI | No Group |
| Cirrhitidae | Cirrhitichthys falco | MI | No Group |
| Labridae | Labroides rubrolabiatus | MI | No Group |
| Cirrhitidae | Neocirrhites armatus | MI | No Group |
| Labridae | Pseudojuloides splendens | MI | No Group |
| Apogonidae | Ostorhinchus novemfasciatus | Pk | No Group |
| Labridae | Pteragogus cryptus | MI | No Group |
| Scorpaenidae | Sebastapistes sp. | Pisc | No Group |
| Scorpaenidae | Taenianotus triacanthus | Pisc | No Group |
| Pomacentridae | Amphiprion perideraion | Pk | No Group |
| Pomacentridae | Chromis fumea | Pk | No Group |
| Labridae | Cirrhilabrus jordani | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|----------------------------------|------------------|-------------------------|
| Blenniidae | Ecsenius bicolor | Pk | No Group |
| Blenniidae | Ecsenius midas | Pk | No Group |
| Blenniidae | Ecsenius opsifrontalis | Pk | No Group |
| Pomacentridae | Lepidozygus tapeinosoma | Pk | No Group |
| Blenniidae | Meiacanthus atrodorsalis | Pk | No Group |
| Apogonidae | Ostorhinchus apogonoides | Pk | No Group |
| Pomacentridae | Plectroglyphidodon lacrymatus | Pk | No Group |
| Pomacentridae | Pomacentrus brachialis | Pk | No Group |
| Pomacentridae | Pomacentrus nigriradiatus | Pk | No Group |
| Pomacentridae | Pomacentrus philippinus | Pk | No Group |
| Pomacentridae | Pomacentrus vaiuli | Pk | No Group |
| Serranidae | Pseudanthias dispar | Pk | No Group |
| Serranidae | Pseudanthias hawaiiensis | Pk | No Group |
| Tetraodontidae | Canthigaster bennetti | Н | No Group |
| Pomacanthidae | Centropyge bispinosa | Н | No Group |
| Pomacanthidae | Centropyge heraldi | Н | No Group |
| Pomacanthidae | Centropyge loricula | Н | No Group |
| Blenniidae | Cirripectes obscurus | Н | No Group |
| Blenniidae | Cirripectes polyzona | Н | No Group |
| Blenniidae | Cirripectes sp. | Н | No Group |
| Blenniidae | Cirripectes springeri | Н | No Group |
| Blenniidae | Cirripectes stigmaticus | Н | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|---------------------------------|------------------|------------------|
| Blenniidae | Cirripectes variolosus | Н | No Group |
| Callionymidae | Callionymidae | MI | No Group |
| Labridae | Labroides phthirophagus | MI | No Group |
| Pomacanthidae | Paracentropyge multifasciata | MI | No Group |
| Blenniidae | Plagiotremus ewaensis | MI | No Group |
| Blenniidae | Plagiotremus goslinei | MI | No Group |
| Scorpaenidae | Sebastapistes coniorta | MI | No Group |
| Monacanthidae | Pervagor melanocephalus | Om | No Group |
| Blenniidae | Plagiotremus laudandus | Par | No Group |
| Blenniidae | Plagiotremus rhinorhynchos | Par | No Group |
| Blenniidae | Plagiotremus tapeinosoma | Par | No Group |
| Labridae | Pseudocheilinus ocellatus | MI | No Group |
| Pomacanthidae | Centropyge flavissima & vroliki | Н | No Group |
| Pomacentridae | Amblyglyphidodon curacao | Om | No Group |
| Pomacentridae | Amphiprion melanopus | Pk | No Group |
| Pomacentridae | Chromis agilis | Pk | No Group |
| Gobiidae | Istigobius sp. | Pk | No Group |
| Pomacentridae | Pomacentrus pavo | Pk | No Group |
| Apogonidae | Pristiapogon fraenatus | Pk | No Group |
| Tetraodontidae | Canthigaster epilampra | Н | No Group |
| Tetraodontidae | Canthigaster solandri | Н | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|----------------------------------|------------------|-------------------------|
| Blenniidae | Cirripectes vanderbilti | Н | No Group |
| Pomacentridae | Stegastes albifasciatus | Н | No Group |
| Pomacentridae | Stegastes aureus | Н | No Group |
| Pomacentridae | Stegastes marginatus | Н | No Group |
| Pomacentridae | Plectroglyphidodon dickii | Cor | No Group |
| Cirrhitidae | Paracirrhites xanthus | MI | No Group |
| Monacanthidae | Paraluteres prionurus | MI | No Group |
| Microdesmidae | Microdesmidae | Pk | No Group |
| Scorpaenidae | Sebastapistes ballieui | MI | No Group |
| Apogonidae | Apogon kallopterus | Pk | No Group |
| Pomacentridae | Chromis weberi | Pk | No Group |
| Labridae | Cirrhilabrus exquisitus | Pk | No Group |
| Syngnathidae | Corythoichthys flavofasciatus | Pk | No Group |
| Pomacentridae | Dascyllus albisella | Pk | No Group |
| Microdesmidae | Gunnellichthys curiosus | Pk | No Group |
| Apogonidae | Pristiapogon kallopterus | Pk | No Group |
| Serranidae | Pseudanthias olivaceus | Pk | No Group |
| Ptereleotridae | Ptereleotris heteroptera | Pk | No Group |
| Ptereleotridae | Ptereleotris zebra | Pk | No Group |
| Pomacanthidae | Centropyge vrolikii | Н | No Group |
| Pomacentridae | Plectroglyphidodon leucozonus | Н | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|-------------------------------------|------------------|------------------|
| Pomacentridae | Plectroglyphidodon johnstonianus | Cor | No Group |
| Labridae | Anampses melanurus | MI | No Group |
| Apogonidae | Cheilodipterus quinquelineatus | MI | No Group |
| Cirrhitidae | Cirrhitichthys oxycephalus | MI | No Group |
| Cirrhitidae | Cirrhitops fasciatus | MI | No Group |
| Labridae | Halichoeres biocellatus | MI | No Group |
| Labridae | Labroides dimidiatus | MI | No Group |
| Labridae | Labropsis micronesica | MI | No Group |
| Labridae | Macropharyngodon negrosensis | MI | No Group |
| Labridae | Pseudojuloides cerasinus | MI | No Group |
| Labridae | Pseudojuloides polynesica | MI | No Group |
| Blenniidae | Aspidontus taeniatus | Par | No Group |
| Tetraodontidae | Torquigener randalli | MI | No Group |
| Pomacentridae | Plectroglyphidodon sindonis | Н | No Group |
| Pomacanthidae | Centropyge potteri | Н | No Group |
| Cirrhitidae | Oxycirrhites typus | Pk | No Group |
| Serranidae | Pseudanthias bicolor | Pk | No Group |
| Ptereleotridae | Ptereleotris microlepis | Pk | No Group |
| Pomacentridae | Stegastes lividus | Н | No Group |
| Labridae | Cirrhilabrus punctatus | MI | No Group |
| Labridae | Halichoeres margaritaceus | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|--------------------------------|------------------|------------------|
| Labridae | Pseudojuloides atavai | MI | No Group |
| Holocentridae | Sargocentron punctatissimum | MI | No Group |
| Monacanthidae | Pervagor janthinosoma | Om | No Group |
| Pomacentridae | Amphiprion clarkii | Pk | No Group |
| Serranidae | Anthias sp. | Pk | No Group |
| Blenniidae | Blenniella chrysospilos | Pk | No Group |
| Chaetodontidae | Chaetodon kleinii | Pk | No Group |
| Pomacentridae | Dascyllus trimaculatus | Pk | No Group |
| Apogonidae | Ostorhinchus maculiferus | Pk | No Group |
| Serranidae | Pseudanthias cooperi | Pk | No Group |
| Gobiidae | Amblygobius phalaena | Н | No Group |
| Tetraodontidae | Canthigaster amboinensis | Н | No Group |
| Tetraodontidae | Canthigaster coronata | Н | No Group |
| Pomacanthidae | Centropyge flavissima | Н | No Group |
| Pomacentridae | Stegastes nigricans | Н | No Group |
| Labridae | Halichoeres melanurus | MI | No Group |
| Labridae | Halichoeres melasmapomus | MI | No Group |
| Labridae | Labroides bicolor | MI | No Group |
| Labridae | Labropsis xanthonota | MI | No Group |
| Cirrhitidae | Paracirrhites arcatus | MI | No Group |
| Labridae | Pseudocheilinus evanidus | MI | No Group |
| Labridae | Pseudocheilinus octotaenia | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|---------------|-------------------------------|------------------|------------------|
| Monacanthidae | Pervagor aspricaudus | Om | No Group |
| Ostraciidae | Lactoria fornasini | SI | No Group |
| Labridae | Pseudojuloides sp. | MI | No Group |
| Pomacentridae | Abudefduf sexfasciatus | Pk | No Group |
| Pomacentridae | Chromis vanderbilti | Pk | No Group |
| Pomacentridae | Chromis xanthura | Pk | No Group |
| Labridae | Cirrhilabrus sp. | Pk | No Group |
| Pomacanthidae | Genicanthus watanabei | Pk | No Group |
| Labridae | Thalassoma amblycephalum | Pk | No Group |
| Pomacanthidae | Centropyge bicolor | Н | No Group |
| Serranidae | Belonoperca chabanaudi | MI | No Group |
| Labridae | Coris centralis | MI | No Group |
| Labridae | Halichoeres ornatissimus | MI | No Group |
| Malacanthidae | Hoplolatilus starcki | MI | No Group |
| Labridae | Macropharyngodon meleagris | MI | No Group |
| Labridae | Oxycheilinus bimaculatus | MI | No Group |
| Labridae | Pteragogus enneacanthus | MI | No Group |
| Labridae | Stethojulis balteata | MI | No Group |
| Labridae | Stethojulis strigiventer | MI | No Group |
| Labridae | Stethojulis trilineata | MI | No Group |
| Pomacentridae | Stegastes sp. | Н | No Group |
| Apogonidae | Apogon sp. | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|---------------------------|------------------|------------------|
| Apogonidae | Apogonidae | Pk | No Group |
| Chaetodontidae | Chaetodon miliaris | Pk | No Group |
| Pomacentridae | Dascyllus auripinnis | Pk | No Group |
| Labridae | Pseudocoris yamashiroi | Pk | No Group |
| Labridae | Stethojulis bandanensis | Pk | No Group |
| Monacanthidae | Cantherhines verecundus | Н | No Group |
| Pomacanthidae | Centropyge interrupta | Н | No Group |
| Pomacentridae | Stegastes fasciolatus | Н | No Group |
| Blenniidae | Exallias brevis | Cor | No Group |
| Labridae | Labrichthys unilineatus | Cor | No Group |
| Labridae | Halichoeres prosopeion | MI | No Group |
| Labridae | Macropharyngodon geoffroy | MI | No Group |
| Gobiidae | Valenciennea strigata | MI | No Group |
| Ostraciidae | Ostracion whitleyi | SI | No Group |
| Scorpaenidae | Dendrochirus barberi | MI | No Group |
| Blenniidae | Blenniidae | Pk | No Group |
| Synodontidae | Synodus binotatus | Pisc | No Group |
| Pomacentridae | Amphiprion chrysopterus | Pk | No Group |
| Serranidae | Pseudanthias pascalus | Pk | No Group |
| Acanthuridae | Ctenochaetus flavicauda | Н | No Group |
| Labridae | Cheilinus oxycephalus | MI | No Group |
| Holocentridae | Sargocentron diadema | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|----------------------------|------------------|------------------|
| Holocentridae | Sargocentron xantherythrum | MI | No Group |
| Labridae | Thalassoma quinquevittatum | MI | No Group |
| Labridae | Iniistius umbrilatus | MI | No Group |
| Labridae | Thalassoma sp. | MI | No Group |
| Pomacentridae | Pomacentridae | Om | No Group |
| Pomacentridae | Abudefduf notatus | Pk | No Group |
| Chaetodontidae | Hemitaurichthys polylepis | Pk | No Group |
| Ptereleotridae | Ptereleotris evides | Pk | No Group |
| Labridae | Anampses twistii | MI | No Group |
| Apogonidae | Cheilodipterus sp. | MI | No Group |
| Labridae | Cymolutes lecluse | MI | No Group |
| Labridae | Halichoeres hartzfeldii | MI | No Group |
| Labridae | Halichoeres marginatus | MI | No Group |
| Pinguipedidae | Parapercis clathrata | MI | No Group |
| Pinguipedidae | Parapercis schauinslandii | MI | No Group |
| Labridae | Choerodon jordani | Om | No Group |
| Monacanthidae | Pervagor sp. | Om | No Group |
| Monacanthidae | Pervagor spilosoma | Om | No Group |
| Pomacanthidae | Apolemichthys arcuatus | SI | No Group |
| Holocentridae | Neoniphon argenteus | MI | No Group |
| Apogonidae | Cheilodipterus artus | MI | No Group |
| Pomacentridae | Chromis ovalis | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|-----------------------------|------------------|-------------------------|
| Labridae | Bodianus mesothorax | MI | No Group |
| Pinguipedidae | Parapercis millepunctata | MI | No Group |
| Labridae | Halichoeres sp. | MI | No Group |
| Serranidae | Cephalopholis leopardus | Pisc | No Group |
| Apogonidae | Cheilodipterus macrodon | Pisc | No Group |
| Pomacentridae | Abudefduf vaigiensis | Pk | No Group |
| Chaetodontidae | Heniochus diphreutes | Pk | No Group |
| Holocentridae | Myripristis vittata | Pk | No Group |
| Caesionidae | Pterocaesio trilineata | Pk | No Group |
| Labridae | Thalassoma hardwicke | Pk | No Group |
| Monacanthidae | Cantherhines sandwichiensis | Н | No Group |
| Tetraodontidae | Canthigaster rivulata | Н | No Group |
| Acanthuridae | Zebrasoma flavescens | Н | No Group |
| Acanthuridae | Zebrasoma scopas | Н | No Group |
| Monacanthidae | Amanses scopas | Cor | No Group |
| Labridae | Anampses chrysocephalus | MI | No Group |
| Labridae | Anampses sp. | MI | No Group |
| Labridae | Bodianus axillaris | MI | No Group |
| Labridae | Bodianus prognathus | MI | No Group |
| Labridae | Coris dorsomacula | MI | No Group |
| Labridae | Coris venusta | MI | No Group |
| Labridae | Cymolutes praetextatus | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|---------------------------------|------------------|-------------------------|
| Labridae | Pseudocoris aurantiofasciata | MI | No Group |
| Labridae | Pseudocoris heteroptera | MI | No Group |
| Scorpaenidae | Pterois antennata | MI | No Group |
| Holocentridae | Sargocentron microstoma | MI | No Group |
| Labridae | Thalassoma jansenii | MI | No Group |
| Nemipteridae | Scolopsis lineata | Om | No Group |
| Zanclidae | Zanclus cornutus | SI | No Group |
| Labridae | Bodianus anthioides | Pk | No Group |
| Chaetodontidae | Hemitaurichthys thompsoni | Pk | No Group |
| Acanthuridae | Zebrasoma rostratum | Н | No Group |
| Kuhliidae | Kuhlia sandvicensis | Pk | No Group |
| Scorpaenidae | Pterois sphex | Pisc | No Group |
| Synodontidae | Synodontidae | Pisc | No Group |
| Pomacentridae | Chromis verater | Pk | No Group |
| Pempheridae | Pempheridae | Pk | No Group |
| Serranidae | Pseudanthias thompsoni | Pk | No Group |
| Balistidae | Xanthichthys auromarginatus | Pk | No Group |
| Acanthuridae | Ctenochaetus binotatus | Н | No Group |
| Labridae | Anampses meleagrides | MI | No Group |
| Labridae | Iniistius aneitensis | MI | No Group |
| Mullidae | Parupeneus chrysonemus | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|---------------------------|------------------|-------------------------|
| Balistidae | Sufflamen chrysopterum | MI | No Group |
| Cirrhitidae | Paracirrhites forsteri | Pisc | No Group |
| Synodontidae | Saurida gracilis | Pisc | No Group |
| Holocentridae | Myripristis kuntee | Pk | No Group |
| Pempheridae | Pempheris oualensis | Pk | No Group |
| Pomacentridae | Abudefduf septemfasciatus | Н | No Group |
| Acanthuridae | Acanthurus nigricans | Н | No Group |
| Acanthuridae | Acanthurus nigrofuscus | Н | No Group |
| Holocentridae | Neoniphon aurolineatus | MI | No Group |
| Pinguipedidae | Parapercis sp. | MI | No Group |
| Labridae | Bodianus sanguineus | Om | No Group |
| Synodontidae | Synodus dermatogenys | Pisc | No Group |
| Synodontidae | Synodus variegatus | Pisc | No Group |
| Pomacentridae | Abudefduf sordidus | Н | No Group |
| Holocentridae | Myripristis earlei | MI | No Group |
| Pomacentridae | Abudefduf abdominalis | Pk | No Group |
| Pomacanthidae | Genicanthus personatus | Pk | No Group |
| Chaetodontidae | Heniochus acuminatus | Pk | No Group |
| Holocentridae | Myripristis chryseres | Pk | No Group |
| Holocentridae | Myripristis woodsi | Pk | No Group |
| Labridae | Thalassoma lunare | Pk | No Group |
| Acanthuridae | Acanthurus achilles | Н | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|---------------|---------------------------------------|------------------|------------------|
| Acanthuridae | Acanthurus achilles & nigricans | Н | No Group |
| Acanthuridae | Acanthurus leucopareius | Н | No Group |
| Acanthuridae | Acanthurus pyroferus | Н | No Group |
| Monacanthidae | Cantherhines pardalis | Н | No Group |
| Labridae | Bodianus diana | MI | No Group |
| Balistidae | Rhinecanthus rectangulus | MI | No Group |
| Holocentridae | Sargocentron caudimaculatum | MI | No Group |
| Holocentridae | Sargocentron ensifer | MI | No Group |
| Labridae | Thalassoma duperrey & quinquevittatum | MI | No Group |
| Labridae | Thalassoma lutescens | MI | No Group |
| Pomacanthidae | Apolemichthys griffisi | SI | No Group |
| Pomacanthidae | Apolemichthys trimaculatus | SI | No Group |
| Pomacanthidae | Apolemichthys xanthopunctatus | SI | No Group |
| Pomacanthidae | Pygoplites diacanthus | SI | No Group |
| Serranidae | Epinephelus hexagonatus | Pisc | No Group |
| Acanthuridae | Acanthurus nubilus | Pk | No Group |
| Muraenidae | Gymnothorax melatremus | MI | No Group |
| Labridae | Pseudodax moluccanus | MI | No Group |
| Labridae | Thalassoma duperrey | MI | No Group |
| Acanthuridae | Acanthurus triostegus | Н | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|---------------|---------------------------|------------------|-------------------------|
| Serranidae | Grammistes sexlineatus | MI | No Group |
| Labridae | Halichoeres hortulanus | MI | No Group |
| Labridae | Halichoeres trimaculatus | MI | No Group |
| Serranidae | Cephalopholis urodeta | Pisc | No Group |
| Cirrhitidae | Paracirrhites hemistictus | Pisc | No Group |
| Acanthuridae | Acanthurus thompsoni | Pk | No Group |
| Siganidae | Siganus spinus | Н | No Group |
| Balistidae | Rhinecanthus lunula | MI | No Group |
| Balistidae | Sufflamen bursa | MI | No Group |
| Ostraciidae | Ostracion meleagris | SI | No Group |
| Acanthuridae | Acanthurus guttatus | Н | No Group |
| Cirrhitidae | Cirrhitidae | MI | No Group |
| Serranidae | Cephalopholis spiloparaea | Pisc | No Group |
| Labridae | Oxycheilinus digramma | Pisc | No Group |
| Scorpaenidae | Scorpaenopsis diabolus | Pisc | No Group |
| Scorpaenidae | Scorpaenopsis sp. | Pisc | No Group |
| Synodontidae | Synodus ulae | Pisc | No Group |
| Caesionidae | Caesio lunaris | Pk | No Group |
| Balistidae | Canthidermis maculata | Pk | No Group |
| Hemiramphidae | Hyporhamphus acutus | Pk | No Group |
| Caesionidae | Pterocaesio lativittata | Pk | No Group |
| Caesionidae | Pterocaesio tile | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|---------------|-----------------------------|------------------|------------------|
| Carangidae | Selar crumenophthalmus | Pk | No Group |
| Balistidae | Xanthichthys mento | Pk | No Group |
| Acanthuridae | Ctenochaetus sp. | Н | No Group |
| Acanthuridae | Naso thynnoides | Н | No Group |
| Balistidae | Balistapus undulatus | MI | No Group |
| Cirrhitidae | Cirrhitus pinnulatus | MI | No Group |
| Labridae | Coris ballieui | MI | No Group |
| Lethrinidae | Gnathodentex aureolineatus | MI | No Group |
| Malacanthidae | Malacanthus brevirostris | MI | No Group |
| Mullidae | Mulloidichthys mimicus | MI | No Group |
| Holocentridae | Myripristis violacea | MI | No Group |
| Labridae | Novaculichthys taeniourus | MI | No Group |
| Balistidae | Rhinecanthus aculeatus | MI | No Group |
| Synodontidae | Saurida flamma | Pisc | No Group |
| Acanthuridae | Paracanthurus hepatus | Pk | No Group |
| Caesionidae | Caesionidae | Pk | No Group |
| Holocentridae | Holocentridae | MI | No Group |
| Priacanthidae | Heteropriacanthus carolinus | Pk | No Group |
| Holocentridae | Myripristis adusta | Pk | No Group |
| Holocentridae | Myripristis amaena | Pk | No Group |
| Labridae | Cheilinus chlorourus | MI | No Group |
| Labridae | Gomphosus varius | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|---------------------------|------------------|-------------------------|
| Lethrinidae | Lethrinus harak | MI | No Group |
| Holocentridae | Neoniphon sammara | MI | No Group |
| Serranidae | Epinephelus melanostigma | Pisc | No Group |
| Serranidae | Epinephelus merra | Pisc | No Group |
| Holocentridae | Myripristis berndti | Pk | No Group |
| Priacanthidae | Priacanthus hamrur | Pk | No Group |
| Priacanthidae | Priacanthus meeki | Pk | No Group |
| Acanthuridae | Acanthurus albipectoralis | Н | No Group |
| Tetraodontidae | Arothron nigropunctatus | Cor | No Group |
| Mullidae | Parupeneus insularis | MI | No Group |
| Mullidae | Parupeneus pleurostigma | MI | No Group |
| Holocentridae | Sargocentron tiere | MI | No Group |
| Labridae | Thalassoma trilobatum | MI | No Group |
| Mullidae | Upeneus taeniopterus | MI | No Group |
| Balistidae | Melichthys vidua | Н | No Group |
| Serranidae | Epinephelus spilotoceps | Pisc | No Group |
| Lutjanidae | Lutjanus semicinctus | Pisc | No Group |
| Serranidae | Pogonoperca punctata | Pisc | No Group |
| Caesionidae | Caesio caerulaurea | Pk | No Group |
| Carangidae | Decapterus macarellus | Pk | No Group |
| Holocentridae | Myripristinae | Pk | No Group |
| Caesionidae | Pterocaesio marri | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|---------------|----------------------------------|------------------|-------------------------|
| Balistidae | Xanthichthys caeruleolineatus | Pk | No Group |
| Labridae | Iniistius pavo | MI | No Group |
| Holocentridae | Neoniphon opercularis | MI | No Group |
| Holocentridae | Neoniphon sp. | MI | No Group |
| Mullidae | Parupeneus crassilabris | MI | No Group |
| Labridae | Anampses cuvier | MI | No Group |
| Labridae | Cheilinus fasciatus | MI | No Group |
| Siganidae | Siganus punctatus | Н | No Group |
| Gobiidae | Gobiidae | MI | No Group |
| Scorpaenidae | Pterois volitans | Pisc | No Group |
| Balistidae | Melichthys niger | Pk | No Group |
| Priacanthidae | Priacanthus sp. | Pk | No Group |
| Monacanthidae | Monacanthidae | Н | No Group |
| Siganidae | Siganidae | Н | No Group |
| Diodontidae | Diodon holocanthus | MI | No Group |
| Mullidae | Mulloidichthys vanicolensis | MI | No Group |
| Mullidae | Parupeneus multifasciatus | MI | No Group |
| Balistidae | Sufflamen fraenatum | MI | No Group |
| Monacanthidae | Cantherhines dumerilii | Om | No Group |
| Pomacanthidae | Pomacanthus imperator | SI | No Group |
| Lethrinidae | Lethrinus rubrioperculatus | MI | No Group |
| Caesionidae | Caesio teres | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|------------------|------------------------------|------------------|-------------------------|
| Balistidae | Odonus niger | Pk | No Group |
| Acanthuridae | Acanthurus nigricauda | Н | No Group |
| Acanthuridae | Acanthurus olivaceus | Н | No Group |
| Acanthuridae | Zebrasoma veliferum | Н | No Group |
| Labridae | Bodianus loxozonus | MI | No Group |
| Labridae | Coris gaimard | MI | No Group |
| Labridae | Hologymnosus annulatus | MI | No Group |
| Labridae | Hologymnosus doliatus | MI | No Group |
| Mullidae | Mulloidichthys flavolineatus | MI | No Group |
| Acanthuridae | Acanthurus maculiceps | Н | No Group |
| Kyphosidae | Kyphosus hawaiiensis | Н | No Group |
| Cheilodactylidae | Cheilodactylus vittatus | SI | No Group |
| Ostraciidae | Ostraciidae | SI | No Group |
| Siganidae | Siganus argenteus | Н | No Group |
| Labridae | Anampses caeruleopunctatus | MI | No Group |
| Serranidae | Epinephelus fasciatus | Pisc | No Group |
| Labridae | Thalassoma ballieui | MI | No Group |
| Labridae | Thalassoma purpureum | MI | No Group |
| Serranidae | Cephalopholis miniata | Pisc | No Group |
| Hemiramphidae | Hemiramphidae | Pk | No Group |
| Acanthuridae | Acanthurus leucocheilus | Н | No Group |
| Ostraciidae | Ostracion cubicus | Н | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|---------------------------|------------------|-------------------------|
| Bothidae | Bothus mancus | MI | No Group |
| Labridae | Cheilinus sp. | MI | No Group |
| Labridae | Cheilinus trilobatus | MI | No Group |
| Malacanthidae | Malacanthus latovittatus | MI | No Group |
| Labridae | Oxycheilinus unifasciatus | Pisc | No Group |
| Labridae | Oxycheilinus sp. | MI | No Group |
| Serranidae | Epinephelus retouti | Pisc | No Group |
| Mullidae | Mulloidichthys pfluegeri | MI | No Group |
| Serranidae | Cephalopholis sexmaculata | Pisc | No Group |
| Serranidae | Cephalopholis sonnerati | Pisc | No Group |
| Serranidae | Gracila albomarginata | Pisc | No Group |
| Mullidae | Parupeneus cyclostomus | Pisc | No Group |
| Belonidae | Platybelone argalus | Pisc | No Group |
| Acanthuridae | Acanthurus mata | Pk | No Group |
| Tetraodontidae | Arothron meleagris | Cor | No Group |
| Balistidae | Balistoides conspicillum | MI | No Group |
| Labridae | Hemigymnus fasciatus | MI | No Group |
| Lethrinidae | Lethrinus obsoletus | MI | No Group |
| Mullidae | Mullidae | MI | No Group |
| Mullidae | Parupeneus barberinus | MI | No Group |
| Holocentridae | Sargocentron sp. | MI | No Group |
| Ephippidae | Platax orbicularis | Om | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|------------------------------|------------------|-------------------------|
| Serranidae | Epinephelus macrospilos | Pisc | No Group |
| Scorpaenidae | Scorpaenopsis cacopsis | Pisc | No Group |
| Kyphosidae | Kyphosus cinerascens | Н | No Group |
| Labridae | Cheilio inermis | MI | No Group |
| Mullidae | Parupeneus porphyreus | MI | No Group |
| Serranidae | Epinephelus socialis | Pisc | No Group |
| Tetraodontidae | Arothron hispidus | MI | No Group |
| Holocentridae | Sargocentron spiniferum | MI | No Group |
| Carangidae | Trachinotus baillonii | Pisc | No Group |
| Labridae | Epibulus insidiator | MI | No Group |
| Serranidae | Epinephelus howlandi | Pisc | No Group |
| Labridae | Bodianus albotaeniatus | MI | No Group |
| Labridae | Bodianus bilunulatus | MI | No Group |
| Acanthuridae | Acanthurus sp. | Н | No Group |
| Serranidae | Aethaloperca rogaa | Pisc | No Group |
| Serranidae | Anyperodon leucogrammicus | Pisc | No Group |
| Serranidae | Cephalopholis argus | Pisc | No Group |
| Serranidae | Cephalopholis sp. | Pisc | No Group |
| Serranidae | Epinephelus maculatus | Pisc | No Group |
| Holocentridae | Myripristis murdjan | Pk | No Group |
| Acanthuridae | Naso brevirostris | Pk | No Group |
| Acanthuridae | Naso maculatus | Pk | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|-----------------------------------|------------------|------------------|
| Acanthuridae | Naso vlamingii | Pk | No Group |
| Kyphosidae | Kyphosus vaigiensis | Н | No Group |
| Muraenidae | Gymnothorax eurostus | MI | No Group |
| Labridae | Hemigymnus melapterus | MI | No Group |
| Balistidae | Pseudobalistes flavimarginatus | MI | No Group |
| Lethrinidae | Lethrinus xanthochilus | Pisc | No Group |
| Acanthuridae | Naso caesius | Pk | No Group |
| Lethrinidae | Monotaxis grandoculis | MI | No Group |
| Serranidae | Variola albimarginata | Pisc | No Group |
| Labridae | Coris flavovittata | MI | No Group |
| Tetraodontidae | Arothron mappa | Om | No Group |
| Carangidae | Carangoides ferdau | Pisc | No Group |
| Carangidae | Carangoides orthogrammus | Pisc | No Group |
| Carangidae | Scomberoides lysan | Pisc | No Group |
| Acanthuridae | Acanthuridae | Н | No Group |
| Lethrinidae | Lethrinus amboinensis | MI | No Group |
| Lethrinidae | Lethrinus erythracanthus | MI | No Group |
| Ephippidae | Platax teira | Om | No Group |
| Serranidae | Plectropomus areolatus | Pisc | No Group |
| Carangidae | Gnathanodon speciosus | Pisc | No Group |
| Serranidae | Epinephelus polyphekadion | Pisc | No Group |
| Serranidae | Epinephelus tauvina | Pisc | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|-------------------------|------------------|------------------|
| Muraenidae | Gymnothorax breedeni | Pisc | No Group |
| Acanthuridae | Naso hexacanthus | Pk | No Group |
| Acanthuridae | Naso sp. | Pk | No Group |
| Kyphosidae | Kyphosus sandwicensis | Н | No Group |
| Kyphosidae | Kyphosus sp. | Н | No Group |
| Balistidae | Balistidae | MI | No Group |
| Balistidae | Balistoides viridescens | MI | No Group |
| Muraenidae | Echidna nebulosa | MI | No Group |
| Haemulidae | Plectorhinchus gibbosus | MI | No Group |
| Balistidae | Balistes polylepis | MI | No Group |
| Tetraodontidae | Tetraodontidae | MI | No Group |
| Monacanthidae | Aluterus scriptus | Om | No Group |
| Ophichthidae | Myrichthys magnificus | MI | No Group |
| Aulostomidae | Aulostomus chinensis | Pisc | No Group |
| Muraenidae | Enchelycore pardalis | Pisc | No Group |
| Sphyraenidae | Sphyraena helleri | Pisc | No Group |
| Muraenidae | Gymnothorax rueppelliae | MI | No Group |
| Oplegnathidae | Oplegnathus fasciatus | MI | No Group |
| Serranidae | Variola louti | Pisc | No Group |
| Haemulidae | Plectorhinchus picus | MI | No Group |
| Haemulidae | Plectorhinchus vittatus | MI | No Group |
| Lethrinidae | Lethrinidae | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|---------------|---------------------------|------------------|------------------|
| Lethrinidae | Lethrinus sp. | MI | No Group |
| Oplegnathidae | Oplegnathus punctatus | MI | No Group |
| Carangidae | Caranx papuensis | Pisc | No Group |
| Muraenidae | Gymnothorax steindachneri | Pisc | No Group |
| Diodontidae | Diodon hystrix | MI | No Group |
| Labridae | Labridae | MI | No Group |
| Belonidae | Belonidae | Pisc | No Group |
| Carangidae | Caranx lugubris | Pisc | No Group |
| Carangidae | Caranx sexfasciatus | Pisc | No Group |
| Scombridae | Euthynnus affinis | Pisc | No Group |
| Scombridae | Grammatorcynus bilineatus | Pisc | No Group |
| Lethrinidae | Lethrinus olivaceus | Pisc | No Group |
| Acanthuridae | Naso annulatus | Pk | No Group |
| Ophidiidae | Brotula multibarbata | MI | No Group |
| Dasyatidae | Urogymnus granulatus | MI | No Group |
| Scombridae | Sarda orientalis | Pisc | No Group |
| Congridae | Congridae | Pisc | No Group |
| Congridae | Heterocongrinae | Pisc | No Group |
| Scombridae | Katsuwonus pelamis | Pisc | No Group |
| Echeneidae | Echeneis naucrates | Pk | No Group |
| Carangidae | Trachinotus blochii | MI | No Group |
| Carangidae | Caranx melampygus | Pisc | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|----------------|------------------------|------------------|-------------------------|
| Muraenidae | Gymnothorax meleagris | Pisc | No Group |
| Tetraodontidae | Arothron stellatus | Cor | No Group |
| Labridae | Coris aygula | MI | No Group |
| Carangidae | Pseudocaranx dentex | Pisc | No Group |
| Muraenidae | Scuticaria tigrina | Pisc | No Group |
| Serranidae | Plectropomus laevis | Pisc | No Group |
| Serranidae | Epinephelus sp. | Pisc | No Group |
| Serranidae | Serranidae | Pisc | No Group |
| Belonidae | Tylosurus crocodilus | Pisc | No Group |
| Carangidae | Alectis ciliaris | Pisc | No Group |
| Muraenidae | Enchelynassa canina | Pisc | No Group |
| Muraenidae | Gymnothorax undulatus | Pisc | No Group |
| Muraenidae | Gymnomuraena zebra | MI | No Group |
| Carangidae | Carangidae | Pisc | No Group |
| Fistulariidae | Fistularia commersonii | Pisc | No Group |
| Carangidae | Caranx ignobilis | Pisc | No Group |
| Carangidae | Caranx sp. | Pisc | No Group |
| Sphyraenidae | Sphyraena qenie | Pisc | No Group |
| Carangidae | Elagatis bipinnulata | Pisc | No Group |
| Chanidae | Chanos | Н | No Group |
| Dasyatidae | Taeniurops meyeni | MI | No Group |
| Dasyatidae | Dasyatidae | MI | No Group |

| Family | Scientific Name | Trophic Group | Functional Group |
|--------------------|--------------------------------|------------------|-------------------------|
| Carangidae | Seriola dumerili | Pisc | No Group |
| Carcharhinidae | Carcharhinus melanopterus | Pisc | No Group |
| Sphyraenidae | Sphyraena barracuda | Pisc | No Group |
| Scombridae | Thunnus albacares | Pisc | No Group |
| Carcharhinidae | Triaenodon obesus | Pisc | No Group |
| Labridae | Cheilinus undulatus | MI | No Group |
| Carcharhinidae | Carcharhinus amblyrhynchos | Pisc | No Group |
| Muraenidae | Gymnothorax flavimarginatus | Pisc | No Group |
| Scombridae | Scombridae | Pisc | No Group |
| Scombridae | Gymnosarda unicolor | Pisc | No Group |
| Muraenidae | Muraenidae | Pisc | No Group |
| Carcharhinidae | Carcharhinus limbatus | Pisc | No Group |
| Muraenidae | Gymnothorax javanicus | Pisc | No Group |
| Muraenidae | Gymnothorax sp. | Pisc | No Group |
| Ginglymostomatidae | Nebrius ferrugineus | Pisc | No Group |
| Myliobatidae | Aetobatus ocellatus | MI | No Group |
| Carcharhinidae | Carcharhinus galapagensis | Pisc | No Group |
| Sphyrnidae | Sphyrna lewini | Pisc | No Group |
| Sphyrnidae | Sphyrnidae | Pisc | No Group |
| Myliobatidae | Mobula sp. | Pk | No Group |
| Scaridae | Scarus fuscocaudalis | Н | Parrotfish |

| Family | mily Scientific Name | | Functional Group |
|----------|--------------------------|---|------------------|
| Scaridae | Calotomus zonarchus | Н | Parrotfish |
| Scaridae | Chlorurus japanensis | Н | Parrotfish |
| Scaridae | Scarus globiceps | Н | Parrotfish |
| Scaridae | Scarus spinus | Н | Parrotfish |
| Scaridae | Scarus psittacus | Н | Parrotfish |
| Scaridae | Scarus dubius | Н | Parrotfish |
| Scaridae | Scarus oviceps | Н | Parrotfish |
| Scaridae | Scarus schlegeli | Н | Parrotfish |
| Scaridae | Chlorurus spilurus | Н | Parrotfish |
| Scaridae | Scarus niger | Н | Parrotfish |
| Scaridae | Scarus festivus | Н | Parrotfish |
| Scaridae | Scarus frenatus | Н | Parrotfish |
| Scaridae | Chlorurus frontalis | Н | Parrotfish |
| Scaridae | Scarus dimidiatus | Н | Parrotfish |
| Scaridae | Calotomus carolinus | Н | Parrotfish |
| Scaridae | Scarus forsteni | Н | Parrotfish |
| Scaridae | Scarus tricolor | Н | Parrotfish |
| Scaridae | Scarus xanthopleura | Н | Parrotfish |
| Scaridae | Hipposcarus longiceps | Н | Parrotfish |
| Scaridae | Scarus altipinnis | Н | Parrotfish |
| Scaridae | Chlorurus perspicillatus | Н | Parrotfish |
| Scaridae | Scaridae | Н | Parrotfish |

| Family | Scientific Name | Trophic Group | Functional Group |
|------------|---------------------------|------------------|------------------|
| Scaridae | Scarus rubroviolaceus | Н | Parrotfish |
| Scaridae | Chlorurus microrhinos | Н | Parrotfish |
| Scaridae | Cetoscarus ocellatus | Н | Parrotfish |
| Scaridae | Scarus ghobban | Н | Parrotfish |
| Scaridae | Chlorurus sp. | Н | Parrotfish |
| Scaridae | Scarus sp. | Н | Parrotfish |
| Scaridae | Bolbometopon muricatum | Cor | Parrotfish |
| Lutjanidae | Lutjanus fulvus | MI | Snappers |
| Lutjanidae | Lutjanus kasmira | MI | Snappers |
| Lutjanidae | Lutjanus gibbus | MI | Snappers |
| Lutjanidae | Lutjanus monostigma | Pisc | Snappers |
| Lutjanidae | Macolor macularis | Pk | Snappers |
| Lutjanidae | Aphareus furca | Pisc | Snappers |
| Lutjanidae | Macolor niger | Pk | Snappers |
| Lutjanidae | Macolor sp. | Pk | Snappers |
| Lutjanidae | Lutjanus bohar | Pisc | Snappers |
| Lutjanidae | Lutjanus argentimaculatus | MI | Snappers |
| Lutjanidae | Aprion virescens | Pisc | Snappers |

APPENDIX B: LIST OF PROTECTED SPECIES AND DESIGNATED CRITICAL HABITAT

Table B-1. Protected species found or reasonably believed to be found near or in Mariana Archipelago waters

| Common name | Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References | | | |
|-------------------------------------|--|--------------------|-------------|------------------|---------------|----------------------------|--|--|--|
| | Seabirds | | | | | | | | |
| Wedge-Tailed Shearwater | Ardenna pacifica | Not Listed | N/A | Uncommon visitor | Both | Wiles 2003 | | | |
| Streaked Shearwater | Calonectris leucomelas | Not Listed | N/A | Rare visitor | Guam | Wiles 2003 | | | |
| Short-Tailed Shearwater | Ardenna tenuirostris | Not Listed | N/A | Common visitor | Both | Wiles 2003 | | | |
| Newell's Shearwater ^a | Puffinus newelli (Puffinus auricularis newelli) | Endangered | N/A | Rare visitor | Both | 40 FR 44149, Wiles 2003 | | | |
| Audubon's Shearwater | Puffinus Iherminieri | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| Leach's Storm-Petrel | Oceanodroma leucorhoa | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| Matsudaira's Storm-Petrel | Oceanodroma matsudairae | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| White-Tailed Tropicbird | Phaethon lepturus | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| Red-Tailed Tropicbird | Phaethon rubricauda | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| Masked Booby | Sula dactylatra | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| Brown Booby | Sula leucogaster | Not Listed | N/A | Uncommon visitor | Both | Wiles 2003 | | | |
| Red-Footed Booby | Sula | Not Listed | N/A | Uncommon visitor | Both | Wiles 2003 | | | |
| Great Frigatebird | Fregata minor | Not Listed | N/A | Rare visitor | Both | Wiles 2003 | | | |
| Lesser Frigatebird | Fregata ariel | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 | | | |

| Common name | Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References |
|---------------------------|-------------------------------|--------------------|-------------|---|---------------|--|
| Black-Headed Gull | Chroicocephalus ridibundus | Not Listed | N/A | Rare visitor | Both | Wiles 2003 |
| Gull-Billed Tern | Gelochelidon nilotica | Not Listed | N/A | Rare visitor | Both | Wiles 2003 |
| Great Crested Tern | Thalasseus bergii | Not Listed | N/A | Uncommon visitor | Both | Wiles 2003 |
| Common Tern | Sterna hirundo | Not Listed | N/A | Rare visitor | Both | Wiles 2003 |
| Black-Naped Tern | Sterna sumatrana | Not Listed | N/A | Rare visitor | Guam | Wiles 2003 |
| Little Tern | Sternula albifrons | Not Listed | N/A | Rare visitor | Both | Wiles 2003 |
| Sooty Tern | Onychoprion fuscatus | Not Listed | N/A | Rare visitor | Both | Wiles 2003 |
| White-Winged Tern | Chlidonias leucopterus | Not Listed | N/A | Rare visitor | Both | Wiles 2003 |
| Brown Noddy | Anous stolidus | Not Listed | N/A | Common resident | Both | Wiles 2003 |
| Black Noddy | Anous minutus | Not Listed | N/A | Common visitor | Both | Wiles 2003 |
| White Tern | Gygis alba | Not Listed | N/A | Common resident | Both | Wiles 2003 |
| Short-Tailed Albatross | Phoebastria albatrus | Endangered | N/A | Breed in Japan and NWHI, and range across the North Pacific Ocean. Potential range includes the Marianas archipelago. | N/A | 35 FR 8495, 65 FR 46643, BirdLife International 2017 |
| Laysan Albatross | Phoebastria immutabilis | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Black-Footed Albatross | Phoebastria nigripes | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| White-Necked Petrel | Pterodroma cervicalis | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Bonin Petrel | Pterodroma hypoleuca | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Black-Winged Petrel | Pterodroma nigripennis | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |

| Common name | Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References |
|-----------------------------|-----------------------------|---|-------------|---|---------------|--|
| Bulwer's Petrel | Bulweria bulwerii | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Christmas Shearwater | Puffinus nativitatis | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Band-Rumped Storm-Petrel | Oceanodroma castro | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Long-Tailed Jaeger | Stercorarius Iongicaudus | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Laughing Gull | Leucophaeus atricilla | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Herring Gull | Larus argentatus | Not Listed | N/A | Rare visitor | CNMI | Wiles 2003 |
| Gray-Backed Tern | Onychoprion lunatus | Not Listed | N/A | Uncommon resident | CNMI | Wiles 2003 |
| | l | | Sea Turtles | | | L |
| Green Sea Turtle | Chelonia mydas | Endangered (Central West Pacific DPS) | N/A | An estimated 1000-2000 turtles forage in Guam/CNMI waters. Particularly common in winter and late spring. | Both | 43 FR 32800, 81 FR 20057, Kolinski et al. 2000, Pritchard 1982, Honigman 1994 |
| Hawksbill Sea Turtle | Eretmochelys imbricata | Endangeredb | N/A | Small population nesting and foraging around Guam. Occur worldwide in tropical and subtropical waters. | Both | 35 FR 8491, NMFS & USFWS 2007, Baillie & Groombridge 1996 |
| Leatherback Sea Turtle | Dermochelys coriacea | Endangered ^b | N/A | Occasional sightings. Occur worldwide in tropical, subtropical, and subpolar waters. | Guam | 35 FR 8491, Eldredge 2003, Eckert et al. 2012 |
| Loggerhead Sea Turtle | Caretta caretta | Endangered (North Pacific DPS) | N/A | No known sightings. Found worldwide along continental shelves, bays, estuaries, and lagoons of tropical, subtropical, and temperate waters. | N/A | 43 FR 32800, 76 FR 58868, Dodd 1990, USFWS 2005 |

| Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References |
|----------------------------|--|---------------------------|--|--|---|
| Lepidochelys olivacea | Threatened (Entire species, except for the breeding population on the Pacific coast of Mexico, which is listed as endangered) | N/A | Believed to occasionally transit through area. | N/A | 43 FR 32800, Starmer et al. 2005 |
| | | Marine mamma | ils | | |
| Mesoplodon densirostris | Not Listed | Non-strategic | Found worldwide in tropical and temperate waters. | CNMI | Mead 1989 |
| Balaenoptera musculus | Endangered | Strategic | No known sightings in CNMI but occur worldwide in tropical and warm-temperate waters. Known to occur in the western North Pacific. | N/A | 35 FR 18319, McDonald et al. 2006, Stafford et al. 2001 |
| Tursiops truncatus | Not Listed | Non-strategic | Distributed worldwide in tropical and warm-temperate waters | Both | Perrin et al. 2009 |
| Balaenoptera edeni | Not Listed | Non-strategic | Distributed widely across tropical and warm-temperate Pacific Ocean. | CNMI | Leatherwood et al. 1982 |
| Ziphius cavirostris | Not Listed | Non-strategic | Occur worldwide. | CNMI | Heyning 1989 |
| Dugong dugong | Endangered | N/A (managed by USFWS) | Extremely rare. One confirmed sighting in Guam in 1975, and multiple anecdotal reports in Guam in 1985. | Guam | Randall et al. 1975, Eldredge 2003 |
| Kogia sima | Not Listed | Non-strategic | Found worldwide in tropical and warm-temperate waters. | Both | Nagorsen 1985 |
| Pseudorca crassidens | Not Listed | Non-strategic | Found worldwide in tropical and warm-temperate waters. | CNMI | Stacey et al. 1994 |
| | Lepidochelys olivacea Mesoplodon densirostris Balaenoptera musculus Tursiops truncatus Balaenoptera edeni Ziphius cavirostris Dugong dugong Kogia sima Pseudorca | Centific name Status | Company Comp | Threatened (Entire species, except for the breeding population on the Pacific coast of Mexico, which is listed as endangered) Mesoplodon densirostris | Threatened (Entire species, except for the breeding population on the Pacific coast of Mexico, which is listed as endangered) Marine mammals |

| Common name | Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References |
|-----------------------------------|-------------------------------|---|---------------|---|---------------|--|
| Fin Whale | Balaenoptera physalus | Endangered | Strategic | Infrequent sightings, occur throughout the North Pacific Ocean. | N/A | 35 FR 18319, Oleson et al. 2015, Mizroch et al. 2009 |
| Fraser's Dolphin | Lagenodelphis hosei | Not Listed | Non-strategic | Found worldwide in tropical waters. | CNMI | Perrin et al. 2009 |
| Humpback Whale | Megaptera novaeangliae | Endangered (Western North Pacific DPS) | Strategic | Occasional sightings in Guam/CNMI waters during winter breeding season. | Both | 35 FR 18319, 81 FR 62259, Guarrige et al. 2007, SPWRC 2008 |
| Killer Whale | Orcinus orca | Not Listed | Non-strategic | Found worldwide. Prefer colder waters within 800 km of continents. | Guam | Leatherwood & Dalheim 1978, Mitchell 1975, Baird et al. 2006 |
| Longman's Beaked Whale | Indopacetus pacificus | Not Listed | Non-strategic | Found in tropical waters from the eastern Pacific westward through the Indian Ocean to the eastern coast of Africa. | CNMI | Dalebout 2003 |
| Melon-Headed Whale | Peponocephala electra | Not Listed | Non-strategic | Found in tropical and warm-temperate waters worldwide, primarily found in equatorial waters. | Both | Perryman et al. 1994 |
| Minke Whale | Balaenoptera acutorostrata | Not Listed | Non-strategic | Uncommon in this region, usually seen over continental shelves in the Pacific Ocean. | CNMI | Brueggeman et al. 1990 |
| Northern Elephant Seal | Mirounga angustirostris | Not Listed | Non-strategic | Females migrate to central North Pacific to feed on pelagic prey | N/A | Le Beouf et al. 2000 |
| Pantropical Spotted Dolphin | Stenella attenuata | Not Listed | Non-strategic | Found in tropical and subtropical waters worldwide. | Both | Perrin et al. 2009 |
| Pygmy Killer Whale | Feresa attenuata | Not Listed | Non-strategic | Found in tropical and subtropical waters worldwide. | CNMI | Ross & Leatherwood 1994 |
| Pygmy Sperm Whale | Kogia breviceps | Not Listed | Non-strategic | Found worldwide in tropical and warm-temperate waters. | Guam | Caldwell & Caldwell 1989 |

| Common name | Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References | | | |
|------------------------------|-------------------------------|--------------------|---------------|---|---------------|---|--|--|--|
| Risso's Dolphin | Grampus griseus | Not Listed | Non-strategic | Found in tropical to warm- temperate waters worldwide. | Both | Perrin et al. 2009 | | | |
| Rough- Toothed Dolphin | Steno bredanensis | Not Listed | Non-strategic | Found in tropical to warm-temperate waters worldwide. | CNMI | Perrin et al. 2009 | | | |
| Sei Whale | Balaenoptera borealis | Endangered | Strategic | Extremely rare. Generally found in offshore temperate waters. | CNMI | 35 FR 18319, Barlow 2003, Bradford et al. 2013 | | | |
| Short-Finned Pilot Whale | Globicephala macrorhynchus | Not Listed | Non-strategic | Found in tropical to warm- temperate waters worldwide. | Both | Shallenberger 1981, Baird et al. 2013, Bradford et al. 2013 | | | |
| Sperm Whale | Physeter macrocephalus | Endangered | Strategic | Found in tropical to polar waters worldwide, most abundant cetaceans in the region. Regularly sighted in waters around CNMI. | Both | 35 FR 18319, Rice 1960, Barlow 2006, Lee 1993, Mobley et al. 2000, Shallenberger 1981 | | | |
| Spinner Dolphin | Stenella longirostris | Not Listed | Non-strategic | Found worldwide in tropical and warm-temperate waters. Occur in shallow protected bays during the day, feed offshore at night. | Both | Norris and Dohl 1980, Norris et al. 1994, Hill et al. 2010, Andews et al. 2010, Karczmarski 2005, Perrin et al. 2009 | | | |
| Striped Dolphin | Stenella coeruleoalba | Not Listed | Non-strategic | Found in tropical to warm- temperate waters throughout the world | Both | Perrin et al. 2009 | | | |
| | Elasmobranchs | | | | | | | | |
| Giant manta ray | Manta birostris | Threatened | N/A | Found worldwide in tropical, subtropical, and temperate waters. Commonly found in upwelling zones, oceanic island groups, offshore pinnacles and seamounts, | Both | Dewar et al. 2008, Marshall et al. 2009, Marshall et al. 2011. | | | |

| Common name | Scientific name | ESA listing status | MMPA status | Occurrence | Guam/ CNMI | References |
|-------------------------|----------------------------|--|-------------|--|---------------|---|
| | | | | and on shallow reefs. | | |
| Oceanic whitetip | Carcharhinus Iongimanus | Threatened | N/A | Found worldwide in open ocean waters from the surface to 152 m depth. It is most commonly found in waters > 20°C | Both | Bonfil et al. 2008, Backus et al. 1956, Strasburg 1958, Compagno 1984 |
| Scalloped hammerhead | Sphyrna lewini | Threatened (Indo-West Pacific DPS) | N/A | Occur over continental and insular shelves, and adjacent deep waters, but rarely found in waters < 22°C. Range from the intertidal and surface to depths up to 450–512 m. Guam's inner Apra Harbor is a nursery habitat. | Both | Compagno 1984, Schulze- Haugen & Kohler 2003, Sanches 1991, Klimley 1993 |
| Corals | | | | | | |
| N/A | Acropora globiceps | Threatened | N/A | Occur on upper reef slopes, reef flats, and adjacent habitats in depths ranging from 0 to 8 m. | Both | Veron 2014 |
| N/A | Acropora retusa | Threatened | N/A | Occur in shallow reef slope and back-reef areas, such as upper reef slopes, reef flats, and shallow lagoons, and depth range is 1 to 5 m. | Both | Veron 2014 |
| N/A | Seriatopora aculeata | Threatened | N/A | Found in broad range of habitats including, but not limited to, upper reef slopes, mid-slope terraces, lower reef slopes, reef flats, and lagoons, and depth ranges from 3 to 40 m. | Both | Veron 2014 |

^a These species have critical habitat designated under the ESA. See Table B-2.

| Common Name | Scientific Name | ESA Listing Status | Critical Habitat | References |
|------------------------------|------------------------------|-----------------------|---|---|
| Hawksbill Sea Turtle | Eretmochelys imbricata | Endangered | None in the Pacific Ocean. | 63 FR 46693 |
| Leatherback Sea Turtle | Dermochelys coriacea | Endangered | Approximately 16,910 square miles (43,798 square km) stretching along the California coast from Point Arena to Point Arguello east of the 3,000 meter depth contour; and 25,004 square miles (64,760 square km) stretching from Cape Flattery, Washington to Cape Blanco, Oregon east of the 2,000 meter depth contour. | 77 FR 4170 |
| Hawaiian Monk Seal | Neomonachus schauinslandi | Endangered | Ten areas in the Northwestern Hawaiian Islands (NWHI) and six in the main Hawaiian Islands (MHI). These areas contain one or a combination of habitat types: Preferred pupping and nursing areas, significant haulout areas, and/or marine foraging areas, that will support conservation for the species. | 53 FR 18988, 51 FR 16047, 80 FR 50925 |
| North Pacific Right Whale | Eubalaena japonica | Endangered | Two specific areas are designated, one in the Gulf of Alaska and another in the Bering Sea, comprising a total of approximately 95,200 square kilometers (36,750 square miles) of marine habitat. | 73 FR 19000, 71 FR 38277 |

Table B-2. ESA-listed species' critical habitat in the Pacific Ocean^a

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^a For maps of critical habitat, see https://www.fisheries.noaa.gov/national/endangered-species-conservation/critical-habitat.

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